BIG GAME STUDIES
VOL. VI BLACK BEAR AND BROWN BEAR
Sterling D. Miller
Alaska Department of $F i s h$ and Game
333 Raspberry Road Anchorage, AK 99518-1599

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## ARCS

HARE RESOURCES




1. SUMMARY OF RESULTS

This study describes the brown bear (Ursus arctos) and black bear (Ursus americanus) populations in the area that would be influenced by a large 2-dam hydroelectric project on the Susitna River in southcentral Alaska. These dams would inundate an area of $185 \mathrm{~km}^{2}$ along an approximately $120-\mathrm{km}-\mathrm{long}$ stretch of river. Estimates of levels of impact are offered where data are adequate to make such estimates. Primary emphasis in this study was to provide baseline data on bear populations prior to project construction. This data could be compared with post-project populations to provide definitive answers on levels of impacts. Most data were based on periodic relocations of radio-marked bears.

This study was conducted in 2 phases. During the first phase it was learned that the Watana Impoundment would likely have a much greater impact on populations of bears than would the Devils Canyon Impoundment. Correspondingly, subsequent efforts emphasized the Watana project area and relatively few data were obtained on the Devils Canyon Impoundment impact area in the second phase of studies.

## 1.A. Brown Bear Results

The area of the proposed project is inhabited by a large population of brown bears. A population density of 2.79 bears $/ 100$ $\mathrm{km}^{2}$ was estimated based on capture-recapture techniques developed during the course of this study. For brown bears, the size of the impoundment-impact area was estimated to be $12,127 \mathrm{~km}^{2}$. This area included the area within 1 mean brown bear home range diameter from the Susitna River. Extrapolation of the density estimate to this area provided an estimate of the number of brown bears that would be affected by the proposed project. This estimate was 327 bears (95\% CI = 295-386) .

Bear use of the impoundment area was analyzed using 3 impoundment proximity zones: 1) within the area that would be flooded; 2) from the shoreline of the proposed impoundment to a distance of 1 mile ; and 3) from 1-5 miles from the impoundment shoreline. Brown bears used the area that would be inundated by the proposed Watana Impoundment over twice as frequently as expected under the null hypothesis that use occurred in proportion to the area of this zone. This selection was evident for males and for females not accompanied by cubs of the year. Females accompanied by newborn cubs showed selection against the area that would be inundated by the Watana Impoundment. Use of the impoundment zone was most pronounced during June. Selection was also
shown for the area that would be inundated by the Devils Canyon Impoundment. However, compared with the Watana Impoundment, the area that would be inundated by the Devils Canyon Impoundment is small and overall influence would be less.

Data on use of impoundment proximity zones formed the basis for my estimate that annual carrying capacity for 43 brown bears would be eliminated due to inundation of habitat by impoundments.

Brown bears, at least in populations that are subject to hunting, tend to develop avoidance reactions to human presence. This avoidance reaction and barriers to movements associated with the impoundments and access roads are expected to result in additional losses of habitat availability for brown bears in the study area. No estimates of the level of such losses are made here. However, the data on pre-project brown bear movements collected in this study provide the basis for making such estimates following completion of post-project studies.

The only anadromous fish stream in the study area was clearly identified as a seasonally critical habitat area for brown bears. Prairie Creek, a small tributary of the Talkeetna River, contains the highest concentration of spawning. king salmon (Onchorhynchus tshawytscha) in the upper Cook Inlet area. Salmon are easily caught by bears in this shallow creek and brown bear movements to this stream were documented from an area of more than $15,000 \mathrm{~km}^{2}$. Most bear use of Prairie Creek occurred in July and early August. The proportion of marked Su-Hydro bears fishing for salmon in Prairie Creek varied from 13\% to $38 \%$ in different years. In 1984 and 1985 $50-60$ bears were estimated to be using the creek at 1 time. The total number of different bears using Prairie Creek at some time during the salmon run was larger than this by some unknown amount. It is anticipated that disturbance displacement of brown bears from Prairie Creek will result from increased human access to the stream from access roads to and across the impoundments. The level of this disturbancedisplacement can range from slight to complete, depending on the limitations that are placed on human uses of the Prairie Creek area. Some of the limitations needed to assure continued brown bear use of Prairie Creek are under the control of the hydro-project developers. The most effective of these limitations would be to prevent access to the south side of the Susitna River in the vicinity of the Watana dam site. If Prairie Creek salmon resources were to become unavailable to project-area bears, a loss of annual carrying capacity for about 41 bears might result.

Reductions in annual carrying capacity for bears would likely be expressed through reductions in bear densities and reductions in reproductive rates. For this reason baseline data on pre-project reproductive rates were described. Separation of mother and offspring occurred when offspring were in their 3rd year of life (2.0+ years old). Mean reproductive interval was at least 3.8 years. Mean age of first litter production for females was 5.5 years (4-8). More bears (44\%) produced first litters at age 6 than at any other age. Litter size averaged 2.1 cubs (1-4), 1.7 yearlings (1-3), and 1.7 -year-olds (1-3). Cub mortality was $37.7 \%$ and yearling mortality was $21.6 \%$.

Mean home range size was $1022 \mathrm{~km}^{2}$ : $1941 \mathrm{~km}^{2}$ for males and 501 $\mathrm{km}^{2}$ for females. A few bears made identifiable movements to caribou calving areas. Subadult males typically disperse from maternal home ranges at age 2 or 3 , while subadult females typically do not disperse.

Annual brown bear harvests by hunters in the project area averaged 32 bears/year during 1983-1985. Hunter harvests are increasing in this area, a probable consequence of increased hunter effort resulting from liberalized seasons and bag limits.

Brown bears are effective predators on moose calves in the study area. No differences in predation rates between different sex and age groups were detected except that females accompanied by newborn calves had lower predation rates ( $\mathrm{P}<0.05$ ). During intensive monitoring we saw radio-marked bears on calf moose kills every 11.8 consecutive observation days. This figure led to an estimate of 3.6 moose calves killed by an average adult brown bear during the spring.

Brown bears typically denned at high elevations away from the impoundment zone. Availability of physically acceptable denning sites was not thought to be a limiting factor in this area. However, there was a tendency for individual bears to den in the same general area in successive years. Displacement of these individuals to denning areas of uncertain acceptability could result in additional mortalities or stress. Such displacement is most likely to result from disturbance occurring on the access road between the Denali Highway and the Watana Dam site. This portion of the access road runs through good brown bear denning habitat. Further displacement could result from equipment working in winter in those borrow areas that are located away from the river near good denning habitat.

## 1.B. Black Bear Results

Black bears were known to occur in the project area when this project started but the population turned out to be larger than anticipated. Correspondingly, study plans were modified
to incorporate black bears. The black bear population in the vicinity of the proposed project can be characterized as typical of a population occurring in marginal habitat: unstable in numbers from year to year with probable periodic declines due to failure of key food crops (notably berries in this area), and low productivity. Black bear habitat is better and bears are more abundant downstream from the proposed impoundments. The population in the area of the impoundments is an upstream extension of the downstream population. This population lives in an increasingly narrow finger of acceptable black bear habitat which follows the course of the Susitna River from Devils Canyon to near the upper limits of the upper impoundment. Studies downstream from the proposed impoundments were also conducted to evaluate the hypothesis that anticipated reductions in salmon-spawning habitat resulting from dam-induced changes in water flow regimes would impact downstream bears.

In the vicinity of the proposed impoundments black bear habitat is largely confined to spruce-forest areas along the river, and to adjacent shrub-lands. The size of this area, determined from movements of radio-marked bears, is $1191 \mathrm{~km}^{2}$. A black bear density estimate of 8.97 bears/100 $\mathrm{km}^{2}$ was obtained in a portion of this area, and extrapolated to the whole area to obtain a population estimate of 107 black bears (95\% CI = 93-122) in the project area during spring 1985. The population at the time this estimate was made (spring 1985) was thought to be below maximum carrying capacity. At this time the population may have been recovering from a decline caused by an apparent berry-crop failure in summer 1981.

Black bears living in the vicinity of the Watana Impoundment selected for the area that would be inundated by this impoundment. This preference was particularly evident in May and June when $52 \%$ and $46 \%$, respectively, of all locations of radio-marked bears were within the area that would be flooded by the impoundment. The population of bears in the vicinity of the Watana Impoundment was estimated to be 59 bears. In the vicinity of the Watana Impoundment, loss of annual carrying capacity for 26 bears was estimated. This loss would result from inundation. Other factors, when combined with this loss of habitat though inundation, led me to conclude that that a resident black bear population could probably not survive in the vicinity of the proposed watana Impoundment. Transient black bears from downstream areas would probably continue to use the area seasonally.

Selectivity for the lower (Devils Canyon) impoundment was much less pronounced. This was because the lower impoundment would have more black bear habitat remaining above the proposed
impoundment shoreline. Only $3 \%$ of point locations of radiomarked black bears were within the area that would be flooded by the Devils Canyon Impoundment; an additional $43 \%$ were within 1 mile of the impoundment shoreline. Under the assumptions used in this analysis, the Devils Canyon Impoundment would result in loss of annual carrying capacity, through inundation, for only 2 black bears.

Downstream from the impoundment area, black bears were found to frequent the vicinity of sloughs used by spawning salmon. Analysis of bear scats collected along these sloughs during late summer revealed that salmon remains were infrequent and that devil's club (Oplopanax horridus) berries were prevalent. Based on these results, impacts on black bear populations resulting from reduced availability of salmon could not be predicted. Such impacts may occur however (especially durịng years when berry crops fail), if salmon are an important buffer food.

Reproductive rates for study-area black bears were low compared with rates from the Kenai Peninsula, the only other area in Alaska where comparable data are available. Mean litter size was 2.1 cubs (1-4) and 1.9 yearlings (1-3). Offspring mortality during the first season out of dens was $35 \%$ and appeared higher in the upstream study area (47\%) thar in the downstream area (6\%). Such mortalities are very rare on the Kenai Peninsula where yearling bears weigh significantly more than in the Su-Hydro area. Intervals between successive production of litters averaged at least 2.7 years. Mean age at first litter production was 6.4 years (5-8); about half of the bears produced their first litters at age 7.

Reported hunter harvests of black bears in the study area averaged 13 bears/year during 1973-1985. Black bear harvests in the upstream study area are thought to be stable and low because of difficulty of access. This situation will change when roads are built to the impoundment area and after use of the impoundment itself, by hunters in boats, begins. Currently, relatively few hunters are thought to be willing to pay for a fly-in hunt for black bear.

Home ranges of black bears averaged $134.6 \mathrm{~km}^{2}, 251.5 \mathrm{~km}^{2}$ for males, and $67.1 \mathrm{~km}^{2}$ for females. Black bears tended to remain in the immediate vicinity of the Susitna River during most seasons except late summer when berries were ripening. At this time bears tended to move into shrub-land habitats adjacent to the forested habitats along the river to forage for ripening berries, primarily blueberries (Vaccinium uliginosum). During years of berry crop failure late-summer movements for some bears are much more extensive and suggest the importance of this food source.

Predation rates for black bear, recorded during periods of intensive monitoring in the spring, were 2 kills/100 consecutive observation days. This rate is lower than observed for brown bears. At this predation rate each adult black bear in the impoundment study area would kill an average of 0.7 moose calves/year.

Unlike brown bear dens, dens of black bears were located in the immediate vicinity of the Susitna River. Over half of the black bear dens in the vicinity of the proposed watana Impoundment would be inundated by the proposed project compared with $3.3 \%$ of the dens in the vicinity of the Devils Canyon Impoundment. Reuse of den sites was common in the study area. This and other observations suggest that competition for good den sites may be occurring at existing black bear densities.

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## 5. INTRODUCTION

5.A. Project Background
5.A.1. Organization and Objectives

This is the final report for black bear (Ursus americanus) and brown bear (Ursus arctos) studies conducted by the Alaska Department of Fish and Game, Division of Game, under contract to the Alaska Power Authority as part of impact assessment studies for the proposed Susitna Hydroelectric Project. Field studies were conducted from 1980 through 1985; analysis was conducted in 1986. The originally stated objectives of these studies were:
-- To determine the distribution and abundance of black and brown/grizzly bears in the vicinity of proposed impoundment areas;
-- to determine seasonal ranges, including denning areas, and movement patterns of bears; and
-- to determine seasonal habitat use by black and brown/grizzly bears.

These objectives were modified and others added during the course of study as information accumulated.

A 2-phase plan of study was developed to meet the project objectives. The first phase (1980 and 1981) was designed to provide an overview of bear movements in the study area. . This overview was intended to identify the bear uses of the impoundment vicinity that were most likely to be affected by project construction and to result in impacts on bear populations. One progress report (Miller and McAllister 1981) and 1 summary report (Miller and McAllister 1982) describing Phase I studies were prepared. Continuation studies during Phase II (1982-spring 1985) were designed to quantify the most significant impacts on bears during Phase I. These results were reported in 2 progress reports (Miller 1984 and Miller 1985a) and in this final report. This report summarizes all pertinent information collected during the project. Publication of additional analyses of peripheral information collected during this project are planned. This analysis will include analyses of habitat selection by bears. These analyses were not completed for this report because project funding was terminated just as habitat-type mapping became available.

During Phase I of this project the proposed Watana Dam was identified as having a relatively large potential for affecting bear populations, compared with the Devils Canyon Dam
(Miller and McAllister 1982). For this reason Phase II studies concentrated on bear populations in the vicinity of the Watana Dam. My plan of study did not include consideration of a project design that included only the Devils Canyon dam and such analyses are not included here.

Prediction of project impacts is a very inexact science and little published work is available. Typically, impact assessment studies do not have a follow-up phase designed to evaluate the accuracy of the predictions that are made. In this project, commitments for such follow-up work were made. Correspondingly, my emphasis was to document, using replicable study designs, the current bear numbers and use patterns of the impact area. With this information available, postconstruction studies could then quantify actual impacts and test the predictions. I have attempted to predict project impacts whenever some reasonable basis for such predictions could be derived. These predictions should be considered hypotheses that need to be tested by post-construction studies. These predictions are also offered as an aid in mitigation planning.

At the time this final report was in preparation it appeared that the construction phase of the proposed project would not soon, and may never, occur. Correspondingly, postconstruction studies designed to evaluate the impact predictions may never result.
5.A.2. Hydro Project Design

This study was designed to evaluate impacts on bears of a proposed 2-dam project on the Susitna River. The lower dam, a concrete arch at Devils Canyon, would have a normal maximum operating level of 1,445 feet above mean sea level (MSL) (maximum $=1466$ feet, minimum $=1,405$ feet). The length of the impoundment would be $41.94 \mathrm{~km}(26 \mathrm{miles})$ and it would have a surface area of $31.58 \mathrm{~km}^{2}(7,800$ acres $)$ at normal maximum operating level (NMOL). The upper impoundment, an earth/ rockfill dam at the Watana Dam site, would have a normal maximum operating level of 2,185 feet above MSL (maximum $=$ 2,202 and minimum $=2,054$ feet). This impoundment would have a length of 77.42 km ( 48 miles) and an area at NMOL of 153.85 $\mathrm{km}^{2}$ (38,000 acres). The NMOLs for each dam are illustrated in Fig. 1 and in other figures in this report where appropriate. Place names used in this report are also illustrated in Fig. 1.

## 5.B. Methods

Only general methods will be described here. Specific methods pertinent to each investigated topic are described along with the results.

Bears were captured with immobilization darts fired from a helicopter. Most bears were immobilized with etorphine (M99) but some were immobilized with Phencyclidine hydrochloride (Sernalyn) or Ketamine hydrochloride (Vetelar) and xylazine (Rompun) mixtures. Bears $<1.0$ year old were captured by hand and were not darted. Most bears were captured early in the year (April-June), but some were captured in August, at which time many bears were in relatively open habitats feeding on berries. Some black bears were immobilized in winter dens to allow replacement of collars and to make cub counts.

During 1980 through 1985, 97 different brown bears were captured. The total number of captures was 151 , and 6 of these captures (4.0\%) resulted in inadvertent capture-related bear mortalities. An additional 3-4 newborn cubs were abandoned and lost, probably as a result of our capture activities. Capture histories of all brown bears are presented in Table 1.

During 1978 and 1979, studies in areas adjacent to the Su-Hydro area were conducted on wolves, bears, moose and vegetation. Where pertinent, references to these results are used to supplement data collected during the course of this study.

During 1980 through 1985, 110 different black bears were captured. The total number of captures was 171 , and 7 of these captures (4.1\%) resulted in inadvertent capture-related bear mortality. Black bear capture histories are presented in Table 2.

All bears were marked with ear tags and lip tattoos. Bears judged to have completed $80 \%$ or more of their growth were fitted with radio collars (Telonics Inc., Mesa Arizona). Radio-marked bears were periodically tracked with fixed-wing aircraft (usually a Cessna 180 or a Super Cub) and locations of bears were recorded on $1: 63,360$ scale ( 1 inch $=1$ mile) USGS maps.

In general, monitoring frequency during periods when bears were out of dens was every $7-10$ days depending on weather conditions. For specialized studies, monitoring frequencies for individual bears were as frequent as twice daily. These specialized studies included density-estimation techniques (spring 1985), predation studies (springs of 1981 and 1984), and estimates of bear numbers at Prairie Creek (summers of 1984 and 1985).

Point locations were digitized and analyzed using geoprocessing software on a Data General computer system. Much of this analysis was done on the computer system maintained by the

Department of Natural Resources. Descriptive information associated with each radiotelemetry point location was used to sort these data and produce plots and figures. Codes and formats associated with this descriptive information are provided in Appendix 5 of this report.
5.C. Acknowledgments

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## 6. THE STUDY AREAS

The area in which bears would be affected by the proposed impoundments was defined as the study area. The size of this area was determined from data collected in this study. The size of this area is an important parameter, as the number of bears that would be affected by the impoundment was estimated by applying a density estimate, obtained in a portion of this area, to the whole area.
6.A. Upstream brown bear study area

The initial capture locations of 53 brown bears that were fitted with radio transmitters is illustrated in Fig. 2. These bears were captured in an area of $2,170 \mathrm{~km}^{2}$ centered approximately at the confluence of the Susitna River and Watana Creek. Movements of these bears, as determined by telemetry (2901 points during 1980-1985), incorporated an area totaling $13,912 \mathrm{~km}^{2}$ (excluding dispersals and atypically large movements to den sites) (Fig. 3).

The area illustrated in Fig. 3 is 1 estimate of the size of the impact area of the proposed impoundments. Another estimate was obtained using the average home .range size. Standard minimum home range grids (Mohr 1947) were used to calculate home range sizes for individual bears and for bears according to sex and reproductive status categories. Mean total home range sizes for males and females were 1941 and 501 $\mathrm{km}^{2}$ respectively, (Section 7.G.3, this report). Circles of this size would have diameters of 49.7 and 25.3 km , respectively. The mean of these 2 diameters was 37.5 km . We defined the area in which brown bears would be affected by the proposed project as the area within 37.5 km on either side of the Susitna River, from the Devils Canyon dam site to the confluence of the Susitna and Oshetna Rivers. This area totaled $12,127 \mathrm{~km}^{2}$ (Fig. 4), a value only slightly lower than the area, mentioned above, that was occupied by radio-marked bears (Fig. 3). Use of an equivalent home range criterion for each of the impoundments, considered separately, yielded an impact area of $9,452 \mathrm{~km}^{2}$ for the Watana Impoundment, $7,121 \mathrm{~km}^{2}$ for the Devils Canyon Impoundment, and $4,425 \mathrm{~km}^{2}$ common to both impoundments (Fig. 4).

Errors are associated with any method of identifying the area in which impacts on bear populations would result. The biases in the method used here result in a conservative estimate of
the affected area's size. This is because home ranges are not circular, as assumed, but are ellipses with (typically) longitudinal axes perpendicular to the river. These longitudinal axes connect spring habitats along the Susitna River with denning habitats in the mountains away from the river.

## 6.B. Upstream Black Bear Study Area

The upstream black bear study area was relatively easy to define based on relocations of radio-marked individuals. This is because black bear habitat is largely restricted to the immediate vicinity of the Susitna River and its major tributaries such as Watana and Tsusena Creeks (Fig. 5). The initial capture locations of 32 bears that were radio-collared incorporated an area of $1,120 \mathrm{~km}^{2}$ (Fig. 5). Subsequent radio locations ( $N=2195$ ) of these bears (excluding dispersers) incorporated an area of $2,950 \mathrm{~km}^{2}$ (Fig. 6). This area is an overestimate of the amount of black bear habitat in the study area as the convex polygon method of delineating home ranges incorporates areas where radio-marked black bears were never located (Fig. 6).

Black bear habitat in the study area was more precisely defined using locations of all bears spotted ( $N=282$ ) and radio-tracked ( $N=2,273$ ) during the period 1980-1984. These points were plotted (1:63,360 scale) and a line was manually drawn around them such that all points were included except those considered to represent erratic movements ( $N=54$ for radio locations and 27 for locations of non-radioed bears). This area totaled $1,191 \mathrm{~km}^{2}$ (Fig. 7).
6.C. Downstream Black Bear Study Area

The area downstream from Devils Canyon was defined as the downstream study area. Bears were studied in this area to determine what impacts anticipated project-related reductions in salmon spawning habitats (especially sloughs) would have on bear populations. Capture locations for 22 downstream black bears that were radio-collared incorporated an area of $250 \mathrm{~km}^{2}$ (Fig. 8). Subsequent relocations ( $\mathrm{N}=616$ ) of these bears incorporated an area of $1,949 \mathrm{~km}^{2}$ (Fig. 9). This area was defined as the downstream black bear study area. Unlike the upstream black bear study area, most of the area incorporated in the polygon illustrated in Fig. 9 is black bear habitat. Bears that moved between upstream and downstream areas were not included for the purposes of defining these study areas.

## 7. BROWN BEAR RESULTS

## 7.A. Number of Bears in Impoundment Impact Zones

In Section 9 of this report I derive an estimate of the number of bears in the impoundment impact zone (Fig. 4). This estimate is based on extrapolation to brown bear habitat in the impoundment impact zone, from a density estimate 12.97 bears/100 $\mathrm{km}^{2}$ ) obtained in part of this zone. The $95 \%$ confidence interval for this density estimate is similarly extrapolated to the impact zone without modifications designed to reflect the extrapolation. The resulting estimate for the number of brown bears in the impoundment impact zone was 327 (295-386). I estimate that 68\% of these bears were 2.0 years old or older (Miller et al. in press, Appendix 2). This is a larger number of bears than $I$ estimated in previous .reports (e.g., Miller and McAllister 1982). This difference is primarily the result of estimates being based on lower bear densities ( 2.44 bears/100 $\mathrm{km}^{2}$ ) estimated in 1979 in an adjacent study area (Miller et al. 1982).
7.E. Use of Impoundment Impact Zones by Brown Bears
7.B.1. Use by season, sex, age, and reproductive status

Miller and McAllister (1982:58-60) provided a preliminary assessment of brown bear use of impoundment area proximity zones; that analysis was combined with data collected subsequently (1980-1984) for the analysis presented here. Three zones were identified for each impoundment area: within the area that would be flooded by the proposed impoundments (zone 1), within 1 mile of the normal maximum operating level (NMOL) shoreline of the proposed impoundments (zone 2), and from 1 to 5 miles from the NMOL shoreline of the proposed impoundments (zone 3). An illustration of these impoundment impact zones is presented in Fig. 10. Data collected farther than 5 miles from the NMOL shoreline of the proposed impoundments ("zone 4") are also reported but not included in the analysis. A vertical north-south line was drawn to separate the 5 -mile polygons of each impoundment which would, otherwise, have overlapped.

The purpose of this analysis was to determine whether bears were selecting for the impoundment area and, if so, at which periods of the year selection occurred. Chi-square analyses were used to make this determination under the null hypothesis that the number of point locations found in each of these 3 zones was in the same proportion as the area in each zone. Not all assumptions of the Chi-square analyses were met because multiple observations were made of the same bear so the data points were not independent of each other. Seasons considered included "spring" (April 1-June 30) and the rest of the year. Data collected in 1980 through 1984 are analyzed.
7.B.1.a. Watana Impoundment

In the Watana Impoundment area, brown bear use of the 3 impoundment zones was significantly different than expected for all months lumped and in the spring (Table 3). Use of the impoundment zone was over twice the expected values (Table 3). No significant variations from expected values were observed during the period July 1-March 31 (Table 3).

Brown bear males also used the 3 Watana Impoundment zones significantly differently than was expected under the null hypothesis (Table 4). In all months and in both periods, use of the impoundment zone was higher than expected values (Table 4).

Brown bear females also used the 3 impoundment zones of the Watana Impoundment differently than expected under the null hypothesis (Table 5). This difference was significant for all months lumped and in the spring period, but did not differ from expected values during the July 1-March 31 period (Table 5).

When a similar analysis was done for brown bear females with cubs-of-the-year, no significant variations from expected values were observed for all periods lumped, or for either of the two time periods (Table 6). This is because these bears tend to stay at higher elevations, well away from the impoundment area, during years when they have newborn cubs. I suspect that this behavioral trait is designed to reduce predation on their cubs, by other brown bears (especially adult males) that are concentrated in lower-elevation habitats early in the year. To test this hypothesis I compared the use of these 3 impoundment zones (both impoundments lumped), during years when the same set of females had cubs-of-the-year with the years when they did not (Table 7). During years when they had newborn cubs these bears utilized these 3 zones differently than during years when they did not have newborn cubs; use of the impoundment zone was less than expected when these females had cubs (Table 7).

The proportion of time spent in the actual impoundment zone was highest during the period 1-15 June for all bears (18.4\%, Table 3), and for female bears (25.5\%, Table 5). The impoundment zone was most heavily used by males during the last 2 weeks of June (23.2\%, Table 4).

The percent of point locations in each proximity zone in each month is illustrated in Fig. 11 for the Watana and Devils Canyon impoundment areas. Comparison of these 2 impoundments illustrates the greater degree of selectivity for the Watana Impoundment zone than for the Devils Canyon Impoundment zone (Fig. 11).

## 7.B.1.a. Devils Canyon Impoundment

Similar analyses were conducted for observations within the 3 proximity zones of the Devils Canyon Impoundment but because of the smaller sample of point-locations in this area and because of the much smaller area that is anticipated to be flooded by the Devils Canyon Impoundment, analyses by season were not possible. Use of these 3 zones (all months lumped) was significantly different for females without cubs-of-theyear and for all bears lumped. Use was not significantly different for males (Table 8). The most significant deviations from expected values were observed in zone 3 , which was used more than expected. Zone 1, the impoundment area, was also used more than expected (Table 8). However, because zone 1 was so small in area, it had only slight use altogether (Table 8).

## 7.B.2. Prediction of impacts

The above analysis demonstrates that the area to be flooded by the proposed Watana Impoundment, as well as the area within 1 mile of the impoundment shoreline, is important habitat to brown bears. Use of this habitat is especially intense during the spring, but is significant throughout the year as well.

Conversion of this evident selectivity to estimates of impacts on the brown bear population when impoundment area habitats are no longer available is not straightforward. I suspect the impact on brown bear populations will be expressed through reductions in bear productivity and in population density. Such reductions from existing population levels might not occur or might be dampened in magnitude if there currently is substantial excess carrying capacity which is not being used by bears and that could be substituted for the habitat that would be lost to the impoundment. Such substitutions would have to be available during the same season. Loss of important spring habitats where bears are foraging for roots and new spring growth, for example, would likely not be fully compensated for by increases f that might result from mitigation efforts for example), in late summer food sources (e.g., salmon or berries). Even if the current population is below carrying capacity, project-related losses of carrying capacity need to be considered in mitigation planning. These losses can be considered loss of bear habitat potential.

The conceptual model $I$ used to estimate impacts from the point location data includes the following assumptions:

1. The proportion of point locations found in a geographic zone represents a corresponding proportion of the bears' total energy budget acquired from resources found in that
zone (this assumption will lead to an underestimate of the importance of the zone in cases where positive selection for that zone is occurring).
2. Substitute resources are not available (in cases where the population is below carrying capacity this assumption will overestimate the impact of loss of the geographic zone).
3. Loss of resources that are especially heavily used during 1 season of the year cannot be made up through extra use, at other seasons, of resources available in other zones (this assumption, also, will probably yield an overestimate of impact).
4. Impact on habitat carrying capacity can be expressed by summing the impacts on individuals (determined in \#1).
5. Radio-marked bears in this study are representative of the population estimated to use the impoundment impact area (Section 7-A of this report).
6. Reduction in carrying capacity would result only from flooding of the impoundment area; no reduction would result from displacement to habitats along the shoreline of the impoundment (this assumption would certainly result in an underestimate of impoundment impacts).

Data obtained in this study were analyzed under these assumptions. Nine radio-marked males and 25 radio-marked females averaged $13.3 \%$ of point locations during the spring period in the impoundment zone; an additional 17.0\% of point locations were within 1 mile of the impoundment shoreline (Table 9). If, as previously estimated, the impoundment impact zone includes 327 brown bears and $13.3 \%$ of the carrying capacity for this population will be eliminated, a decline in carrying capacity for an estimated 43 bears would be expected from habitat inundation under the above-listed assumptions.

Because some substitution of resources would undoubtedly occur, I expect that this estimated impact is more likely to be an overestimate than an underestimate of the project's impact resulting from inundation of habitat. This expectation is supported by the observation that 14 of the radio-marked bears (41\%) had no point locations in the impoundment-impact area (Table 9). Nine of these bears (26\%) had no locations within the 1 -mile proximity zone either (Table 9). Although these bears may have used these zones without being detected, it is probable that these data indicate availability of spring food resources outside of the immediate impoundment impact area.

## 7.B.3. Mitigative Measures

Potential measures to mitigate for loss of spring foraging habitats resulting from inundation include:

1. Increasing the abundance of foods used in the spring in substitute areas;
2. substitution of foods utilized during other seasons for losses of spring carrying capacity; and
3. indirect mitigation (e.g., bear habitat protection elsewhere or transference of mitigation values to other species).

It is uncertain if measure \#2 would be efficacious. Implementation of either measure 1 or 2 would be experimental as little is known about how to accomplish increases in bear habitat carrying capacity (Proceedings--Grizzly Bear Habitat Symposium, Missoula, Montana, 1985, Intermountain Research Station, Ogden, Utah, General Tech. Report INT-207 252pp.).

## 7.C. Disturbance-Displacement from Remaining Habitat

The degree to which brown bears are compatible with increased human presence is not completely clear. In most areas. it appears that brown bears will tolerate the proximity of humans better than humans will tolerate the presence of brown bears. In large National Parks, like Denali National Park, where grizzlies are not hunted and special efforts are made to accommodate grizzly bear needs, bears remain abundant regardless of high levels of human use. More typically, however, increasing human activity in an area correlates with declines in grizzly numbers (Herrero 1985; Pulliainen 1972 and 1982; Horejsi 1986; Horejsi, in press; Elgmork 1983). Pulliainen (in press) observed that the population of bears in Finland declined as human populations and impacts increased. However, the decline was followed by an increase in absolute numbers resulting from immigration from Russia. Mattson et al. (in press) documented a retreat of grizzlies, especially females, from roads and developments in Yellowstone National Park. Archibald et al. (in press) also documented avoidance by adult female grizzly bears following logging development of an area.

Some of these declines result from humans killing bears in both sport and nonsport circumstances. Increased killing by sport hunters is a direct consequence of improvements in accessibility and interest in hunting; increased killing in nonsport circumstances results from intolerance or inability of humans to coexist with bears (Miller and Chihuly, in
press). In addition, I suspect there is strong selective pressure for bears in populations that are heavily hunted, to learn to avoid man. Bears that fail to learn this behavior at an early age are easier prey for hunters. If this theory is correct, then increased human presence in the project area will result in abandonment of the area by adult bears that are displaced as a result of intolerance of people. This abandonment may also occur in areas where bears are not hunted (see Jope 1983), but is probably more evident in areas like the project area where bear hunting occurs. Young bears that have not learned this avoidance behavior may be especially vulnerable to nonselective hunting effort (Bunnell and Tait 1980).

Although most bear biologists would agree that disturbance displacement occurs, there is little direct quantitative documentation. The number of visitors to the bears' fishing area at McNeil River State Game Sanctuary is limited. This limitation is based on observations that too many visitors resulted in fewer bears visiting the portion of the sanctuary where bears were most concentrated (Faro and Eide 1974). In their preliminary assessment of the effects of construction of the Terror Lake Hydroelectric project on movements of Kodiak bears, Smith and Van Daele (1985) observed short-term shifts of activity areas of individual brown bears, away from construction sites. These authors observed no major movements away from construction activities and 1 bear denned within 0.4 km of an access road. Bear problems resulting from contractors' inadequate disposal of garbage were observed in this Kodiak study (Smith and Van Daele 1985).
7.C.1. Impoundments, access roads, and accidental mortalities

Although bears swim readily and are known to swim across impoundments, movements across the impoundment will probably be restrained, to some degree, compared with movements bears currently make across the river. Simpson (1986:21) studied movements of grizzly bears in the vicinity of the Revelstoke Reservoir in British Columbia and noted that "grizzlies would cross a river but not the reservoir." At Revelstoke, Richard L. Bonar (April 18, 1985, interview transcribed by Bill Steigers of the Susitna Project Group of LGL) noted "the radio-collared bears [both species] haven't crossed as often as they did before the water came up."

Although some impact is probable, it is impossible to guess how much movements across the river will be restrained by the Susitna impoundments. In this study we concentrated on documenting how frequent crossings were during the preconstruction phase so comparisons could be made during a post-construction study. Such comparisons will permit more accurate predictions of effects in future impact assessment studies.

The number of river crossings for each radio-marked bear in each year with $>5$ non-den observations varied from 0 to 10 (Table 10). Clearly, the number of documented river crossings is directly related to frequency of observation, so the number of observations is also provided in Table 10. For the purpose of this analysis a "bear-year" was defined as a year in which we obtained more than 5 radio-locations of a radio-marked bear away from its den site. For males, crossings were observed for 27 of 32 bear-years (84.4\%); for females crossings were observed for 38 of 77 bear-years (49.4\%) (Table 10). Of 658 point locations for males, 98 (14.9\%) had a documented crossing of the Susitna River after the preceding location (Table 10). Of 1,668 point locations for females, 152 (9.1\%) had a documented crossing of the Susitna River after the preceding location (Table 10). No doubt these values were larger for males than for females because males had larger home ranges and, as a result, the home ranges of a higher proportion of males incorporated both sides of the river. Movements of bears living north of the river to the Prairie Creek salmon fishing area could be restrained by the impoundment and associated facilities.

In addition to inhibiting movements across the reservoir, movements up and down the river would likely be restricted to some degree by inundation of tributaries. These tributaries, such as Watana Creek (Fig. 1), can be easily crossed at present.

Increased human activity in the vicinity of the impoundment would also likely act to displace bears from habitats along the reservoir shoreline. This disturbance would be greatest in the vicinity of communities established to house construction and operation workers.

Disturbance would also be significant in the vicinity of recreational facilities established as outlined in the recreational plan. The objective of these facilities is to provide increased recreation opportunities for as many people as possible. I suspect this objective is inimical to maintaining the present population of adult brown bears in the project area. The area affected by the proposed recreation plan is much larger than the area that would be directly affected by impoundments and construction facilities. Anticipated recreational developments and trails are expected to be built many miles away from the dam sites, reservoirs, and access roads.

The proposed route of the access road (Fig. 1) is in heavily used brown bear habitat along most of its length from the Denali Highway to the Devils Canyon dam site. This route would bisect the home ranges of many brown bears. Miller and Ballard (1982b) noted that movements of transplanted brown
bears appeared to be inhibited by roads and it is probable that the access road would also modify normal bear movements in the impoundment area. Smith and Van Daele (1985) observed little displacement of brown bear by traffic on roads built for construction of the Terror Lake hydroelectric project.

Increased human presence in brown bear habitat is likely to result in additional mortalities of bears through killing of nuisance or dangerous bears (Miller and Chihuly, in press, Appendix 3) and accidents. Such mortalities and problems were observed for both species of bears during construction of the trans-Alaska oil pipeline (Follmann and Hechtel, in press). Many of these problems resulted from feeding of bears and from inadequate garbage disposal (Follmann and Hechtel, in press). During construction of the Terror Lake hydroelectric project on .Kodiak Island no mortalities from these causes were documented but bear problems resulting from inadequate garbage disposal were observed (Smith and Van Daele 1985).
7.C.2. Levels of impact and mitigation measures

Maximum estimated level of impact from disturbance displacement was estimated in the same manner as loss of carrying capacity due to inundation. For this purpose it was assumed that all carrying capacity in the zone from the proposed impoundment shoreline to a distance of 1 mile (Zone 2 in the proximity analysis) would become unavailable to brown bears as a result of disturbance displacement. Point locations in this zone totaled $17 \%$ of all point locations (Table 9). For the brown bear population estimate of 327 in the impoundment area, a loss of $17 \%$ of carrying capacity would result in an estimated decline of carrying capacity for 60 brown bears. This estimate is subject to the same qualifications outlined above for loss of carrying capacity due to inundation. In addition; I suspect that loss of carrying capacity due to disturbance displacement would be proportionately less than loss of carrying capacity due to inundation; more bears could coexist with disturbance than could obtain forage from flooded habitats.

The most effective mitigation measures designed to minimize losses of habitat due to disturbance displacement will be those that restrict human activities and facilities to the smallest possible area. Concentration of construction facilities and human habitations will have this effect, as will minimizing the area in which access by the public will be facilitated. Disturbance-displacement of brown bears in the area between Kosina Creek and Prairie Creek can be minimized, for example, if public access by road to the south side of the Susitna River is not provided and if recreation facilities in this area are not built. Strict enforcement of state
regulations regarding feeding of wildife and disposal of garbage will also help reduce incidence of bear problems and killing of bears that have become nuisances.

## 7.D. Brown Bear Use of Prairie Creek Fishing Area

## 7.D.1. Level and time of use

Each year many brown bears in the Su-Hydro study area move in July and August to Prairie Creek, a tributary of the Talkeetna River that runs out of Stephan Lake. The purpose of these movements is to fish for king salmon (Oncorhynchus tshawytscha) which run in this small creek at this time. Sport fisheries biologists with the Department of Fish and Game report that Prairie Creek supports the most concentrated king salmon spawning area in the upper Cook Inlet region (Larry Engle, pers. Commun.). Salmon are relatively easy for bears to catch in Prairie Creek compared with larger rivers like the Gulkana.

Radio-marked brown bears have been documented moving from an area of $15,300 \mathrm{~km}^{2}$ to utilize Prairie Creek salmon resources (Fig. 12). For just radio-marked males the area was 15,285 $\mathrm{km}^{2}$, for just females it was $3,300 \mathrm{~km}^{2}$. The actual area of attraction to brown bears is larger than this because these data are biased as a result of tagging radio-marked bears only in the Su-Hydro study area which is north and east of Prairie Creek. Bears moving to Prairie Creek from south and west directions would have had no chance of being radio-marked in this study. One radio-marked bear (G407) moved to Prairie Creek to fish for salmon from upper Gold Creek (downstream from Devils Canyon) at a time when pink and chum salmon (O. gorbuscha and O. keta) were abundant and much closer in lower Gold Creek. This movement may indicate that the king salmon in Prairie Creek may be preferred over salmon resources elsewhere.

The proportion of radio-marked Su-Hydro study area bears that have been documented moving to Prairie Creek to fish for salmon has ranged from $13 \%$ in 1981 (a year when little moritoring was done as a result of poor flying conditions) to $38 \%$ in 1984. (Table 11). This proportion appears higher for radio-marked males (50\% in 1984, excluding dispersers) than for radio-marked females (33\% in 1984) (Table 11).

In summer 1984 and 1985, efforts were made to estimate the number of bears at Prairie Creek at 1 time during the salmon rurı. This number is difficult to determine from direct counts because of dense vegetation along the banks of Prairie Creek. This vegetation makes it very difficult to spot the bears from the air as bears need only to move a few feet from the creek
to be well hidden from sight in alders. Correspondingly, we attempted to census the bears in this area using the ratio of radio-marked to unmarked bears spotted during intensive search efforts along the length of the creek between upper Murder Lake and the Talkeetna River. The search area was a strip of about 1 km on each side of Prairie Creek and about 0.5 km on each side of salmon-carrying tributaries of Prairie Creek. Marked bears that were spotted were identified by their radio frequencies but radio-tracking gear was not utilized in finding the bears during the search effort. The search pattern flown was a circular one overlapping Prairie Creek from both sides and following the tributaries on both sides of Prairie Creek to the limit of salmon spawning. Subsequent to the search effort, radio-tracking gear was utilized to determine how many radio-marked bears were present in the area previously searched. These surveys were flown by. experienced bear spotters in both years: pilot Al Lee (Lee's Air Taxi) in 1984 and Harley McMahan in 1985. I was present as spotter and radio-tracker both years.

Results of flights on 29 July and 1 August 1984 are presented in Table 12. On 29 July an estimate of 48 bears (95\% confidence interval $=12-180$ ) was obtained; on 1 August an estimate of 33 bears ( $95 \%$ confidence interval $=10-62$ bears) was obtained (Table 12). These estimates include only bears that were not accompanied by their mothers for bears at least 2.0 years old). An estimate including these subadults would be 30-40\% higher, or about 44-65 bears. The large confidence intervals of this estimate result from a low number of marked bears being present in the search area when the census was conducted (only 4-5, Table 12).

Equivalent data were collected in mid summer 1985 (23-27 July) during replicated morning and evening flights in a Piper Super Cub (PA 18), for a total of 8 counts. On 6 August another flight was conducted in a Cessna 180 flown by Larry Rogers (Kenai Air Alaska) with Randy Fairbanks, Richard Fleming, and me as observers. This flight was incomplete at the lower end of Prairie Creek because of fuel shortage. The 6 August flight was poorest in terms of visibility because of the larger airplane and increased number of observers; however, it may have provided the best estimate because of the larger number of marked bears that were present (Table 13). Summarized results of these 9 flights are presented in Table 14.

The data in Table 26C were used to calculate 9 separate Petersen Indices. These estimates varied from 27 to 107 bears and averaged 51 bears. The $95 \%$ confidence interval for this average was +22 bears or $43.7 \%$. Another estimate was obtained using the bear-days estimator (Miller et al.r in press, see

Appendix 2). Using this estimator, the estimate for the average number of bears present in the search area was 59 with a $95 \%$ CI of +23 bears (Table 14). These estimates include subadults.

The estimates from 1984 and 1985 both indicate that an average of $50-60$ brown bears used Prairie Creek at any 1 time. Because some bears were just out of the search area and because bears come and go from Prairie Creek, the total number of different individuals that use Prairie Creek during the salmon-spawning period (1 July-15 August) is higher than this estimate by some unknown amount. My guess is that 70-120 different brown bears may use Prairie Creek salmon resources at some time during the king salmon run.

The areas occupied by 6 radio-marked brown bears during the period 23 July-6 August 1985 are illustrated in Fig. 13. These 6 bears moved an average of 2.4 km between successive locations during this period (range $=0.2-7.4 \mathrm{~km}$ ). The mean distance between points 24 hours apart was 3.3 km (range $=$ $0.4-7.9 \mathrm{~km}$ ). Only points on the periphery of these movements are illustrated in Fig. 13. Locations of all bears spotted between 23 July and 6 August are illustrated in Fig. 14.

I believe that most bears that utilize Prairie Creek are offspring of females that used Prairie Creek. However, my sample of marked subadults is too small to demonstrate this. Some bears that live near Prairie Creek (e.g., female 299 in the Fog Lakes area) do not go there, while others travel from great distances (e.g., female 407 from upper Gold Creek). Some bears find out about Prairie Creek on their own. Male 382 was weaned in 1983, at age 2, from a mother that did not use Prairie Creek (313). This subadult male stayed near his maternal home range (centered on Tsusena Butte) in 1983 and 1984, but in 1985 he dispersed south and fished along lower Prairie Creek. This bear shed his drop-off collar at Prairie Creek in August 1985 and his subsequent movements are unknown.
7.D.2. Potential impacts of project on brown bear use of Prairie Creek

The amount of disturbance which will occur in the Prairie Creek area is uncertain, as are the relative impacts of different levels of disturbance on bears. Increasing levels of disturbance through increased recreational use of the area are currently evident and likely to continue regardless of whether the dam is built. If the dam is built, however, the improved access to the area will result in greatly accelerated disturbance impacts. There is a real potential that this disturbance will become so great that bears may be excluded altogether from this habitat. This has nearly happened
elsewhere in Alaska; for example, along sections of the Kenai and Russian Rivers that are currently heavily utilized by humans during salmon runs.

Our work at Prairie Creek was designed to estimate the number of bears using Prairie Creek during the salmon run. I also wanted to provide the baseline data needed to document the anticipated decline in bear use of Prairie Creek, which will occur if the impoundment is built and the Prairie Creek area is developed. This documentation will result from replicated surveys flown subsequent to construction. These surveys should reveal whether development has resulted in the anticipated exclusion of many brown bears from this resource. In order to assist in this documentation, the human habitations present in 1985 in the Prairie Creek-Stephan Lake area are documented in Fig. 15. Many of these habitations were built in recent years and it is clear that human presence and impact in this area is increasing.

The exclusion of brown bears from Prairie Creek will result, in part, from increased numbers of non-sport brown bear kills by the increased number of recreational users who will have access to the area subsequent to construction of access routes from the Denali Highway to and across the impoundment. More important, however, will be the effects of disturbance exclusion wherein brown bears will abandon the area because of the anticipated large increase there in numbers of humans. Increased disturbance-displacement will result from increased recreational use of the Prairie Creek area by boaters (especially those floating down Prairie Creek from Stephan Lake), fishermen, hikers, and other recreational activities, as well as from increased industrial activities (mining, logging, tourist lodges, etc.). These activities will increase markedly in the Prairie Creek area once public access is provided by means of the proposed access road to the project area. Disturbance to the Prairie Creek area can be minimized if public access by roads crossing the Watana dam site is not allowed.

All of these activities are not inherently incompatible with bears. In Katmai National Monument, tourism and recreational activities coexist with many salmon-fishing brown bears at Brooks Camp (B. Gilbert and K. Jope, pers. commun.). One important difference between Brooks Camp and the Susitna project area is that bears are protected from hunting in national parks. Where hunting is legal, bears likely develop a more wary reaction to human presence.

## 7.D.3. Level of impact on brown bear

The worst-case scenario is used here to estimate impacts of the project on brown bears using Prairie Creek. Research subsequent to the project will likely reveal less of an
impact, but at this time, I have no realistic method of estimating how much less this could be. The worst-case scenario is that 100-120 brown bears use Prairie Creek salmon resources annually and that the project and related disturbances will accelerate development of the Prairie Creek area until bears are completely excluded from Prairie Creek, the only salmon stream with readily catchable fish that is available in the study area around the Watana Impoundment. Absence of this food resource would likely act to reduce bear density in this area and to lower the reproductive rates of remaining bears (see Section 7.G.1, this report). No estimate of how much lower reproductive rates might be is offered here; this would probably be expressed as a longer reproductive interval.

Assuming that all of the difference in bear density between the Su-Hydro study area $\left(2.79 / 100 \mathrm{~km}^{2}\right)$ and the upper Susitna River study area (2.44/100 $\mathrm{km}^{2}$ ) (Miller and Ballard 1982a) results from availability of Prairie Creek salmon, a reduction in density of about 0.35 bears/l00 $\mathrm{km}^{2}$ is indicated. In the Su-Hydro study area of $11,704 \mathrm{~km}^{2}$ this would mean an estimated elimination of average annual carrying capacity potential for 41 bears. By these calculations $59 \%$ of the estimated 100 bears currently using Prairie Creek salmon resources would find acceptable alternatives to these resources.

This model of impact levels is certainly simplistic as, among other things, there are no data indicating bears are currently at carrying capacity. If bears are currently below carrying capacity, reduction in availability of any single food resource would have less impact on the existing population. However, this estimate provides a reasonable starting place for mitigation planning.

## 7.D.4. Potential mitigation efforts

Prairie Creek is the clearest example of a critical habitat for brown bears that $I$ found in the vicinity of the proposed hydroelectric project. As such, protection of this area from the impacts discussed above offers an obvious opportunity to mitigate for losses of brown bear habitat that will occur as a result of the project. This mitigation could be achieved if the area surrounding Prairie Creek were obtained by the State and put into an appropriate land-use designation such as a state game refuge. This protection would not result in any absolute increase in numbers of brown bears in the study area. Protection of Prairie Creek as a salmon fishing area for bears probably would, however, help maintain larger populations of bears than would be able to exist in this area without such protection of this habitat. As this is the only kind of mit:igation that is likely to be effective for the losses that
the project would cause to brown bear populations in the study area, protection of Prairie Creek as a food source for salmon-fishing brown bears should receive the attention of mitigation planners. The factors necessary to adequately protect Prairie Creek from exclusion impacts include:

1. Restrictions on human use (including float traffic on Prairie Creek) between 1 July and 15 August, at least; and
2. Minimal human development and impacts in the larger area surrounding Prairie Creek, such as the Fog Lakes area.

It is noteworthy that the recreational plan currently under consideration as part of the Federal Energy Regulatory Commission. license application would most likely be incompatible with either of these requirements. Among other things it is highly questionable whether, for example, there would be any point in protecting Prairie Creek as a state game refuge or critical habitat area if road access to the south side of the Susitna River is provided as a result of the project. Such access would almost certainly result in levels of increased human use of the Prairie Creek area. This increased use would, in my view, result in reduced brown bear use of the area and the degree of reduction would be directly related to the level of disturbance.
7.E. Downstream Impacts, Brown Bears

During this study little emphasis was given to brown bear populations downstream from the Devils Canyon Dam site. As part of downstream black bear studies (Section 8E, this report) and from observations of 3 radio-marked brown bears, however, some insights into potential sources of impact in this area were gained.

Brown bear populations occur along the Susitna River to its mouth on Cook Inlet. It is my impression that these populations become progressively less dense downstream from the Devils Canyon Dam site. Brown bear tracks along the salmon-spawning sloughs off the Susitna River were very common, especially above the confluence with the Indian River. I expect most of this use was by locally residing bears, because except for 1 dispersing subadult (342), no brown bears radio-marked upstream from Devils Canyon moved downstream during this study. Such downstream movements might become evident if upstream bears were displaced from Prairie Creek (Section 7D, this report).

The project's major downstream impact on brown bears would likely result from the anticipated reduced availability of salmon in these sloughs. Estimates of the levels of salmon
reduction that would occur are not available. Correspondingly, much speculation on potential secondary impacts on bears is not warranted. It is noteworthy, however, that there has been a dramatic increase in the resident human population in the area between Devils Canyon and Talkeetna in recent years; most of this increase is the result of state land disposals in the area. I expect that the effect of this human presence on bear populations in the downstream area will be many times greater than effects resulting from construction of the impoundments. These human-caused impacts would be the result of increased sport and non-sport kills and disturbance displacement.
7.F. Cumulative Impacts, Brown Bear

The proposed project's cumulative effects on brown bears may be greater than the sum of individual effects. This is because impact mechanisms that would have little or no impact considered separately may act synergistically and, in total, produce significant impacts. Methodology to identify and quantify such cumulative impacts on brown bears has been described by Christensen (1985), Young (1985), Winn and Barber (1985), and Weaver et al. (1985). An effort to conduct similar cumulative effects analyses should be accomplished as part of environmental impact assessments undertaken for the Susitna Hydroelectric Project. In this report only some examples of such impacts will be discussed.

Adequate high-quality food is probably the single most important life requisite for bears of both species. This is because bears have only 5-7 months of activity. During this time bears must obtain the energy reserves needed to reproduce and to sustain themselves in their dens. If a pregnant female does not attain a sufficient threshold of condition to permit successful rearing of a litter of cubs prior to den entrance, then she should not invest energy in gestation and lactation. In such cases implantation of the embryo into the uterus may not occur and the female will "try again" the following year.

Energy budgets of bears have not been adequately studied, but it is reasonable to assume that super-abundance of foods in 1 season cannot completely compensate for substandard foods in another season. In such a model, superabundance of late summer foods (berries and salmon for example) would not compensate for loss of early spring foods (through inundation by impoundments, for example). In similar fashion, reduced availability of early spring foods combined with reduced quality or availability of late summer foods (loss of Prairie Creek salmon or blockage of travel corridors to berry feeding areas, for example) would likely have synergistic effects on bear numbers. The net impact would be greater than the sum of the individual parts.

In preceding sections I made estimates of carrying capacity losses that might result from various impact mechanisms. Loss of bear habitat carrying capacity would cause reductions in the existing bear populations only if these populations are currently at or above carrying capacity of the habitat. If not, these estimates represent losses in carrying capacity potential. Carrying capacity is a useful theoretical concept but techniques to evaluate it are lacking for most species. Density can be a direct estimate of carrying capacity, as existing density must be at or below carrying capacity unless the population is declining, or about to decline, as a result of lack of resources.

I do not know how to measure bear carrying capacity in the Su-Hydro area or elsewhere but $I$ can subjectively evaluate where the existing population is relative to its theoretical carrying capacity based on density, reproduction, and resource-availability comparisons with other areas. Brown bear density and reproductive rates are high in the Su-Hydro area compared with other interior Alaskan areas (Miller and Ballard 1982a; Miller et al., in press, Appendix 2; and Section 7.G.1 of this report). The most obvious difference in resource availability between the Su-Hydro area and other interior Alaskan areas is the seasonal availability, to many bears, of salmon in Prairie Creek.

The high productivity of the existing Su-Hydro bear population indicates that this population is certainly not above the habitat's carrying capacity. At present the primary factor that could cause existing bear populations. to be below carrying capacity in the Su-Hydro area is hunting. Since 1980 liberalized seasons and bag limits in Unit 13 have resulted in increased bear harvests in the study area and elsewhere in Unit 13 (Section $7 . G .2$ of this report). It is probable that these increased harvests have reduced bear population density in the study area below levels that existed prior to 1980. If this is true, excess carrying capacity may exist which could buffer the existing population from project-related reductions in carrying capacity.

## 7.G. Brown Bear Biology

7.G.1. Brown bear productivity

Along with changes in bear numbers and density, I suspect that reductions in food supply that would result from the project would cause changes in productivity. Currently this population appears to be one of the most productive that has been documented. The primary factor in this high productivity is the short reproductive interval; females were never observed to keep their offspring with them longer than 2.8
years. This leads, commonly, to a reproductive interval of 3 years. In no case during this study did a female enter a winter den with 2-year-old offspring. In. Denali National Park, $7 \%$ of litters (5 of 69) of 2 -year-olds remained with their mothers another year (Murie 1981). Entering dens with 2-year-old or older offspring is common for brown bears in other areas (Bunnell and Tait 1981; Reynolds and Hechtel 1976, 1984, and 1985), including areas where bears live in apparently much more productive habitats such as Kodiak Island (Smith and Van Daele 1985 and 1986, Barnes 1985) and the Alaska Peninsula (Glenn et al. 1976).

Data on productivity are provided in this section to provide the baseline data needed to measure changes if the proposed project is completed. No estimates of project-caused changes in productivity are offered. I suspect an increase in reproductive interval and age at first reproduction would be the parameters most likely to be affected. In a study just north of the Alaska Range from our study area, Reynolds and Hechtel (1985) found that some females entered dens with 2-year-old offspring. Their study area is equivalent in many respects to our study area except that salmon are unavailable in their area. Salmon are available to some Su-Hydro study area bears at Prairie Creek (Section 7.C. 2 this report).

## 7.G.1.a. Litter Size and Offspring Mortality

Thirty-eight litters of newborn cubs that were observed following their emergence from dens averaged 2.1 cubs (range $=1-4$ ) (Table 15). These data exclude project-related mortalities. Twenty-two of 59 cubs were lost before they emerged from their dens in the following year (37.7\% mortality) (Table 15). The mortality rates for newborn brown bears observed in this study were near the upper limit for the studies reviewed by Bunnell and Tait (1985), at 30\%-40\%. Higher mortality rates have been found in southeast Alaska (Schoen, pers. commun.).

Causes of mortality were investigated using expandable dropoff: transmitter collars (Strathearn et al. 1984). These transmitters were on very slow pulse when active (17 pulses/ mirute or "ppm"), speeding up to about 45 ppm on inactive mode. This pulse rate was acceptable because as long as these cubs were with their mother and on active mode, the mothers' col.lars could be used for radio-tracking. These collars were placed on 6 cubs in 3 litters in 1983 (females 281, 283 and 299) and on 7 cubs in 4 litters in 1984 (females 340, 337, 423, and 281). Seven of these 13 cubs survived to their yearling year ( $46 \%$ mortality). Cause of death for 5 cubs was determined to be predation by unknown brown bears. Cause of death for the remaining 2 cubs was not determined as the
bodies could not be found when their radio-signals disappeared. I suspect that these cubs were either drowned and swept downriver during river crossings or that they were preyed upon and their transmitters destroyed. In one of these cases of unknown cause of mortality, the lost cub was markedly the smallest cub in a litter of 4 (with female 423); the other 3 cubs survived.

It is noteworthy that 4 of the lost radio-marked cubs were with female 281 who had litters of 2 newborns in 1983 and in 1984. In both years this female left her high-elevation den site and moved to lower elevations along the Susitna River early in the year, following the typical pattern for bears not accompanied by newborn cubs. In both years she lost her cubs ( 3 to brown bear predation, 1 to cause unknown) within days of moving to lower elevations (cubs were lost on 1 June in 1983 and on about 28 May 1984). This was a young female that had her first litter in 1983. In 1985 she had another litter of 2 and followed the same pattern of moving to lower elevations; this time she lost one of her cubs between 5 June and 26 June; the other survived through September 1986.

An additional 2 cubs were radio-marked with female 388 in 1984. This capture resulted in a capture-induced separation which ended in the death of the cubs despite 3 efforts we made to reunite this family. Separation occurred on 16 May and reunion efforts occurred on 18,23 and 24 May. In the first effort we herded the female toward the cubs with a helicopter. In the second we air-dropped the cubs about 10 feet from a helicopter near the female. In the third effort we immobilized the female with Sernylan and released the cubs nearby; the cubs began to nurse immediately. At this last effort 1 cub had a nose full of porcupine quills which we pulled. One cub died on 29 May, most likely of starvation. Nearby feces of the other cub were full of overwintered Empetrum berries. The other cub survived until mid-June at least; its collar was picked up on 23 June but no sign of the cub was found nearby. This collar was unexpanded, evidence indicating the cub was killed by a predator rather than having shed the collar. On other occasions reunion efforts like those described above were successful. The lack of success in this case may have resulted from the delay in attempting the reunion; the female may have physiologically changed from a lactating mode to an estrous mode. She was seen with another large bear on 3 June and with a known male on 7 June and she had cubs again the following year.

Thirty-six litters of yearling cubs observed following emergence from dens averaged 1.7 offspring (range $=1-3$ ) (Table 16). Eight of 37 yearlings (21.6\%) were lost before their mothers emerged from their dens in the following year
(Table 16). I suspect most or all of these were mortalities but it is possible that some of the yearlings defined as "lost" may have separated from their radio-marked mothers as yearlings. None of the bears defined as "lost" as yearlings have subsequently appeared in the hunter harvest.

Implant transmitters were surgically implanted in 6 yearlings (in 3 litters) in an effort to determine causes of mortality. Only 1 of these bears died before transmitter failure the following year; the body of this bear could not be found to determine the cause of death as a fox carried the transmitter away from the carcass (determined from tooth marks on the transmitter). Causes of yearling mortality are largely unknown, but Dean et al. (1986) documented 2 instances in Denali National Park where yearlings were killed by adult .males.

Twenty litters of 2-year-old offspring averaged 1.7 offspring (range $=1-3$ ) (Table 17). All but 1 of these litters separated from their mothers prior to den entrance the following fall. Female 337 may prove to be an exception, as she still had her 2-year-olds when last seen on 24 September 1986 (Table 17). Separation from the mother at age 2 was defined as "weaning."

Reproductive histories of individual females are given in Table 18. A summary of losses of cubs and yearlings in these litters is given by year in Table 19. Measurements of cubs and yearlings handled in this study are given in Tables 20 and 21.

## 7.G.1.b. Reproductive Interval

There are numerous ways to calculate reproductive interval. The interval between successive production of litters of newborn cubs is not a good statistic.because complete loss of a litter of cubs would frequently yield an interval of 1 year. Inclusion of such intervals in a calculation of mean reproductive interval would underestimate the interval that is needed to calculate population growth rate. The best interval to use would be the interval between successive successful separations ("weanings") of offspring from their mothers; however this method requires many years of data. Reproductive histories for individual radio-marked female brown bears are given in Table 22. Reproductive status of bears was determined during visual observation of radio-located females.

Reynolds and Hechtel (1985) defined reproductive interval as the period between successful breeding (as evidenced by cub production the following year) and the next successful separation of mother and offspring ("weaning"). Their method
provides intervals that are 1 year longer than the one used in this study. I defined reproductive interval as the interval between production of a litter (as evidenced by observation of that litter following emergence from the den) and the next successful weaning of a litter. This interval definition will be shorter than that used by Reynolds and Hechtel (1985), as our definition does not include years of apparent conception failure unless these instances occurred subsequent to a successful weaning. With my definition $I$ was able to include intervals for those females initially captured in the spring and accompanied by yearling offspring (back-dated to the year these yearlings were born); these intervals will be biased toward short intervals as litters could have been lost prior to the litter first observed as yearlings. We defined successful separation as occurring when 2 -year-olds separated from their mothers after den exit (no cases of females entering dens with 2 -year-old offspring were observed although female 337 still had her 2-year-old offspring with her in September 1986).

Following this definition $I$ observed 17 reproductive intervals; 14 of these were 3 years (Table 22). The year in which 1 capture-related loss of a cub litter occurred (388 in 1984 [Table 22]) was not counted. Intervals of longer than 3 years were observed in 3 cases. In all of these, intermediate litters were completely lost in the year of their birth or in the following year (Table 22). Of these intervals, 1 was 4 years, 1 was 5 years, and 1 was 6 years. The mean reproductive interval for these 17 cases was 3.4 years (Table 22).

This estimate of mean reproductive interval is an underestimate as it is biased toward 3-year intervals, the minimum possible in natural conditions (Bunnell and Tait 1985). This bias results from shortness of the study period, losses of radio-marked bears, and back-dating from litters first observed as yearlings. For example, 5 females would have had intervals >3 years. These intervals were not counted because a complete interval, according to the above definition, was not obtained. Failure to complete these intervals resulted because the study ended, because the bear was shot by a hunter, or because the radio transmitter failed before the interval was completed. These incomplete intervals resulted from complete loss of a litter; the intervals would have been at least 4-7 years in different cases (Table 22). If the minimum values for these incomplete intervals are included, the estimated mean interval for 17 complete and 5 incomplete intervals would be 3.8 years (Table 22). This is still an underestimate as minimum possible values were used for incomplete intervals (396, for example, lost litters of newborns in 1984 and 1985, and was alone in 1986; the minimum interval of 6 years was obtained for her by assuming she will have cubs in 1987 and will successfully wean this litter in 1989) .

Other methods of calculating reproductive interval are possible. The interval from birth of a litter which was successfully weaned and birth of the next litter was observed for 3 cases (312, 337 , and 420); all of these intervals were 3 years (Table 18). The interval between successful weaning of 1 litter and successful weaning of the next litter was observed in 1 case (337); in another case (388) this interval should be completed in 1987. In both cases the interval was (or will be) 3 years (Table 18). As above, these intervals are biased toward the short intervals by the limited period of study.

## 7.G.1.c. Age at First Reproduction

Ages used in calculating age at first reproduction were estimated from counting cementum lines in a sectioned and stained premolar extracted during tagging. Some error in these estimates (probably nonsystematic) is likely. Age at first successful breeding is 1 year less than the age at first litter production.

As with reproductive interval, age at first reproduction (defined as production of a litter seen at emergence from natal den, not as breeding activity) can be calculated in different ways. The best way is to annually observe bears from immaturity through the time they are seen with litters. This is difficult because: 1) problems exist with attaching transmitters to subadults; 2) it requires long-term studies; and 3) it requires not utilizing data from other sources. Four bears aged as subadults when originally captured were followed to production of their first litter; all first produced cubs at age 6 (Table 18). Another bear in this category (407 at age 8) produced no litters I could see when she was age 4 through age 7 (Table 23). The earliest 407 could produce a litter would be in 1987 when she will be age 8. For these 5 bears, mean age at first reproduction (including 407) averaged 6.4 years (Table 23).

Young adults accompanied by cub, yearling, or 2-year-old offspring when first captured, can be back-dated to determine their mother's age at the time that litter was born (data in Table 22). With these data there is no way of knowing for certain. whether a litter was previously produced and lost. This source of error would yield overestimates of age of first litter production. Using such back-dated data, I calculated that 4 bears produced their first observed litters at age 4; 4 at age 5; 4 at age 6; and 1 at age 7 (Table 23). For these 13 bears, apparent age at first reproduction averaged 5.2 years. These data were back-dated from newborn cub litters ( $\mathrm{N}=4$ ), from yearling litters ( $\mathrm{N}=7$ ), and from litters with 2-year-old offspring $(N=2)$ (Table 23). No back-dating of
litters to determine mothers' age when the litter was born was included for bears aged $\geq 8$ years old when first captured. Such bears had a high likelihood of having had litters prior to the one they had when first captured.

When these two data sets are combined, an estimate of 5.5 years was obtained for average age of females producing first litters ( $N=18$ female brown bears; range 4-8) (Table 23). This is not the same as mean age at first reproduction, as this statistic is based on the proportion within each age class producting first litters. The frequency distribution for these combined data shows that age 6 is the most common age for production of first litter (44\%) (Table 23).
7.G.2. Sources of brown bear mortality

The Su-Hydro study area is in Game Management Unit 13. Since 1980 brown bear hunting regulations have been liberalized in GMU 13 in an effort to increase bear harvests, and thereby, to accelerate moose population growth. These changes have increased reported bear harvests in the study area to an average of 32 bears/year in 1983-85 compared with 14.3 in the period 1.978-80 (Table 24). In Table 24, harvests in the Su-Hydro study area are compared with harvests in the Denali Highway areas used for comparison. The locations of the areas used in these comparisons are illustrated in Fig. 16. Harvests along the Denali Highway have been relatively constant since 1980 although harvests have doubled in the Su-Hydro area (Table 24).

Frequency with which marked bears are taken by hunters is an index to harvest effort. Data on hunter kills of bears marked during the period 1978-1986 are presented in Tables 25-27, and summarized in Table 28. Percentage values in Tables 25-27 are underestimates because there are unrecorded natural mortalities of marked bears and because some marked bears are not recognized as marked during the sealing process. The percentage values are not harvest rates of the whole population because cub and yearling bears which compose a large proportion of the bear population were not considered.part of the marked population.

The minimum percentage of marked bears shot in a year in the Su-Hydro area varied from $3 \%$ to $15 \%$ (Table 28). This is an underestimate because it assumes no natural mortalities or failure to recognize marks when bears are sealed. A more probable estimate, based on bears known to be alive and including bears suspected (not just known) to have been shot, was 4\%-22\% (Table 27). Frequency with which marked bears are shot has increased in recent years (Table 27). This is in line with increasing harvests of bears in the study area as discussed above (Table 24).

Three cases of apparent natural mortalities of adult radio-marked brown bears were observed during the course of this study. These instances are described in Table 29. Mortality rates for subadult brown bears are discussed in Section 7.G.1.a of this report.
7.G.3. Brown bear movements
7.G.3.a. Home range size

Home range was calculated using the standard minimum grid described by Mohr (1947). Data for individual bears in inclividual years and for all years lumped are given in Table 30; these data are summarized by sex; age and reproductive status in Table 31. When years are lumped for individuals with more than 1 year's data, home ranges averaged $1,022 \mathrm{~km}^{2}$ $\left(1941 \mathrm{~km}^{2}\right.$ for males and $50.1 \mathrm{~km}^{2}$ for females) (Table 31). Home range variances determined by standard minimum grids were large (Table 31). Males' home ranges varied little between years while home ranges for females without newborn cubs varied more (Fig. 17).
7.G.3.b. Movements to hunting and fishing areas

Peak of caribou calving occurs 20-25 May for the Nelchina herd, but calves can be born through 15 June. The main caribou calving area used by Nelchina caribou during the period of this study was between Kosina Creek and the Oshetna River (Pitcher, in press). This area is southeast of the largest part of the Watana Impoundment and outside the home ranges of most radio-marked bears. For this reason, movements of bears to the caribou calving area at the time caribou calves are available can reasonably be interpreted as movements motivated by intent to prey on caribou calves. Murie (1981:173) noted that although grizzlies could catch some calves, "... [I] noted no special movement of bears into a calving area for the purpose of preying on calves." Murie suggested that such movements could occur for some bears in circumstances where calving is concentrated. Reynolds and Garner (in press) noted such movements on Alaska's north slope. Histories of individual bears that made such movements are given below.

Brown bear female 340 (age 3 in 1981 when first captured) was intensively monitored in spring 1981. Until 14 June, she lived in the Deadman Creek-Watana Creek area; on 15-16 June she moved to the Clarence Lake area and then returned. This movement was not classified as related to caribou predation because it occurred 2-3 weeks after the peak of caribou calving. In late May 1982 this bear moved into the Kosina Creek calving area, returning by 9 June. Between 15 May and

23 May 1983, this bear was twice located in caribou calving areas on lower Kosina Creek In 1984 this bear had newborn cubs and was again intensively monitored in the spring (starting 28 May), but no movements to caribou calving areas were documented. In 1985, with yearling offspring, she was in the caribou calving areas on 23 May (no locations were made between 16 May and 23 May). On 24 May 1986 this bear (without offspring) was again located on Gilbert Creek in the midst of the caribou calving area, and although a kill was not seen, blood was seen on the snow near her. Except during the caribou calving period, this bear was never found south of an east-west line through Watana Mountain. I conclude that this bear regularly, probably annually, moved to caribou calving areas to prey on caribou.

Female brown bear 331, age 6 when captured in 1981 with two 2-year-old offspring, weaned her young after 15 May. She was next seen on 15 June in the upper Oshetna River country where . she remained until the end of June when she returned to her normal home range along Tsusena Creek (Fig. 18). This bear made no similar movements in spring 1982 although she left her home range after 29 June and in mid-August was found dead on Tsisi Creek, of unknown causes. I considered the movement in 1981 a movement to the caribou calving grounds.

Male 280; age 5 in 1980, was originally captured in the upper Kosina caribou calving grounds in early May 1980. Subsequently, most of its movements were between Tsisi Creek and upper Watana Creek except on 16 May, 1983, when it moved to the caribou calving area around Gilbert Creek, and in early June 1984, when it was around Clarence Lake. I considered these movements probable forays into the caribou calving area.

Movements into caribou calving areas (less clearly motivated by predation) were made by bears $293,381,384$, and 299. These bears all had year-round home ranges near or overlapping the caribou calving area.

There are only a few instances of clearly defined movements to caribou calving grounds in the Su-Hydro study area. When such movements occurred, bears typically spent little time in these calving areas. These data suggest that the impoundments' blockage of bear movements to caribou calving areas is likely to have little impact on bear nutrition. It is possible that Su-Hydro area bears are little motivated to move very far to caribou calving grounds because numerous moose calves are equally good prey and these can be found within their annual home ranges (Section 7.G.4, this report).
7.G.3. Brown bear dispersal

The pattern for brown bears in the Su-Hydro. area is for subadult males to disperse from maternal home ranges as 2- or 3 -year-olds. Female subadults typically set up home ranges within their maternal home ranges. Subadult dispersal was studied using drop-off radio collars and surgically implanted transmitters.

One male (342) dispersed as a 2-year-old from the Watana dam site to the Kashwitna River in 1981 (Fig. 19). This dispersal was in a southwesterly direction and covered, in a direct line measurement, a distance of about 120 km . In subsequent years this bear gradually worked his way back toward the study area and was last found on Prairie Creek in July 1984.

Two 2-year-old sibling males (391 and 392) dispersed about 70 km in a northeasterly direction from their maternal home range following weaning in spring 1983. They stayed together until just prior to den entrance. Another bear thought to be a female sibling of these bears (393) remained near her maternal home range (Fig. 20).

A different pattern was found for 2 male 2-year-old siblings in spring 1983. One male (389) dispersed about 80 km in an easterly direction following weaning while the other (390) remained within the maternal home range at least until the following spring (Fig. 2i).

Another 2-year-old male (386) dispersed in a northerly direction from its maternal home range in spring 1983. The dispersal distance was approximately 52 km (Fig. 22).

These movements suggest that the Su-Hydro study area is a source of recruits through emigration to surrounding areas. There is evidence as well that subadults from surrounding areas immigrate to the Su-Hydro area. Male 214 was originally tagged as a 2-year-old during earlier studies in 1978. The tagging location was north of the Denali Highway on Valdez Creek. In spring 1980 this bear was recaptured near Clarence Creek (between Vee Canyon and Jay Creek). A similar pattern was observed for female 273, originally captured and transplanted from north of the Denali Highway in 1979 as a 3-year-old. This bear returned to its capture site (Miller and Ballard 1982b), but was recaptured in the middle of the Su-Hydro study area in 1985.

I suspect that reduction of brown bear carrying capacity in the Su-Hydro area will likely decrease the number of emigrants available for dispersal to surrounding areas as a result of lowered productivity. I also suspect that survivorship of
immigrants to the Su-Hydro area will be lowered as a result of the anticipated decline in carrying capacity resulting from the proposed project.

## 7.G.4. Brown bear predation on ungulates

Earlier studies have shown that brown bears are significant predators on newborn calves in Game Management Unit 13 (Ballard et al. 1981 and 1985). Black bears were also shown to be important predators on moose calves on the Kenai Peninsula (Franzmann et al. 1980). Just north of the Alaska Range, in Unit 20, wolf predation was shown to limit predation in a system where bears are rare (Gasaway et al. 1983, Ballard and Larson, in press). Previous studies on predation by bears have not been conducted in an area, such as the Su-Hydro location, where each of these 3 predator species is abundant. Our predation studies were initiated in an effort to better understand the dynamics of predation on moose in a system that includes all 3 predators. The information obtained can be used to test hypotheses about the effects, on predators and on prey, of impoundment-related impacts which alter predator-prey ratios.

Brown bear predation on ungulates was evaluated by intensive monitoring of radio-marked bears. Intensive monitoring was conducted on 21 May-23 June 1981 (Miller and McAllister 1982), on 28 May-7 June 1984, and on 29 May-1 August 1984 (Miller 1985a). Monitoring was done once per day except during 29 May through 7 June 1984 when bears were monitored twice per day. Coordinated studies of causes of mortality of radio-marked moose calves were conducted in spring 1984 (Ballard et al. 1985). These studies were similar to those conducted in 1978 and 1979 near the headwaters of the Susitna River and elsewhere in Game Management Unit 13 (Ballard et al. 1981). Papers on these data are in preparation (Ballard and Miller, in prep., and Ballard et al., in prep.).

Results from intensive monitoring of brown bears during spring studies are presented in Table 32. For the purposes of these analyses, "consecutive observation days" summed all days in periods of $>2$ consecutive days when a radio-marked bear was seen at least once.

In 1978 spring predation rates were 1 kill/4.9 consecutive observation days or 1 moose calf kill/8.4 consecutive observation days (Table 32) (Ballard et al., in prep.). In our spring 1981 and 1984 studies, observed kills were less frequent: 1 kill/7.5 consecutive observation days and 1 moose calf kill/l1.8 consecutive observation days (Table 32). Rates of loss of radio-marked moose calves to brown bear predation was similar in the 1977-1978 Unit 13 studies and in the 1984

Su-Hydro studies (Ballard et al. 1985). In both studies predation accounted for $86 \%$ of natural mortalities, with brown bears responsible for $65 \%$ of mortalities in 1984 and $79 \%$ in the earlier studies (Ballard et al. 1985). Of predatorrelated mortalities, brown bears accounted for 75\% in 1984 compared with 91\% in 1977-78 (Ballard et al. 1985).

Uniike these earlier studies, the Su-Hydro studies were undertaken in an area where black bears were abundant. Here black bears accounted for $12.5 \%$ of predator-related deaths in 1984 (Ballard et al. 1985). In 1984, then, black and brown bears were responsible for $87.5 \%$ of predator-related deaths, almost equal to the 1977-78 figure of 91\%. In both studies moose calf losses were largely confined to the 6 weeks following birth. In the Su-Hydro studies, predation was much lower during late July through August, 1984 (Table 33).

In the 1978 studies significant differences could not be detected between bear predation rates (on ungulates), based on sex or reproductive status categories, but it was suspected that female bears accompanied by offspring older than 1.0 years could have higher predation rates than other bears (Spraker et al. 1981). Predation rates (all known and probable kills of ungulates throughout a year) based on all visual observations during radio-tracking (except those at den sites) for radio-marked bears from 1978 through 1985 are presented in Table 34. For these analyses the presence of a bear on a kill was assumed to reflect predation. This assumption is biased to the degree that bears usurp kills made by other species, or other bears; or scavenge natural mortalities.

Chi-square analyses indicate no differences between sex and reproductive status groups in the 1978 studies ( $P<0.10$ ). No differences in observed predation rates were observed between males and females in 1978, in 1981 and 1984 combined, or in combined results ( $\mathrm{P}>0.10$ ). Neither were there significant differences in predation rates between females with yearling offspring and females without offspring (includes those with 2 -year-olds in early spring) in either study or in combined results ( $\mathrm{P}>0.10$ ). In combined data from these 2 studies, females with newborn cubs had lower predation rates than either females without offspring or females with yearling offspring ( $P$ < 0.05). In the Su-Hydro data ("area 1"), females with newborn offspring had significantly lower predation rates than females with yearlings ( $\mathrm{P}<0.05$ ) but not lower than rates for females without offspring ( $P>0.05$ ). These analyses support the conclusions that females with newborn cubs tend to have lower predation rates on ungulates (moose and caribou) than other bears, and that all other brown bear categories, based on sex or reproductive status, have
similar predation rates. Similar analyses were done for observations of brown bears on moose calf kills (Table 34). Again, there were no differences between male and female predation rates ( $\mathrm{P}>0.10$ ) or between females with yearlings and females without offspring ( $P>0.05$ ). Females with newborn cubs, again, had lower predation rates than either single females or females with yearling offspring ( $P<0.05$ ).

The lower predation rates observed for females with newborn cubs probably reflect the geographic separation of this group from prey concentrations (see Section 7.B, this report). Females with newborn cubs tend to remain at higher elevations near their den sites for $3-8$ weeks longer than other bears (including years when the same females have older offspring or no offspring). Moose calve at lower elevations where they are available to bears that move down in the spring in the typical pattern, but not to the bears that remain at higher elevations. This behavior pattern by females with newborn cubs may minimize predation on cubs by other bears; some females, such as 281 and 396 , which did not follow this pattern, had especially high rates of cub loss (Section 7.G.1, this report).

During intensive monitoring in spring 1981 and 1984 we saw radio-marked brown bears on 25.5 moose calf kills during a total of 302 consecutive observation-days (Table 32) (half kills result from joint occupancy, with another predator, of a kill site). This provides a minimum estimate of predation rate ( 1 calf kill/11.8 consecutive observation days) because unobserved kills could easily occur between observations and because kills cannot always be seen or identified. Regardless, this estimate can be combined with other data to estimate the total number of moose calves killed by brown bears in the study area.

If all predation on moose calves occurred during a 6-week period in the spring, at an average rate of 1 kill/11.8. days, an average bear would kill 3.6 calves. If, as estimated in Section 7. A of this report, there are 327 brown bears in the impoundment impact zone and $32 \%$ of these are cubs and yearlings (Miller et al., in press), then there are about 222 brown bears age 2 or older in the study area. At the above predation rate these bears would kill 799 moose calves/year. Similar estimates were independently derived from models of moose populations (Ballard et al., 1984).
7.G.5. Brown bear denning ecology

Den sites of radio-marked brown bears were located during winters of 1980-81 through 1984-85. Dens were initially located from fixed-wing aircraft and most dens were visited on the ground in May or June following bears' emergence from
dens. During these visits dens were measured, and slope, aspect, and other characteristics recorded when possible. These measurements have been described by Schwartz et al. (in press). Dens were frequently collapsed when visited in the spring; interior measurements were impossible in these cases. In some cases where dens were collapsed, the den site was not physically visited and slope, aspect, and elevation were recorded from a helicopter hovering at the den site. Some data were also collected from dens made by bears that were not radio-marked; these dens were spotted during aerial tracking flights.
7.G.5.a. Den entrance and emergence dates

Entrance and emergence dates were estimated from the radio telemetry data in 3 ways. For entrance dates, the last time a bear was seen outside its den was considered the minimum (earliest) entrance date and the first time a bear was found in its den was considered the maximum (latest) possible entrance date. The midpoint between these 2 dates was considered the "most likely" entrance date for use in calculating means. Similar procedures were followed for den exit dates. The maximum period a bear spent in its den was the period between its minimum entrance date and maximum exit date; the minimum period was that between its maximum entrance date and minimum exit date. The midpoint for period spent in the den was that period between the "most likely" entrance and exit dates. Data on entrance and exit dates for each radio-marked bear for each year of the study are provided in Tables 35-39.

Based on most likely dates, the earliest den entrance was 24 September (pregnant female 313 in 1980) and the latest was 10 November (male 400 in 1984). The average most likely entrance date varied from 6 to 18 October in different years (Tables 35-39).

The earliest den exit date based on "most likely" calculations was 11 April (for downstream females 379 and 403 in 1984) and the latest exit date was 28 May (for female 388 with newborn cubs in 1985). The average most likely exit date varied from 23 April in 1980 to 10 May in 1985. Heavy spring snowfall was thought to delay den exit for brown bears in spring 1985. Available data on snow conditions are based on once-a-month readings of 4 snow stations in the impoundment vicinity by the U.S. Soil Conservation Service. These data (illustrated in Fig. 23) are inadequate to document the abnormally late and heavy snow conditions in spring 1985 but these conditions were evident to me.

Using the most likely dates for den entrance and emergence, average number of days spent in dens varied from 187 in 1980-81 to 214 in 1981-82 (Tables 35-39). Using these most likely dates, I calculated the average time spent in dens for 74 bear-years during the study to be 201 days (S.D. = 16.6).
7.G.5.b. Characteristics of dens

Measurements, and other characteristics of 96 brown bear dens for which some data are available, are presented in Table 40. Only 2 dens were in natural cavities and one of these was partially excavated. Dug dens totaled 75; undetermined cavity types totaled 19 (Table 40). Dug dens predominated in dens on Kodiak Island examined by Lentfer et al. (1972), and natural cavity dens were more common in parts of southeastern Alaska (Schoen et al., in. press) and northern Alaska (Reynolds et al. 1976).

Brown bear den sites were found on all aspects, but dens on south aspects were approximately twice as common as on any other aspect (Fig. 24). South aspects seemed to be more strongly selected by females who were pregnant at den entrance than for females who were not, or for males (Fig. 25).

No brown bear den sites were found in the area that would be inundated by either of the proposed impoundments. Elevations of den sites in the upstream study area ranged from 2010 to 5330 feet (Table 41): The lowest den site would have been inundated if it had been in the vicinity of the Watana Impoundment but it was in the vicinity of the lower, Devils Canyon, impoundment. This den site, that of pregnant female 396, was so atypical for a brown bear that I initially thought it represented a shed collar or dead bear rather than a den site. This female lost her litter of newborn cubs shortly after emergence from this den. Den sites were lower in the downstream study area (Table 41) where higher elevations were not as available to bears.

Locations of den sites in upstream and downstream study areas are illustrated in Fig. 26 and Fig. 27. The impoundment itself will likely have little impact on brown bear denning habitat but winter activities along the access road, borrow sites, and other construction areas that occur in brown bear denning habitat could disturb denning bears. Reynolds et al. (in press) observed responses in denning bears to disturbances within 1.6 km and suggested rerouting aircraft and other disturbances away from known den sites during denning. I found no evidence that availability of denning habitat was a limiting factor for brown bears in the study area. Bears may be able to find adequate den sites away from the source of disturbance. If disturbance causes bears to abandon dens
after the period of den entrance, however, these bears may find it very difficult to find and dig dens in alternative areas when the soils are frozen.

Most bears showed a tendency to den in the same general location year after year but considerable variation was observed. Den sites used in different years by the same individual were separated by a mean distance of 3.8 miles (Table 42). One bear, male 400, moved from his spring home range near Watana Creek to den sites north of the Denali Highway on the upper McLaren River in 3 successive winters. There could be strong selective pressures on bears to return to areas that are known, based on previous experience, to be good denning areas, rather than risk denning in an area with equivalent characteristics but where an individual had no previous experience. Good sites are those where wind currents assure that the den entrance will be well-sealed with deep snow and where soil and permafrost characteristics are such that dug dens are unlikely to collapse during the winter.

## 8. Black Bear Results

8.A. Number of black bears in impoundment impact zone

In part 9 of this report I derived an estimate of the number of bears in the impoundment impact zone. This estimate was based on extrapolation to black bear habitat in the entire zone from a density estimate ( 8.97 bears/100 $\mathrm{km}^{2}$ ) obtained in part of this zone. The 95\% confidence interval for this density estimate was similarly extrapolated to the impact zone without modifications designed to reflect the extrapolation. The area defined as black bear habitat (1191 $\mathrm{km}^{2}$ ) was determined by drawing a line around point locations of radio-marked bears (Section 6.B of this report). The resulting estimate was 107 black bears (95\% CI = 93-122). I estimated that $35 \%$ of these bears were cubs and yearlings (Miller et al., in press; see Appendix 2). This estimate was lower than earlier estimates I made for this area based on a rough density estimate of 24 bears/100 $\mathrm{km}^{2}$ (Miller and McAllister 1982), perhaps because the population declined significantly during the course of this study. This decline may have resulted from the poor berry crop in 1981 (Miller 1983, 1984, and 1985a).

Because the impact zones of each impoundment overlap, over half of the estimated population in the 2 -impoundment area would be in the impact zone of either impoundment considered separately. However, it is difficult to estimate the size of the zone of overlap. In order to divide the whole study area into impact areas for each impoundment a line between the impoundments was drawn. This was a north-south line through
the confluence of Tsusena Creek and the Susitna River (this location is about 2.5 miles downstream from the Watana dam site). Within the area defined as black bear habitat (Fig. 7), the area east of this line ( $658 \mathrm{~km}^{2}$ ) was defined as the area inhabited by the Watana Dam population of black bears, and the area west of this line ( $533 \mathrm{~km}^{2}$ ), as the area inhabited by the Devils Canyon population. At the aboveestimated density the Watana Dam population would then have had 59 black bears (51-67), and the Devils Canyon population 48 (42-55).
8.B. Black Bear Use of Impoundment Proximity Zones
8.B.1. Levels and seasons of use.

Black bear use of nested zones of proximity to the Devils Canyon and Watana Impoundments was analyzed using the same methods and procedures previously discussed for brown bears (see Section 7.B of this report and Miller and McAllister 1982). In this analysis relocations of radio-marked bears were allocated to 1 of 4 zones: within the area that would be flooded (zone 1), from the impoundment high water line to 1 mile from this line (zone 2), from 1 to 5 miles from the high water line (zone 3), and more than 5 miles from the high water line (zone 4). Use of these 4 zones for each month for the impoundment zones of each proposed impoundment is illustrated in Fig. 28. Monthly percentage use of the area to be flooded (zone 1) is higher for the Watana Impoundment zone than for the Devils Canyon zone (Fig. 28).

Black bear use of the areas that would be inundated by the Watana Impoundment was highly significant when compared with the adjacent zone or the 2 adjacent zones (Table 43). Overall, $42 \%$ of the observations of radio-marked black bears made in the vicinity of the Watana Impoundment were in the area that would be inundated by that dam (Table 43). This percentage value was highest in May and June ( $52 \%$ and $46 \%$, respectively), the same time period when brown bear use of the impoundment area was highest (Fig. 11). No doubt at this time the black bears and brown bears are using the same spring food resources that are available earliest on the south-facing slopes along the Susitna River and its tributaries: carrion, newly-emerged plants, overwintered berries, and moose calves.

This same pattern is not evident for the Devils Canyon Impoundment. This is probably because of the very small area that would be inundated by this impoundment (only $3.3 \%$ of the area within 5 miles of the Susitna River along the section of the river that would be inundated by the Devils Canyon Impoundment) (Table 44). In the spring period when the Devils Canyon Impoundment zone is most used (May 1-June 30), observed
use was lower than expected values for zone 1 , for the comparison between zones 1 and 2 (Table 44). In the area around the Devils Canyon Impoundment the distribution of acceptable black bear habitat is much wider than farther upstream and as a result, dependence of bears on the habitat in the immediate vicinity of the river is less in the lower portion of the study area.

## 8.B.2. Prediction of impacts

Reductions in black bear populations, resulting from habitat loss, were estimated for black bears in the same manner as for brown bears (see Section 7.B.2). Rather than using just spring data, however, data on annual use were used for the black bear analysis because less seasonal variation in use of the impoundment zone was evident for black bears than for brown bears (Figs. 11 and 28).

Radiotelemetry data for 17 male and 14 female black bears using the Watana Impoundment impact area show that $43 \%$ of all point locations were within the zone that would be inundated; an additional $36 \%$ were within 1 mile of the impoundment shoreline (Table 45). Under the assumptions used for these analyses (Section 7.B.2), I estimate that the carrying capacity for the estimated Watana population of 59 black bears would be reduced by $43 \%$ due to habitat inundation; this is a reduction of 26 bears.

Radiotelemetry data for 9 male and 10 female black bears using the Devils Canyon Impoundment impact area show that only $3 \%$ of point locations were within the zone that would be inundated, and an additional 43\% were within 1 mile of the impoundment shoreline (Figure 45). Under the assumptions used in this analysis, the carrying capacity of Devils Canyon's estimated population of 48 black bears would be reduced by $3 \%$ due to habitat inundation, this is a reduction from existing numbers, of only 2 bears (existing numbers are not necessarily at carrying capacity, however).

Considering both impoundments together, $30 \%$ of point locations were within the area that would be inundated by one of the impoundments (Table 45). Using this value; I estimated that the carrying capacity of the whole study area's population of 107 black bears would be reduced by 32 bears. This estimate is close to that obtained by summing the values for each impoundment separately (28 bears).

Of the 31 bears used for the Watana Impoundment analysis, 24 (77\%) had point locations within the area that would be inundated by the proposed impoundment (Table 45). Of the 19 bears used for the Devils Canyon Impoundment analysis, 8 (42\%)
had point locations within the area that would be inundated by this impoundment (Table 45). These data.may indicate that inundation by the impoundments could result in a more severe decline in availability of bear habitat than I estimated above (using the proportion of point locations in the impoundment zone).
8.B.3. Mitigation measures

As with brown bears, potential measures to mitigate for loss of black bear habitat resulting from inundation are limited. Possibilities include:

1. Increasing the abundance of foods used by black bears throughout the year; or
2. Indirect mitigation (out-of-kind substitution of other benefits for the resources, for bears, that are lost as a result of the project).

One of the reasons black bears may utilize so little of the habitat available in the study area, compared with brown bears, may be competitive exclusion of black bears by brown bears. To the degree that this is a factor, the anticipated reduction in brown bear numbers through habitat loss and displacement disturbance may make more habitat available for black bears. Although this is possible, I consider it unlikely, as in most cases, I suspect that black bears' recognition of acceptable black bear habitat is genetically based (most black bears are unlikely to venture into more open areas even if brown bears are not present).

Prairie Creek may be an exception to this rule. Black bears make only slight utilization of Prairie Creek salmon resources. This is probably because of competitive exclusion by the many brown bears utilizing the area. If, as anticipated (see Section 7.D of this report), brown bear use of Prairie Creek greatly declines because of displacement disturbance caused by humans, I would expect that black bears would exhibit increased utilization of Prairie Creek. This is because black bears are more tolerant of humans than brown bears are and because humans are more tolerant of black bears than they are of brown bears. Prairie Creek is in a forested area that, except for the presence of brown bears, seems to be good habitat for black bears.

## 8.C. Other Impacts

8.C.1. Berry-foraging areas

In the 6-8 weeks prior to denning, berries constitute a highly important source of food for bears. Berries are highly
digestible and easily converted to fat (Bunnell and Hamilton 1983; Bunnell, in press) and therefore they are particularly appropriate foods for the period of hyperphagia prior to den entrance (Nelson et al., in press). In the upstream study area the most abundant and important berry for bears of both species is probably blueberry (Vaccinium uliginosum). Lowbush cranberry (V. vitis-idaea) is also abundant in the upstream study area. In the downstream area devils club (Oplopanax horridus) is heavily utilized (Section 8.E of this report). Based on scats collected in the early spring, overwintered berries (especially crowberries, Empetrum nigrum) appear to be important foods in spring as well (Sections 8.E and 8.G.4). During August, movements of black bears become more extensive and many bears travel to habitats little utilized at other times of the year. These habitats are the semi-open shrublands adjacent to the spruce forests.

During years of berry crop failure, such as in 1981, movements of some bears may become much more extensive and include utilization of very open habitats distant from forests that are more typically utilized by brown bears (Section 8.G.3, this report).

The limited data we gathered during 1982-1984 on berry abundance in these shrublands is consistent with a hypothesis that blueberries are more abundant in this habitat than in the adjacent spruce forest where bears spend most of their time during the rest of the year (Section 8.G.4.b). Information on abundance of berries and berry-producing bushes is presented in Section 10 of this report.

These shrubland sites used in late summer by black bears foraging for berries are the favored sites for construction camps, borrow areas, and permanent residences. The area between Tsusena Creek and Deadman Creek will be especially heavily affected by these activities as this is a highly favored foraging area for black bears during late summer. Although black bears are not as prone to disturbance displacement resulting from these activities as brown bears, it is likely that black bears will come into conflict with man in these sites.

## 8.C.2. Blockage of movements

As discussed previously for brown bears (Section 7.C), black bears swim readily and are known to swim across impoundments. Movements across the reservoir will probably be restrained to some degree, relative to movements bears currently make across the river. Simpson (1986:21) studied movements of grizzly
bears in the vicinity of the Revelstoke Reservoir in British Columbia and noted that "grizzlies would cross a river but not the reservoir." Relative to this same reservoir, Richard L. Bonar (18 April, 1985, interview transcribed by Bill Steigers of the Susitna Project Group of LGL) noted "...the radio-collared bears [both species] haven't crossed as often as they did before the water came up."

Although some impact is probable, it is impossible to guess how much movements across the river will be restrained by the Susitna impoundments. Movements across impoundments are not the only movements that may be inhibited. Black bears frequently make extensive seasonal movements both up and down the river and, unlike brown bears, these movements occur largely in and along the forested corridor of the Susitna River. Following flooding of the impoundment, such movements will require crossing or circling around inundated tributaries. The greatest barrier to these movements following filling of the reservoir will be the large bay at what is now Watana Creek.

In this study I concentrated on documenting frequency of crossing so that these data from the preconstruction phase could be compared with data collected during a postconstruction study. Such comparisons will permit more accurate predictions of impacts in future impact assessment studies.

The number of river crossings for each radio-marked bear in each year with $>5$ non-den observations varied from 0 to 12 (Table 46). For purposes of this analysis, a "bear-year" was defined as a year in which a radio-marked bear received more than 5 radio locations lexcluding observations at its den site). For males, crossings were observed for 36 of 56 bear-years ( $64 \%$ ); for females crossings were observed for 18 of 57 bear-years (32\%) (Table 46). The average number of crossings for males that crossed was 3.3; for females it was 3.8 crossings (data in Table 46).
8.C.3. Mitigative measures

The potential methods of mitigating for loss of berry foraging areas or for inhibition of movements resulting from impoundments are very limited. It would be advantageous to establish facilities and communities in areas where they are not in the middle of bear movement corridors. However, I doubt that efforts to situate these facilities in areas where they are distant from the river and, correspondingly, distant from black bear transportation corridors, can be justified on the basis of certainty that this effort would significantly benefit the black bear population remaining after the postimpoundment period. This is because such relocation would
likely be very costly and because the black bear population in the vicinity of the upper impoundment will probably be so greatly reduced by other impoundment-related impacts that few bears will be left to benefit. It is worth noting that most black bear movements up- and downstream occur on the north side of the river. Correspondingly, facilities situated on the south side are likely to have less impact than those on the north side.

## 8.D. Interspecific Effects

8.D.1. Moose and brown bears

As with brown bears, it is difficult to estimate the effects on black bears of project-caused changes in abundance of other species. Nevertheless, such impacts are likely to occur and their probable direction can be reasonably predicted.

The predicted reductions in numbers of brown bears, as a result of the project, could only be beneficial to remnant black bear populations. Brown bears are suspected of kiliing some black bears and attacks have been documented in this area (Miller 1985b). Also, I suspect that with reduced brown bear populations, black bears would probably forage somewhat further from forested escape habitats. If this happened, it would effectively expand the amount of habitat available for black bears. Conversely, black bears forced to move into more open habitats as a result of flooding of current habitats could be more exposed to predation from brown bears.

Reduction of brown bears may increase the number of moose calves available as prey to black bears. Black bears in the Susitna area currently kill fewer moose calves than black bears on the Kenai Peninsula (see Section 8.G.4 of this report). In part, at least, this may be because brown bears are much more abundant in the Susitna area than on the Kenai. This possible increase in spring food supply would result only if moose populations remained constant or increased. If moose populations declined as a result of the project (Ballard et al. 1985), then more calves would not necessarily be available to black bears regardless of reduced brown bear predation on moose calves.

## 8.D.2. Human/bear interactions

Compared with brown bears, black bears are tolerant of human presence (Herrero 1985). Correspondingly, I would expect much less human-caused disturbance displacement to occur for black bears than for brown bears. Because of this tolerance, however, black bears are likely to thoroughly explore the food-producing potential of the new human communities in the impoundment area. In this way bears will inevitably come into
conflict with man. Problems, including killing of nuisance bears, can best be minimized by very careful handling of garbage and other human foods and by strict enforcement of regulations against feeding wildlife. The recommendations of Bromley (1985) should be reviewed and followed during construction and operation of the project to minimize these conflicts. Especially in the vicinity of the Watana Impoundment, the amount of forested habitat that remains along the fringe of the impoundment shoreline will be greatly reduced by impoundment flooding. Black bears will be increasingly valnerable to hunting by humans in the remaining forested habitat.

## 8.E. Downstream Impacts on Black Bears

Negative impacts on black bears downstream from the proposed impoundments were anticipated during Phase I of this project (Miller and McAllister 1982). I thought these impacts would likely result primarily from reduced availability of salmon, especially spawning salmon, in sloughs and tributaries between Talkeetna and Devils Canyon and especially between Curry and Devils Canyon (Miller and McAllister 1982). Only rarely are salmon able to swim upstream through Devils Canyon so reduction of salmon is not a consideration in the upstream study area.

I anticipated reductions of salmon in the downstream area based on fisheries studies then occurring as part of Su-Hydro investigations. No final report on these studies of projectrelated impacts on salmon in the Susitna River is available. Correspondingly, without a documented level of reduction of salmon availability, I am unable to predict impacts on bears. Given this lack of information, it is fortunate in terms of prediction of impact on bears, that the data I collected on bear use of salmon in the downstream study area suggest salmon availability is not as important as hypothesized earlier.

Studies of bears downstream from Devils Canyon began in 1982. Additional bears were captured and marked in 1983. Radiotracking data on these bears revealed that most utilized the slough and riparian areas along the main Susitna River especially heavily during the July-August period when salmon were spawning in these areas (Miller 1983, 1984, and 1985a). Correspondingly, in 1982, 1983, and 1984 I visited this area, inspected the sloughs, and collected fresh bear scats. Most scats collected in mid-August were found along the Susitna River or sloughs along the Susitna in the zone between Curry and Portage Creeks. Nomenclature of sloughs follows Su-Hydro fisheries studies for the anadromous adult project. Analyses of scats were made by Paul Smith following procedures outlined by Smith (1984). Data on contents of the scats collected each
year are presented in Tables 47-49. In most cases it was impossible to differentiate between black bear and brown bear scats; efforts to develop differentiation techniques were unsuccessful (Appendix 4). Numbers of salmon counted in sloughs and tributaries by Su-Hydro fisheries staff in each year from 1981 through 1984 are presented in Table 50.

Fish were present in identifiable amounts in only 3 of 76 scats collected in the downstream study area. In 2 of these, fish were present in trace amounts and in one it was present in "category 2" amounts (6-25\% of scat contents). The low number of fish remains in these scats was puzzling to us as we saw many fish that had been killed and partially eaten by bears during our inspection of the downstream sloughs (Tables 51 and 52). Fame (1974) observed heavy use of salmon by black bears in Prince William Sound, Alaska. I doubt that the absence of salmon in the scats we analyzed resulted from lack of ability to recognize salmon in scats due to differential digestibility or other reasons. At McNeil River and along Prairie Creek I have seen many scats from bears that have been eating salmon and have noted that these are readily identifiable based on superficial inspection. These scats frequently contain bones, are diarrhetic, and have a distinctive unpleasant smell.

By far the most abundant item in the scats collected in the downstream area in August was berries of devil's club (Oplopanax horridus) which occurred in 75 of the 76 scats. Amount of scat represented by devils club was: trace (3\% of scats), 6-25\% (9\%), 26-50\%(25\%), 51-75\%(17\%), and 76-100\%(45\%). Devil's club was not an abundant plant in the downstream area. It occurred primarily in the zone between the scoured riparian flats and the adjacent forest. Farther upstream from Devils Canyon, in the upstream study area, this plant was rarely found and seldom seen with berries. Based on available data it appears that the July-August movements of black bears to riparian areas (movements documented with telemetry data) were more likely motivated by the presence of ripening devil's club berries than by spawning salmon. On the Kenai Peninsula, Schwartz et al. (1983a, 1983b) have documented late summer movements of black bears to hillsides of mature upland forests containing devil's club. In these summer feeding areas black bear scats indicated bears were feeding almost exclusively on devil's club berries (Schwartz et al. 1983a \& b). The relative absence of devil's club in the upstream study area may cause or contribute to this area's carrying capacity being much lower, in average years, than in the downstream area or in the Kenai Peninsula area studied by Schwartz.

Our data may not accurately represent the importance of salmon to bears in the downstream study area. It is possible that bear use of salmon in downstream sloughs was more prevalent in

July and early August than in late August when we collected most of our scats. In late August it is possible that bears switch from an earlier and greater dependence on salmon to ripening berries. It is also possible that salmon are an important buffer food source that is more heavily used in years of berry-crop failure. Finally, bears may use both salmon and berries in a daily cycle that makes it unlikely that salmon-rich feces would be found at the salmon-spawning areas. Based on available information, however, there is no reason to conclude that reduction from salmon availability in sloughs and tributaries downstream of the impoundment area would impact carrying capacity for black bear populations in this area.
8.F. Cumulative Impacts, Black Bears

For black bears, cumulative impacts of the proposed project may be greater than the sum of individual impacts. Methodology, to identify and quantify such cumulative impacts on brown bears has been described by Christensen (1985), Young (1985), Winn and Barber (1985), and Weaver et al. (1985).

No effort to conduct similar cumulative-effects analyses was made as part of this report, but such an effort should be undertaken as part of environmental impact assessments for the Susitna hydroelectric project. I suspect that such analyses would lead to the conclusion that the combination of habitat destruction through inundation, reduced berry-foraging areas because of construction sites and other facilities, reduced availability of good den sites, increased disturbance and hunting in the remaining habitat, increased destruction of "nuisance" bears, road kills on access routes, and other factors, will, in total, result in the complete elimination of the black bear population in the vicinity of the Watana Impoundment. As discussed elsewhere in this report, I think the upstream black bear population is only marginally secure at present and may be subject to periodic wide fluctuations in numbers, based on annual environmental differences. Superimposition of additional sources of stress on such a marginal population would likely result in complete loss of the ability of the habitat to support black bears.
8.G. Background Information on Black Bear Biology

## 8.G.1. Black Bear Productivity

As for brown bear (Section 7.G.1), I suspect that the impoundment will result in declines in availability of foods currently utilized by black bears and that these declines will be reflected in changes in bear numbers as well as in declines in productivity. Changes in productivity are difficult to
predict, so my effort has concentrated, primarily, on documenting existing levels of productivity so that changes can be measured during post-impoundment studies. Currently, the upstream population is less productive than a Kenai Peninsula population of black bears intensively studied by Schwartz et al. (1983b). The major difference in these 2 areas is that cub mortality is much higher in the upper Susitna. I suspect that the major difference in food supply between the Kenai and upper Susitna populations is that devils club berries, important on the Kenai and lower Susitna River in late summer, are essentially not available to black bears in the impoundment area. I also suspect that black bears in the upper Susitna are highly dependent on blueberry crops and have fewer buffer foods to turn to when blueberry crops fail (Section 8.G.4.a, this report).

Reproductive data discussed in this section are derived largely from observations of radio-marked bears. This source of data is subject to sighting errors. Such errors were especially likely in the downstream study area where heavy vegetation frequently prevented visual observation of the bear at the time it was radio-located. Reproductive status could not be confirmed unless the bear was seen. Especially in the early spring, newborn black bear cubs frequently hide in trees when approached by radio-tracking aircraft. This made sighting and counting of cubs very difficult. These problems are much more likely with the black bear data than with the brown bear data discussed earlier because brown bears were more frequently in open country where they, and their offispring, could be easily seen.

## 8.G.1.a. Litter Size and Offspring Mortality

Mean litter size at the time radio-marked females were first observed for 42 litters of newborn cubs was 2.1 (range $=1-4$ ) (Table 53) and for 28 litters of yearling offspring it was 1.9 (range $=1-3$ ) (Table 54). At time of first observation 74 告 of litters had 2 cubs; 17\%--3 cubs; 7\%--1 cub; and 2\%--4 cubs (Table 53). Litter sizes were approximately equivalent on the Kenai (1.9 for 15 litters of newborns, Schwartz et al. 1983). Sex ratios of newborn cubs handled ( $N=44$ ) was 76 males:100 females, and for 10 yearlings the ratio was 100:100 (Tables 55 and 56).

In Su-Hydro studies, I defined as "mortalities" cases in which a female was observed with newborn offspring (either in her den or following emergence) but did not have the same number of offspring at the time of entrance into her next den. For 60 newborn cubs in both the upstream and downstream study areas, $35 \%$ experienced such mortalities (Table 57). This percentage was much higher in the upstream study area (47\%
mortalities for 43 cubs) than in the downstream study area (6\% mortalities for 17 cubs) (Table 57). In Kenai Peninsula studies, no mortalities were observed for 13 newborn cubs between ages 0.3 (emergence) and 1.7 years (separation from mother), but a third of 9 radio-marked yearlings died (Schwartz et al. 1983b). We had only 2 radio-marked yearlings and one of these died during its yearling summer; the other (329) survived into adulthood.

Schwartz et al. (1983a \& b) provided weights for 16 yearlings captured in dens or shortly after emergence in the period February-June 1983. These bears ranged in weight from 29 to 126 lbs (mean $=83 \mathrm{lbs} ., \mathrm{S} . \mathrm{D} .=30 \mathrm{lbs}$ ). During the course of my studies in the upstream black bear study area, I weighed 7 yearlings and estimated weights during handing for 3 more during April through June of different years. These 10 bears weighed an average of 24 lbs (range $=14-33$ lbs., S.D. $=7$ lbs.) (Table 56). Although these data sets are of different sizes and represent somewhat different periods they suggest that Kenai Peninsula black bears are in much better condition following their first summer than are upper Susitna bears. The high mortality of newborn black bear cubs in the upper Susitna and the relatively slow growth rate of these cubs in their first year of life most likely reflects relatively poorer habitat and foraging conditions for black bears in the upper Susitna compared with the Kenai Peninsula. Two of the lightest Kenai yearlings (20 and 22 pounds--Schwartz et al. 1982) died of malnutrition as yearlings (Schwartz et al. 1983).

There are other factors which may contribute to high cub mortality in the upstream Susitna area. Some black bear mortality in the Su-Hydro area is probably caused by brown bear predation. Brown bears are much less common in the Kenai Peninsula area studied by Schwartz. It is also possible that the Kenai Peninsula area as well as the downstream Susitna study area have lower cub mortalities than the upstream Susitna area because the proportion of adult male bears is lower as a result of relatively high hunter effort. Bunnell and Tait (1980) noted that hunting typically results in skewed sex ratios and Young and Ruff (1982) observed apparent increases in cub survivorship following experimental reduction of adult males in an Alberta black bear population. Tietje et al. (1986) noted an instance of interspecific predation on young black bears.

Measurements of newborn cubs are presented in Table 55.

## 8.G.1.b. Reproductive Interval

Methods of measuring reproductive interval were discussed in Section 7.G.1 of this report. Following Reynolds and Hechtel (1985) I defined reproductive interval as the period between
successful breeding (as evidenced by cub production the following year) or successful weaning of a previous litter and the next successful separation of mother and offspring ("weaning"). Intervals based on females initially captured with yearlings were not counted by back-dating this litter. I considered it to be a successful separation if the adult female was seen with those yearling offspring following emergence from the den shared with her yearling offspring. With this definition it is usually not possible to distinguish between mortality experienced by yearlings while accompanied by their mothers and "successful separation". Since in most cases separation occurs relatively early, in May or June, this source of error is probably small. Separation from yearling offispring occurred in 23 cases (289 [3 cases], 290, 301 [2], $317[2], 321,327,349,354,363,364,369,375,376,378,402$ [2], 411 [2], and 432) and from 2-year-old offspring in 2 cases (verified in den for female 361 and based on sightings for female 405) (Table 58).

In some instances a female would separate from her yearling offspring in the spring, during breeding season, but they would apparently reunite later in the summer (sometimes just before den entrance). At least in cases where the female was pregnant it appeared that the yearling and its mother would not den together following such a reunion (e.g. 289 in 1984, and 317 in 1985). In some cases, the female was apparently not pregnant (had no newborn upon exit) but was seen with a smaller bear (probably her 2-year-old offspring) at exit from the den the following year (e.g., 317 in 1981, 364 in 1984, and 376 in 1984). In these cases I am uncertain whether the bears denned together or whether they denned near each other. Denning together by unrelated bears has been recorded but is rare (Schwartz et al., in press).

Reproductive histories of individual females are presented in Table 58. Reproductive intervals based on these histories are summarized in Table 59. Counting only reproductive intervals for which complete data were available ( $N=25$ ), I found that intervals ranged from 2 to 5 years and averaged 2.4 years for bears in upstream and downstream areas combined (Table 59). As previously mentioned for brown bears, using only complete intervals underestimates the true reproductive interval. This is because many intervals are incomplete and, in a short study period, the incomplete intervals tend to be those that are longer than minimum length. If one assumes no more skipped years or lost litters for the bears with currently incomplete intervals ( $\mathrm{N}=15$ ), the calculated mean interval for these bears averages 3.1 years (Table 59). When completed, some of these intervals will be longer than the minimum value. For example, 9-year-old female 441 was alone when captured in 1985; she apparently bred in that year but did not have cubs
in 1986 (Table 58). If she has cubs in 1987 and weans these in 1988, she will have had an interval of 3 years and this is the value included for her "incomplete interval" (Table 59). Combining available complete intervals and minimum values for incomplete intervals ( $\mathrm{N}=40$ ) provides an average reproductive interval estimate of 2.7 years (range $1-5$ years) (Table 59). Intervals appear equivalent in the downstream study area (2.6 years, $N=12$ ) and upstream (2.7 years, $N=28$ ) study areas (Table 59). Counting incomplete intervals, 2-year intervals were most common (53\%), followed by 3-year intervals (33\%), 4-year intervals (10\%), and 5-year intervals (5\%) (Table 59).

Schwartz et al. (1983b) reported 1 interval of 2 years and 5 intervals of 3 years on the Kenai Peninsula. This yields an average interval of 2.8 years for his data. Schwartz did not report incomplete intervals which would probably have raised this average value. Based on available information I cannot conclude that reproductive intervals were different in the Kenai and Susitna studies.

## 8.G.1.c. Age at First Reproduction

In this study $I$ defined "age at first reproduction" as the age when the first observed litter was produced. This definition will overestimate actual age at production of first litter when whole litters are lost before they are observed. Other errors may be introduced through errors in aging based on cementum lines.

Limited data are available for age at first reproduction because few transmitters were placed on subadult bears. Black bear 329, tagged as a yearling in 1981, still had not produced a verified litter through 1986 when she was 6 years old (Table 58). She was seen with males during breeding seasons when she was 3, 4 , and 5 years old (Table 58). The earliest this bear could produce a litter is age 7 (in 1987). For all other, bears, age at first reproduction is based on cementum age. Bear 448 had no observed litters when it was either 6 or 7 years old (Table 58). If we assume no litter was produced before she was captured at age 6, the earliest this bear could produce a litter is at age 8 (in 1987). In the following calculations bears 329 and 448 are assumed to produce first litters in 1987 when they will be 7 and 8 years old respectively. Summary data used in calculating age at first reproduction are presented in Table 60. For 14 black bears for which reasonable data are available (Table 60), mean age at first reproduction was 6.4 years. Half of these bears produced first litters at age 7 (Table 60).

On the Kenai Peninsula Schwartz et al. (1983b) found 6 females that produced first litters at age 4 while 2 others had not produced litters yet by ages 4 and 5. If we assume that these last 2 females produced cubs the following year, the mean age at first reproduction was 4.4 years (range $=4-6$ ). Based on these data, Kenai Peninsula black bears reach reproductive maturity at a younger mean age than bears in my study area ( $t=25.9$, 20 d.f., $P<0.001$ ). This result could be predicted from the slower growth rate of Su-Hydro bears as indicated by lighter weights of yearlings in the Su-Hydro area, discussed above.

## 8.G.2. Sources of black bear mortality

As for brown bears, hunter kills of black bears in the Su-Hydro study area have generally increased during the period 1973-85. Reported kills averaged 13 bears/year during this period (Table 61). This is lower than the hunter kill of brown bears which averaged 19/year in the same area during the same period (Table 24). In the last 5 years (1981-1985) hunters have killed an average of 14.6 black bears and 27.6 brown bears (Tables 24 and 61). I suspect that at least some of the increase in bear harvest in this area, especially for black bears, resulted from augmented interest in and knowledge of the area on the part of staff working on various projects associated with the proposed Susitna hydroelectric dams. This suspicion is based on personal knowledge of hunting by such staff. Increases in harvest are expected when formerly remote areas are opened up by improved access or publicity of available game. Additional increases can be expected if roads to the dam sites are built. Under these circumstances regulations may need to be adopted to prevent harvests of bears and other wildife from exceeding acceptable levels. Because black bears inhabit the forested fringe along the shores to the proposed impoundment, remnant black bear populations in the impoundment area would be especially vulnerable, in the very narrow post-impoundment fringe of forested habitat, to hunters using boats on the reservoirs.

The proportion of the marked black bear population that is taken by hunters is an index to the population exploitation rate. These data are provided in Table 62. If both upstream and downstream black bears are included, annual kill rates of marked black bears ranged from 6\% to l7\% (Table 62). Exploitation rates were higher in the downstream study area than upstream from Devils Canyon (Table 63). This is probably because downstream from Devils Canyon, bears can be hunted easily from a river boat while upstream from Devils Canyon access is primarily by float plane. Natural mortality of radio-marked black bears during the study period was high compared with that of brown bears (Table 29). A total of 13 black bears died, mostly from unknown causes (Table 29). I
suspect a couple of these deaths may have resulted from gunshot wounds. Available indications suggested that others resulted from natural causes including predation by brown bears (Table 29). The apparent high natural mortality of adult bears in the upstream study area is another indication suggesting that this area may be marginal habitat for black bears.
8.G.3. Black bear movements
8.G.3.a. Home range size

As for brown bears, black bear home ranges were calculated using minimum home range polygons (Mohr 1947). In many cases these home ranges were not accurate representations of the areas utilized by individuals. This was because black bears were largely restricted to movements up and down the river, but since the river does not run in a straight line, the minimum home range polygons include areas not utilized by bears between river meanders. This point is illustrated in Figures 29-33 for annual home ranges of 5 black bears. Home ranges for individual bears in specific years, and for all years combined, are presented in Table 64. Annual home ranges for all bears averaged $134.6 \mathrm{~km}^{2}$; male home ranges ( $251.5 \mathrm{~km}^{2}$ ) were larger than female home ranges (67.1) ( $t=13.1$, 121 d.f. $P<0.001$ ). Home ranges of females in years they had newborn cubs ( $69.2 \mathrm{~km}^{2}$ ) were not significantly different from those of females in years they did not have cubs $\left(77.3 \mathrm{~km}^{2}\right)(t=0.05$, 64 d.f., $P>0.5$ ) (Table 65).

Average male home range size varied little in different years of the study except for the first year (Fig. 34). The first year had a lower average because some bears were not captured until August. Home range for females without newborn cubs was larger in 1981 than in other years (Fig. 34). In 1981 there was an apparent failure of the berry crop which probably accounted for the larger home ranges in that year.

## 8.G.3.b. Seasonal movements

The basic seasonal pattern for black bear movements in the study area is for black bears to remain in the forested riparian zone along the river for denning and during spring and early summer. When berries are ripening in late summer and fall, black bear movements become more extensive in both upstream and downstream directions. At this time black bears may also venture out of the forested zone into the adjacent shrub zone.

Variations in this pattern were observed in 1981 when, in response to an apparent berry crop failure, bears moved much more extensively in both upstream and downstream directions
(Figs. 29-33). Most bears did not make equivalent movements in other years but male 343 (Fig. 32) continued to make similar movements downstream each year in late summer. These movements were probably motivated by increased availability of devil's club berries downstream or, possibly, the availability of salmon in downstream sloughs.

Another variation in this pattern was observed in spring 1985, when black bears appeared to be more abundant at higher elevations away from the Susitna River. I suspect this difference was related to availability of overwintered berries. Overwintered berries, especially crowberry (Empetrum nigrum) are an important spring food for bears. Winter 1984-85 had little snow cover at lower elevations along the river until February. I suspect that lack of snow cover reduced overwinter survival of berries at lower elevations, forcing some bears to forage at higher elevations distant from the riparian forest. These areas are thought to be less preferred by black bears as they may be more vulnerable there to attack by brown bears.

## 8.G.3.c. Dispersal from study area

Only 1 dispersal into or out of the study area was documented for subadult black bears. Little effort was made to obtain such documentation by placing radio-transmitters on subadult black bears. Only 1 yearling was radio-marked and survived for more than 5 months; this bear (female 329) did not disperse. Another male marked as a 2-year-old in the upstream study area in 1980 (323) did not disperse and was shot by a hunter in September, 1983. A male marked in the upstream study area (Clark Creek) in May 1980 did disperse. This bear, 307, was shot by a hunter 1 year later near Hurricane on the Parks Highway.
8.G.4. Black bear food habits
8.G.4.a. Predation rates

Black bears are known to be effective predators on moose calves (Franzmann et al. 1980) but, in 1 case at least, black bears were observed to be inhibited, compared with brown bears, in killing moose calves (Miller 1985b). In this case a black bear watched a cow moose with 2 newborn calves for over 24 hours without successfully attacking, but a brown bear attacked and killed the calves as soon as it found them (Miller 1985b). Simultaneous with intensive monitoring of brown bears (Section 7.G.4.b this report), radio-marked black bears were intensively monitored in 1981 and 1984 to estimate predation rates (Table 66). During periods of intensive monitoring in the spring, 16 black bears were observed on 13
calf moose kills, 1 adult caribou kill, and 1 probable kill during a total of 460 visual sightings. This translates to 2.8 moose calf kills/100 visual sightings, 4.1 kills of all kinds/100 observation-days, and 5.4 kills (all kinds)/100 consecutive observation-days (Table 66). An "observation-day" was defined as a day on which a bear was seen at least once and a "consecutive observation-day" summed all periods of $>2$ consecutive observation-days.

This kill rate is about $25 \%$ of that observed for brown bears (Section 7.G.4, this report). Brown bears were observed during intensive monitoring at the same time on 16.5 kills/ 100 consecutive observation-days (Table 32), compared with 4.1 for black bears. If one considers just moose calves, brown bears were observed on 9.9 kills/100 consecutive observation-days and black bears on 1.9 (Tables 66 and 32).

A kill rate of 2 calves/100 consecutive observation-days during a 5 -week period when moose calves are most vulnerable would result in an average estimated kill of 0.7 calves/bear/ year. In section 8.A of this report I estimated black bear populations in the impoundment impact area to be 107 bears. If one assumed $35 \%$ of this population was cub and yearling bears (Miller et al., in press; Appendix 2), about 70 bears were available to prey on moose calves. At 0.7 calf kills/bear, these bears would kill about 50 calves/year in the Su-Hydro study area.

These kill rates are minimum estimates because it is easy to miss kills during radio-location flights. Regardless, it. appears probable that at this low kill rate predation on moose calves by adult black bears is unlikely to contribute significantly to the spring nutrition needs of these black bears. It may be a more significant source of nutrition for some individuals that are particularly adept at killing calves. For example, of the 13 calves observed killed, 7 were killed by 2 of the 16 intensively monitored bears.

## 8.G.4.b. Annual variation in berry abundance

As discussed in Miller and McAllister (1982), a berry-crop failure apparently occurred in summer 1981. I first suspected a berry crop failure because movements of black bears in late summer of that year appeared much more extensive than in 1980; radio-locations in subsequent years verified that movements in 1981 were extensive. In late summer 1981, black bears made atypical movements in both upstream and downstream directions. These movements were discussed for each individual in Miller and McAllister (1982:103) and are illustrated, for 4 bears, in Figs. 29-33). Observations on the ground in late summer 1981 provided subjective verification that berry crops were
exceptionally low in 1981 compared with other years of this study (Table 67). Years during which these data were collected were subjectively appraised as "near typical" for the upstream study area. This is different from the preceding year, 1981, when berry crops in black bear habitat were thought to have had a widespread failure (Table 67).

## 8.G.4.c. Scat analyses

Food-habits data based on scat analyses were of limited value because few scats were collected in upstream areas, and because of the difficulties in differentiating between black and brown bear scats (Appendix 4). Most scats were collected along sloughs and streams in the downstream study area in an effort to evaluate the importance of salmon to bears in this area (Section 8.E, this report). Scat data are presented in Tables 47-49.

## 8.G.5. Black bear denning ecology

My data on the denning ecology of black bears have been analyzed and contrasted with data from 2 other parts of south central Alaska by Schwartz et al. (in press, see Appendix 1). Only those components of the black bear denning data that are directly related to the proposed hydroelectric project will be discussed in this report.

Den entrance and emergence dates for each individual black bear in each year are given in Tables 68-72. No differences were observed between males and females but pregnant females entered dens earlier than males or non-pregnant females (Schwartz et al., in press).

Locations of black bear dens in upstream and downstream study areas are illustrated in Figs. 35-36. Characteristics of these dens are presented in Table 73 and the tendency to prefer southern aspects is illustrated in Fig. 37. History of den use by individual bears is presented in Table 74 and by individual dens in Table 75. These data demonstrate a high rate of reuse of individual dens by bears in the upstream Su-Hydro area compared with other study areas (Schwartz et al., in press) and suggest that good den sites may be limited in the upstream study area.

Forty-four different dens were found in the vicinity of the Watana Impoundment; 55\% of these were dug, $41 \%$ were in natural cavities, and $2 \%$ were of unknown cavity type (Table 75). Of these dens, $55 \%$ would be flooded by the proposed impoundment and $46 \%$ would not be flooded (Table 75).

Thirty different dens were found in the vicinity of the Devils Canyon Impoundment; $33 \%$ of these were dug, $43 \%$ were in natural cavities, and $7 \%$ were of unknown cavity type (Table 75). Of these dens only 1 (3.3\%) would be flooded by the proposed impoundment (Table 75).

In the downstream study area 29 black bear dens were found. Compared with the upstream area, fewer downstream dens were in natural rock cavities and more were dug (Table 75).

These data suggest that the Watana Impoundment would probably result in a reduction of acceptable denning sites for black bears resident in this area. This factor might become limiting for black bear populations in this area if populations remained at pre-impoundment levels. Since black bears in the Watana Impoundment area are expected to decline greatly in number based on reductions in habitat and carrying capacity, it is likely that the population will actually be limited by habitat shortage before the bears are limited by a shortage of den sites. The Devils Canyon dam is likely to have little impact through inundation on black bear denning habitat.

Black bears den in the forested habitats along the Susitna River in the vicinity of both the upper and lower impoundments. Pre-inundation clearing of forests in and adjacent to the proposed impoundment during the denning period would probably result in disturbance of many black bears and additional mortalities, to some individuals, resulting from den abandonment. If logging occurs during the denning period, as anticipated, black bears should be radio-marked and monitored prior to the clearing in order to document the impact of this source of disturbance.

## 9. BEAR DENSITY AND POPULATION ESTIMATION

Standardized methods for estimating bear numbers have not been developed. Even in very intensively studied populations where all bears are marked or radio-collared, it can be difficult to convert these data to meaningful density estimates (Schwartz et al. 1983a).

In this study I attempted to estimate black bear density using Lincoln-Petersen Indices where radio-marked bears constituted the marked sample. In summer 1982, when black bears were in relatively open habitats feeding on berries, and in spring 1983, before leaf emergence restricted visability, I attempted to estimate bear numbers using ratios of marked to unmarked bears observed in a single flight. In these efforts the number of marked bears present in the search area was determined through radio-tracking flights before and after the
observation flight. Estimates with very large variance were achieved with this procedure, probably because observability was so low (see Miller 1984 for these results).

Work conducted in spring 1985 was designed to provide an improved density estimate for both black and brown bears in the Su-Hydro study area. This work was essentially a series of replications, in a well-defined smaller area, of the technique used in the 1982 and 1983 studies. Consecutive days of search effort were combined to provide a series of independent estimates over time and a single combined estimate of the number of bears present in the search area during an average day of the search period. This technique has been published (Miller et al., in press, see Appendix 2) and only those site-specific details not included in this publication will be repeated here.

The search area and quadrats used to allocate search effort are illustrated in Fig. 38; time spent actually searching in each quadrat is presented in Table 76 (commuting time and time spent circling bears prior to capture is excluded). We were forced to base this census effort from Talkeetna which greatly increased commuting time to the search area. Total fixed-wing charter time was 264 hours, twice the number of hours spent in actual search (Table 76). Because this was a newly developed technique some errors were made which should be avoided in future applications. The most serious of these errors was failure to search each quadrat on each day of the search effort (Table 76). This was not considered a problem at the time because $I$ originally intended to combine a number of days' data to obtain an estimate for that period. If this had been done the missed quadrats on a single day would not have been such a serious problem if all quadrats were searched equally over the period.

The problem with combining days, however, is that one could potentially have more marked bears seen during a period than were "present" during that period (where presence for each bear is a fraction equaling the proportion of time the marked animal spent in the search area). In illustration, a marked bear that was present half of the time in the period would be counted as 0.5 marked bears present, but if seen one or more times it would be counted as 1.0 marked bears seen.

This problem was eliminated through use of the bear-days estimator described by Miller et al. (in press, Appendix 2). This estimator provided a brown bear density estimate of 2.79 bears/100 $\mathrm{km}^{2}\left(95 \% \mathrm{CI}=2.52-3.30\right.$ bears $/ 100 \mathrm{~km}^{2}$ ) and a black bear density estimate of 8.97 bears/100 $\mathrm{km}^{2}(95 \% \mathrm{CI}=$ 7.74-10.21 bears/100 $\mathrm{km}^{2}$ ).

These density estimates were extrapolated to the area identified as that in which bears would be affected by the proposed hydroelectric project. This extrapolation provided an estimate of the number of bears that would be impacted by the proposed project. Evidence based on relocations of radio-marked brown bears during 1980 through 1984 illustrate that all of the search area was brown bear habitat (Fig. 40). The density estimate for brown bears represented density in habitats below 5,000 feet elevation; the amount of area below 5,000 feet elevation in the brown bear impact area was 11,704 km2 (12,127 minus 423 km 2 above 5,000 feet elevation). For just Devils Canyon the impact area was $6,833 \mathrm{~km} 2 \quad$ (7,120 minus 287 above 5,000 feet) while for just the Watana Impoundment the area was $9,056 \mathrm{~km} 2$ ( 9,452 minus 398 above 5,000 feet). At the density estimated above, the estimated number of bears in the impoundment study area was 327 (95\% CI = 295-386) .

The density estimate for black bears was extrapolated to the area (1195 km2) identified as black bear habitat based on radio-locations of marked bears and habitat considerations (Figure 7), resulting in an estimate of 107 black bears in the impoundment impact area (95\% CI $=$ 93-122). Because of overlaps of the impoundments' impact zones, over half of this value would be within the impact zone of either impoundment considered separately.

The 1985 estimated population of 107 black bears may be less than maximum carrying capacity of this habitat following a series of good years for food crops. I suspect the poor berry crop in 1981 resulted in a reduced black bear population in this area, although there is little objective data available to support this conclusion. I based my suspicion on less frequent sightings of black bears, in 1982 and subsequently, than in 1980 and 1981.

## 10. BERRY ABUNDANCE AND CANOPY COVERAGE

Personnel conducting Su-Hydro studies designed to measure moose forage biomass in the impoundment area (Becker and Steigers 1986) simultaneously collected information on plants producing berries eaten by bears, as well as on horsetail (Equisetum spp.). The bear data were collected during 1 $\overline{1}$ July-25 August 1986. Information was collected on transects including randomly spaced plots of 1 square meter. Transects were also identified as within willow (Salix spp.) biomass strata and plots were identified as being within vegetation types based on both vegetation mapping and on-ground classifications at the time data were collected. Transects were run from the Susitna River up to elevations of 3400 feet. Details of sampling schemes and mathematical treatments of these data are presented by Becker and Steigers (1986). Data on canopy coverage of berry-producing plants (as
well as Equisetum), on berry abundance, and on berry ripeness were collected for blueberry (Vaccinium uliginosum), crowberry (Empetrum nigrum), and lowbush cranberry (also called lingonberry) ( Vaccinium vitis-idaea). Six canopy-coverage categories were used: Absent, trace-5\%, 6-25\%, 26-50\%, 51-75\%, and 76-100\%. Four berry-abundance categories were used: None, trace, 5-20 berries, and $\mathbb{I} 20$ berries. Five ripeness classifications were also used to represent average ripeness in each plot: green, starting, tart, sweet, and past. The first 2 and last 2 categories were lumped in my analysis of berry-ripening phenology. This analysis did not take elevation, slope, or habitat types into consideration (these factors may influence ripening phenology). For analysis of ripeness, data were lumped into 6 intervals of approximately 1 week each.

Data were weighted by willow biomass strata to reflect differing sampling intensities in these strata, and were analyzed to produce statistics on mean canopy coverage and berry abundance in each of 3 "populations" (within the flooded zone for each impoundment and outside of this zone up to an elevation of 3400 feet). This design was not optimal for collection of data on bear foods because this objective was incidental to the main purpose of the browse survey. I gratefully acknowledge the assistance of Earl Becker (ADF\&G) and Bill Steigers (LGL) and their crew in collecting these data; Earl Becker also assisted in the analysis of these data.

Phenology
In 1985, phenology of berry ripening was similar for blueberry and crowberry; the incidence of green berries dropped rapidly during the first week of August and the incidence of sweet berries increasing rapidly during the third week of August (Figs. 40a \&40b). For lowbush cranberry, this ripening pattern was about 2 weeks delayed and few plots with ripe berries were found during the 3rd week of August when the study ended (Fig. 40c). Since most black bears in this area enter dens during the last week of September and first week in October (Section 8.G.5, this report), these data illustrate that ripe berries are available to this population of black bears for a period of only 4-6 weeks.

Abundance and Canopy Coverage
The estimated proportion of berries and berry bushes and the standard error for this estimate (corrected for covariance effects) was calculated according to the methods described by Becker and Steigers (1984). These data are presented and illustrated in Figures 41-47. The estimated proportion was converted to a whole number by multiplying by the number of
transects in each population (47 in the Devils Canyon vicinity below 2200 feet elevation, 165 in the Watana vicinity below 2200 feet, and 126 above 2200 feet). Following this multiplication, categories with $<5$ "observations" were lumped with the next lower category and Chi-square tests run. Results of these Chi-square tests are given in Figures 41-47.

For blueberry abundance and canopy coverage, the null hypothesis that the 3 populations were equivalent could not be rejected (Figures 41 and 45).

The null hypothesis for crowberry canopy coverage (Fig. 42). By inspection of Fig. 42 (lumping last 3 categories) it can be seen that the area outside of the impoundment had fewer crowberry bushes. These data are consistent with a hypothesis that the impoundment area may be especially important for spring foraging by bears for overwintered crowberries. Sample size was inadequate to say much about crowberry abundance, but berries appeared more abundant in Population A (Watana Impoundment) than in B (above 2200 feet elevation) and more abundant in $B$ than in $D$ (Devils Canyon Zone).

Lowbush cranberry bushes were unequally distributed in the 3 populations, with more cover in populations $B$ and $D$ (Devils Canyon and outside impoundments, respectively) than in $A$ (Watana Impoundment) (Fig. 43). With reference to berry abundance, Population $B$ is the most productive with $A$ and $D$ having equivalent productivity.

For Equisetum canopy coverage the categories with $>5 \%$ coverage had to be lumped and the null hypothesis of equivalent distribution of Equisetum in the 3 populations was rejected (Fig. 44). This resulted from greater frequency of categories with $>5 \%$ in the zone outside of the impoundments than within the impoundment zone (Fig. 44).

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Appendix 1.
A COMPARISON OF DENNING ECOLOGY OF THREE BLACK BEAR POPULATIONS IN ALASKA

Charles C. Schwartz, Alaska Department of Fish and Game, Box 3150, Soldotna, AK, 99669.
Sterling D. Miller, Alaska Department of Fish and Game, 333 Raspberry Road, Anchorage, 99503.
Albert W. Franzmann, Alaska Department of Fish and Game, Box 3150, Soldotna, AK, 99669.

Abstract: Between 1978-1985, denning ecology of the black bear (Ursus americanus) was studied in the Kenai Peninsula, the Susitna River basin, and Prince William Sound, Alaska. All these populations are near the northern extension of their range. In different years the mean number of days spent in dens varied from 189 to 233 days; the maximum time spent in a den by an individual bear was 247 days. Timing of emergence in the spring and entrance in the fall appeared most related to time of year, and secondly, to weather, snow accumulation and melt, and food availability. Bears in the more severe climate along the Susitna River entered dens almost 2 weeks earlier and emerged later than bears on the warmer Kenai Perinsula. Chronology of denning differed between pregnant females and other sex and age groups, but overlap occurred with all age and sex groups. Site selection, vegetation type, and den type (cave, tree, excavated) varied between areas and was related to winter weather conditions (rain vs. snow), soil type (deep vs. shallow and rocky), and topography of the areas (mountains vs. flats). Den morphometry was compared between areas. Denning chronology was compared with that of other black bear populations in North America and with current theory on why bears den.

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Appendix 2.
BLACK AND BROWN BEAR DENSITY ESTIMATES USING MODIFIED CAPTURE-RECAPTURE TECHNIQUES IN ALASKA

Sterling D. Miller, Alaska Department of Fish and Game, 333 Raspberry Rd., Anchorage, AK. 99518-1599.
Earl F. Becker, Alaska Department of Fish and Game, 333 Raspberry Rd., Anchorage, AK. 99518-1599.
Warren B. Ballard, Alaska Department of Fish and Game, P.O. Box 1148 Nome, AK. 99762-1148.

Abstract: Population density estimates were obtained for sympatric black bear (Ursus americanus) and brown bear (U. arctos) populations inhabiting a search area of $1,325 \mathrm{~km}^{2}$ in south-central Alaska. Standard. capture-recapture population estimation techniques were modified to correct for lack of geographic closure based on daily locations of radio-marked animals over a 7-day period. Calculated density estimates were based on available habitat in the search area $\left(1,317 \mathrm{~km}^{2}\right.$ for brown bears and $531 \mathrm{~km}^{2}$ for black bears). Calculated density was 2.79 brown bears $/ 100 \mathrm{~km}^{2}\left(2.52-3.30\right.$ bears $/ 100 \mathrm{~km}^{2}$ ) and 8.97 black bears $/ 100 \mathrm{~km}^{2}$ (7.74-10.21 bears/100 $\mathrm{km}^{2}$ ). Calculated 95\% confidence intervals were $+13.7 \%$ of the estimate for black bears and $9.9 \%$ to $+18.5 \%$ of the estimate for brown bears. Probabilities of capture based on calculated sightability indices were not equal in some instances, so confidence intervals should be interpreted cautiously. Increasing the number of marked bears during the study period resulted in altered brown bear estimates and smaller confidence intervals, but because closure was a relatively good assumption for black bears in our study area, had little effect on black bear estimates or confidence intervals. When telemetry data were used to correct input values for lack of geographic closure, the Schnabel estimator and the mean of 7 separate daily estimates both yielded estimates close to our results. We recommend our technique for additional testing as a method to objectively compare bear densities between different areas or between different times. These procedures may also be appropriate for use with other species.

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Appendix 3.
CHARACTERISTICS OF NONSPORT BROWN BEAR DEATHS IN ALASKA
Sterling D. Miller, Alaska Department of Fish and Game, 333 Raspberry Rd., Anchorage, AK 99518-1599.
Mark A. Chihuly, Alaska Department of Fish and Game, 333
Raspberry Rd., Anchorage, AK 99518-1599.
Abstract: The sex, age, and other characteristics of 668 brown bears (Ursus arctos) killed in nonsport circumstances in Alaska during the period 1970-85 were examined. These data represent an unknown fraction of total nonsport kills as not all kills were reported. Both sport harvests and nonsport kills are increasing in Alaska. Nonsport harvests averaged 5.1\% of total sport and nonsport. kills. Areas with the highest human density had the highest ratio of nonsport to sport harvests. Nonsport harvests are most common during periods when most people are in remote areas to hunt or fish. Males predominate in the nonsport kills of younger bears and females in the nonsport kills of older bears. Regulations and other factors make adult male bears more vulnerable to sport hunters than adult female bears. Partially as a result, nonsport kills contain more adult females than sport kills. An analysis based on affidavits from 224 persons killing bears revealed that bears were shot to avoid perceived danger (72\%), to protect property (21\%), and to eliminate nuisances (7\%).

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Appendix 4.
Abstract of "Differentiation of Brown and Black Bear Scats: An Evaluation of Bile Acid Detection by Thin Layer Chromatography" by Enid Goodwin, ADF\&G (full text of report in Appendix 1 of Miller 1984).

SUMMARY: A thin-layer chromatographic technique (TLC) for separation and detection of fecal bile acids was evaluated for use in differentiation of black bear scats from brown bear scats. Fecal samples from 22 known black bears and 19 known brown bears were tested. Bile samples from 4 black bears and 3 brown bears were also examined using TLC. Statistical analysis of $R f$ values obtained from the fecal samples indicated no significant difference between brown bear and black bear chromatograms. The numbers of bile samples were too small for statistical analysis, but indications of possible differences were noted. Variations among individuals within a species were documented, as were significant variations within individuals. Variations were hypothesized to be primarily caused by dietary influences on bile acid production mechanisms. Pigment removal methods were also evaluated. Alkaline distilled water was found to be effective in removing berry pigments, while hexane was a preferred solvent for removal of other types of plant pigments.

# Susitna Hydroelectric Project Big Game Study 

Data Component Descriptions and Coding Schemes Black and Brown Bears

Alpha codes are left-justified, numeric codes are right-justified.

1. Species: $1=$ moose

2 = sheep
$3=$ caribou
4 = brown bear
$5=$ wolf
$6=$ black bear
$7=$ goat
$8=$ coyote
9 = wolverine
2. Project: A one-digit code project ID:
$1=$ upstream
2 = downstream
$3=$ GASAWAY
4.= Denali Hwy.

5 = Noatak
3-8. Individual ID: An integer number of up to six digits which will be unique for the individual animal it represents within the project. For Su-Hydro bears it is the tattoo number. If a bear is unmarked, ID=99.

9-12. Age (in years, no decimal).
13. Age code $=A$ (decimal age).
14. Sex code: $M=M a l e, F=$ Female, blank $=$ unknown.

15-17. Observation number: An integer number up to three digits which uniquely identifies the sighting of an individual animal. The value must be right-justified.

18-23. Date: Two-digit integer for each: month, day, and year, respectively, each right-justified.

24-27. Time: Military time (by 24-hour clock), right-justified.
28. Visual: Was the individual actually sighted, or located only by radio?
( ${ }^{3} 1 / 16^{\prime \prime}$ on $1: 63,360$ )
$H=$ able to map with a high $\quad B=$ located only within degree of accuracy a broad range
( ${ }^{3} 1 / 8^{1 \prime}$ on $1: 63,360$ )
$\begin{aligned} M= & \text { able to map with a moderate } \quad I= \\ & \text { degree of accuracy }\end{aligned}$
degree of accuracy intermediate range
( ${ }^{2} 1 / 8^{1 \prime}$ on $1: 63,360$ )
$\begin{aligned} L= & \text { able to map only to a low } \\ & \text { degree of accuracy }\end{aligned} \quad C=\begin{aligned} & \text { located within a } \\ & \\ & \text { close range }\end{aligned}$
$\mathrm{Y}=$ yes; level of mapping $\quad \mathrm{N}=$ no; not sighted, with accuracy not recorded no record of accuracy of radio relocation
29. Activity:

| $\mathrm{A}=$ agonistic | $\mathrm{O}=$ other |
| :--- | :--- |
| $\mathrm{B}=$ bedded | $\mathrm{P}=$ apparent den site (bear |
| $\mathrm{D}=$ at den site | $\mathrm{n}=$ not seen) |
| $\mathrm{E}=$ digging | $\mathrm{R}=$ running |
| $\mathrm{F}=$ feeding | $\mathrm{S}=$ standing |
| $\mathrm{H}=$ hiding | $\mathrm{T}=$ treed |
| $\mathrm{I}=$ in den | $\dot{W}=$ walking |
| $\mathrm{J}=$ den of unmarked bear | $\mathrm{X}=$ swimming |
| $\mathrm{M}=$ mating | $\mathrm{Y}=$ fishing |
| $\mathrm{N}=$ nursing | $\mathrm{Z}=$ sitting |

42-45. Elevation: The elevation of the terrain upon which the animal was sighted, expressed in feet; up to four digits.
46. Slope: A code for the range of slope of the terrain upon which the animal was sighted.

```
F=flat (0' - 10足 ( R = w/in riverbank
G = gentle (1\mp@subsup{1}{}{\circ}-3\mp@subsup{0}{}{\circ})
M = moderate (31' - 60 )
S = steep (61' - 90 )
```

48-49. Aspect: A code for the general direction of exposure of the terrain upon which the animal was sighted: $N$, NW, E, SE, S, SW, W, NW, or
$\mathrm{F}=\mathrm{flat}$
$\mathrm{R}=$ ridgetop
$G=g u l l y$
the code is left-justified.

55-56. Number of young/age class: The number of young within a specific age class, for as many as two different age classes, sighted with (and directly associated with) the reported individual. Right-justified.

0 = young-of-the-year
1 = yearlings
2 = 2-year-olds
57-58. Same as 55-56, used if more than 1 age class of young is with bear.

59-62. Group size: The total number of individuals (of the same species) sighted within the group associated with the reported individual. Always will be at least 1 unless bear not seen (in this case leave blank).

63-65. Number of adult males: The total number of adult males (of the same species) within the group sighted in association with the reported individual.

66-68. Number of adult females: The total number of adult females (of the same species) within the group sighted in association with the reported individual.

69-71. Number of young: The total number of offspring (of the same species) within the group sighted in association with the reported individual.
72. Other species: If another species with the individual, enter the code for that species (see \#1).
74. Status:
$A=$ probably dead or shed
$B=$ capture site of new bear or bear w/o functioning transmitter
C = see comment (use for "special" points)
D = known nonhunter mortality
$\mathrm{F}=$ probably subsequent collar failure
$\mathrm{H}=$ known hunter kill subsequently
$\mathrm{S}=$ known shed collar
$\mathrm{U}=$ uncollared, but marked bear
86. Species: A code for the species of a killed animal on which the recorded predator was found.

B = beaver
$\mathrm{C}=$ caribou
$\mathrm{F}=\mathrm{f} 1 \mathrm{sh}$
$\mathrm{H}=$ snowshoe
$\mathrm{M}=$ moose

```
S = small mammal
U = unidentified
O = other
Age class: A code for the estimated age of the prey.
0 = young-of-the-year
1 = yearling
2 = 2-year-old
3 = adult
4 = unknown
88. Sex: Sex of the prey animal.
M = Male
F = Female
U - Unknown
89. Killed by: A code for the species which actually killed the
prey, or how it was killed.
U = unknown W}=\mathrm{ wolf
B = black bear }\quadV=\mathrm{ wolverine
G = grizzly A = accidental
S = winter kill O = other
90. Freshness:
    F = fresh
0 = old
Percent consumed: The approximate percent of the prey that has been consumed.
```


## 95-100. Habitat:

```
\begin{tabular}{|c|c|c|}
\hline SPRUCE & SHRUBLANDS & TUNDRA \\
\hline 1. Sparse-TALL & 10. Riparian willow & 18. Sedge-grass \\
\hline 2. Mod.-TALL & 11. Upland willow & 19. Alpine herbaceous \\
\hline 3. Dense-TALL (rip.) & 12. Willow/birch & 20. Shrub (d. birch) \\
\hline 4. Sparse-MEDIUM & 16. Alder & 21. Mat \& Cushion \\
\hline 5. Mod.-MEDIUM & & \\
\hline 6. Dense-MEDIUM & OTHER & OTHER FOREST \\
\hline \multicolumn{3}{|l|}{7. Sparse-LOW} \\
\hline 8. Sparse-LOW & 15. Marsh & 13. Aspen \\
\hline \multirow[t]{3}{*}{9. Dense-LOW} & 17. Rock/ice/snow & 14. Ripar. hardwood \\
\hline & 22. Gravel bar & 23. Mixed birch/spruce \\
\hline & & 24. Birch (trees) \\
\hline
\end{tabular}
```

101. Movement: codes for suspected direction of bear movements, inferred after the fact, based on best guess.
$\mathrm{N}=\mathrm{No}$ specialized movements suspected
$B=$ In seasonal activity area -- caribou calving grounds
$\mathrm{C}=$ En route to or from caribou calving grounds
$\mathrm{D}=\mathrm{In}$ season activity area -- salmon fishing area
$\mathrm{E}=\mathrm{En}$ route to or from salmon fishing area
F = In seasonal activity area searching for food resources that are scarce in that year within normal home range (especially bad berry years) -- summer feeding grounds
$G=E n$ route to or from above area
$\mathrm{H}=\mathrm{In}$ seasonal activity area -- denning behavior outside of known nondenning range
I = En route to or from above denning area
$\mathrm{J}=\mathrm{In}$ seasonal activity area -- generalized early spring lowland foraging
$\mathrm{K}=$ Suspected dispersal movements
$\mathrm{L}=$ Initial capture site or recapture site of nonradioed bear
$M=A t$ or en route to or from den site within normal home range
$0=$ Movement outside normal area based on suspected reproductive activity
102. Reproductive status codes -- Inferred after the fact, based on subsequent sightings.

A = With newborn cubs
B = With yearling offspring
C = With 2-year-old offspring
D = With 3-year-old offspring
$E=$ Presence or absence of offspring unknown (had them previously but not subsequently)
F = Probable or known estrous female or breeding male (usually accompanied by another bear in the case of males)
$\mathrm{G}=$ Inactive, unknown or alone (cubs lost or weaned)
H = Subadult
$\mathrm{M}=$ Movement outside normal use based on suspected reproductive activity

Table 1. Brown bears captured in Susitna Dam Studies as of Nov. 1986.

Comments

| Capture |  |  | Wt. (pounds) | Date | Ear Tags |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Tattoo | Sex | Age (years) |  |  |  | Comments |
| (277) | F | 10.5 | 225* | 4/10/80 | 1065/1066 | w/2 ylgs, not marked, collar shed 80/81 den |
| (278) | M | 9.5 | 375* | 4/19/80 | -- -- | capture mortality |
| (279) | M | 9.5 | 400* | 4/20/80 | 1100/1099 | collar shed by 6/12/80, recaptured 5/18/83, shot 9/84 |
| 280 | M | 5.5 | 300* | 4/20/80 | 1097/1098 | recollar next spring |
| 214 | M | 4.5 | 300* | 4/22/80 | 1072/1071 | collar shed 9/9/80, recaptured 6/85 |
| 281 | F | 3.5 | 250* | 4/22/80 | 16175/15950 | not turgid, see 5/81 recapture |
| 282 | M | 4.5 | 325* | 4/22/80 | 1079/1080 | see 6/82 recapture |
| 283 | F | 12.5 | 280* | 4/22/80 | 690/689 | w2 @2.5: 284 and 285 |
| (284) | M | 2.5 | 180* | 4/22/80 | 1074/1073 | w/283 see 5/5/81 recapture |
| 285 | M | 2.5 | 180* | 4/22/80 | 687/688 | w/283 |
| 286 | M | 3.5 | 264 | 5/1/80 | 1081/1082 |  |
| 292 | F | 3.5 | 174 | 5/2/80 | 1322/1321 | turgid |
| (293) | M | (3.5) | 277 | 5/2/80 | 1116/1115 | recaptured 8/81, 5/83, shot spring '85 |
| (294) | M | 10.5 | 607 | 5/2/80 | -- -- | died on 8/6/81 recapture |
| (295) | M | 12. 5 | 589 | 5/3/80 | 1303/1304 | collar shed by 5/4/80 |
| 299 | F | 13.5 | 285 | 5/4/80 | 1109/1110 | w/2 ylgs, turgid, recaptured 5/7/81 |
| (297) | M | 1.5 | 65 | 5/4/80 | (1301/1302) | w/299, shot by hunter on 9/18/81 |
| 298 | M | 1.5 | 65 | 5/4/80 | 1318/1317 | W/299 |
| 306 | F | 3.5 | 163 | 5/4/80 | 1319/1320 | turgid |
| (308A) | M | 6.5 | 480 | 5/6/80 | (1126/1125) | shot 9/83 |
| (308B) | F | 5.5 | 240 | 5/6/80 | 1096/1095 | turgid (?) - died on 8/6/81 recapture |
| 309 | M | 12.5 | 600 | 5/6/80 | 1117/1118 | collar shed by 5/14/80, recaptured 6/85 |
| (312) | F | 10.5 | 319 | 5/7/80 | 1312/1311 | w/311 |
| (311) | M | 2.5 | 227 | 5/7/80 | -- $=$ | shot on 9/16/80 |
| 313 | F | 9.5 | 286 | 5/7/80 | 1119/1120 | w/314 @2.5 |
| 314 | F | 2.5 | 154 | 5/7/80 | 1049/1050 | w/313, recaptured 6/1/85 |
| 315 | F | 2.5 | 90* | 5/7/80 | 1127/1128 | alone, recaptured 5/18/83 |
| (284\#2) | M | 3.5 | 125 | 5/5/81 | 1074/1073 | near $283 \mathrm{w} / 2 \mathrm{c}$, shot by hunter on 5/18/81 |
| (331) | F | 6.5 | 172 | 5/5/81 | (1296/1295) | w/332 and 333, died August 1982 |
| (332) | M | 2.5 | 79 | 5/5/81 | (1215/1216) | W/331 and 333, shot by hunter on 9/5/82 |
| (333) | M | 2.5 | 67 | 5/5/81 | (1240/1239) | $w / 331$ and 332, shot by hunter on 9/3/81 |
| 334 | F | 10.5 | 325 | 5/5/81 | 1292/1291 | estrous, missing in 1982 |
| 335 | F | 3.5 | 194 | 5/5/81 | 1220/1219 | recaptured 5/14/83 and 6/86, age changed +1 '83 tooth |
| 281\#2 | F | 4.5 | -- | 5/6/81 | 1201/1202 | estrous? recaptured 5/15/83 |
| 283\#2 | F | 13.5 | 261 | 5/6/81 | 1089/1090 | W/338 and 339, recaptured 5/14/83 |
| 338 | F | 0.5 | 12 | 5/6/81 | 1224/1223 | w/283, sex switched to female |
| (339) | M | (0.5) | 13 | 5/6/81 | 1222/1221 | w/283, recaptured 6/85, sex switched to male; shot 9/85 |
| 312\#2 | F | 11.5 | 280 | 5/6/81 | 1300/1299 | w/2c @0.5 (not captured), recaptured 5/14/83 |
| $313 \# 2$ | F | 10.5 | 284 | 5/6/81 | 1120/1119 | w/336, recaptured 5/14/83 |
| 336 | F | 0.5 | -- | 5/6/81 | 1237/1238 | w/313, not drugged (abandoned) |
| 337 | F | 13.5 | 321 | 5/6/81 | 1294/1293 | w/3c reunited on 5/9/81, recaptured 5/14/83 |
| 340 | F | 3.5 | 190 | 5/6/81 | 1225/1218 | not estrous, recaptured 5/15/83 |
| 280\#2 | M | 6.5 | 394 | 5/7/81 | 1097/1267 | w/F 341, recaptured 5/16/83 |
| 341 | F | 6.5 | 224 | 5/7/81 | (1208/1207) | W/M 280, collar failed, recaptured 6/81 |
| 299\#2 | F | 14.5 | 291 | 5/7/81 | 1109/1110 | w/2 @2.5 (297 and 298 - not recaptured), not estrous, recaptured $8 / 6 / 81$ |
| (342A) | M | 2.5 | 220 | 5/7/81 | 1228/1227 | alone, see 5/25/82 recapture, died 7/84 |
| 344 | F | 5.5 | - | 5/8/81 | 1204/1203 | W/2 cubs subsequently, recaptured 5/14/83 |
| (345) | M | 7.5 | 495 | 5/8/81 | -- -- | capture mortality |

Table 1. (continued)


Table 1. (continued)


Table 1. (continued)


* Weight estimated, () indicates shed collar or dead bear; \# recapture; - collar or mark replaced subsequently; last tattoo $=425$; last cub $=\# 25$.

Table 2. Black bears captured in Susitna Dam Studies as of Nov. 1986


Table 2. (continued)


Table 2. (cont inued)


Table 2. (continued)

| Tattoo | Capture |  | Wt. (pounds) | Date | Ear Tags | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 412\#3 | M | 2.0 | 80* | 2/25/85 | same | w/361 in den, applied green visual dropoff |
| 413\#3 | F | 2.0 | 65* | 2/25/85 | same | w/361 in den, applied red visual dropoff |
| 414\#3 | F | 2.0 | 55* | 2/25/85 | same | w/361 in den, applied white visual dropoff |
| 349\#3 | F | 8.0 | -- | 2/28/85 | same | in den w/at least 2@0, collar loosened $\mathrm{l}_{2}$ |
| 001 | M | 0.0 | 1.8 | 2/28/85 | -- | w/349, at least one stbling not handled |
| 289\#4 | F | 14.3 | -- | 3/1/85 | same | w/at least $2 @ 0$ in den, cubs not handled |
| 328\#3 | F | 11.3 | -- | 3/29/85 | same | w/3@0 in den, loosened collar $1 \frac{1}{2}$ notches, rubbed |
| 002 | M | 0.3 | 5.0 | 3/29/85 | -- | w/B328 and siblings |
| 003 | M | 0.3 | 4.1 | 3/29/85 | -- | w/B328 and siblings |
| 004 | F | 0.3 | 4.1 | 3/29/85 | -- | w/B328 and siblings |
| 404\#2 | F | 13.3 |  | 3/30/85 | same | w/3@O in den, collar fine, died (shot?) spring 1985, coys dispatched |
| 005 | M | 0.3 | 4.1* | 3/30/85 | -- | w/B404 and stiblings |
| 006 | M | 0.3 | 4.1* | 3/30/85 | -- | w/B404 and siblings |
| 007 | F | 0.3 | 3.5* | 3/30/85 | -- | w/B404 and siblings |
| (426) | M | (2.5) | 75* | 6/1/85 | -- | capture mortality |
| 428 | M | 5.5 | 175* | 6/1/85 | 2109/2167 | rot-away canvas spacer |
| 430 | M | 9.5 | 285* | 6/2/85 | (2093/2088) | rot-away canvas spacer, pulled off collar 1986 |
| 431 | F | 11.5 | 116 | 6/2/85 | 1519/1520 | ---- |
| $310 \# 2$ | M | 7.5 | 225* | 6/2/85 | 2185/2183 | rot-away canvas spacer. |
| 432 | F | 6.5 | 124 | 6/2/85 | 1558/1557 | w/ylg. 434 |
| 434 | F | 1.5 | 33 | 6/2/85 | 1552/1572 | w/B432 |
| 433 | M | 3.5 | 68* | 6/2/85 | 1647/2081 | ---- |
| (435) | M | (7.5) | 200* | 6/2/85 | 2182/2186 | ----, shot 9/85 |
| 436 | $\stackrel{M}{\text { M }}$ | 2.5* | 40* | 6/3/85 | --/2121 | w/B364-mother? |
| 438 | F | 8.5 | 130* | 6/3/85 | 1516/1521 | w/B439 \& uncaptured sibling |
| 439 | M | 2.5* | 40* | 6/3/85 | --1-- | w/B438-and sibling, dart injured leg |
| 441 | F | 9.5 | 195 | 6/4/85 | 2361/2362 | ---- |
| 351\#2 | M | 4.5 | 140 | 6/4/85 | 2169/2175 | old tags left in too (516/515) |
| 444 | M | 3.5 | 78 | 6/4/85 | 2154/2153 | dropoff visual collar |
| 445 | M | 8.5 | 250* | 6/4/85 | 2068/2164 | dropoff collar |
| (446) | F | 5.5 | 99 | 6/5/885 | --/- | capture mortality |
| 448 | F | 6.5 | 100 | 6/5/85 | 1544/1533 | break-away collar |
| 318\#4 | F | 10.5 | -- | 6/5/85 | same | w/2@l (not captured), recapture, old collar shed |
| 449 | M | 6.5 | 165* | 6/9/85 | 1640/2188 | alone |
| 451 | F | 2.5 | 54 | 6/10/85 | 2408/2484 | alone |

* Weight or age estmated; () shed or replaced collar or dead bear; \# recapture; __subsequently changed; last tattoo used = 425; last $\mathrm{cub}=25$.

Table 3. Number of observations of radio-marked brown bears (older than 2.0 years) within nested proximity zones of the Watana Impoundment (den-related activies are not included).

ZONE 1
ZONE 2
ZONE 3
ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) ( $1-5$ miles) (over 5 miles) TOTAL

| 1. | April 1-30 | 6 | 1 | 8 | 9 | 24 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2. | May 1-15 | 12 | 8 | 19 | 69 | 108 |
| 3. | May 16-31 | 31 | 27 | 65 | 108 | 231 |
| 4. | June 1-15 | 70 | 67 | 154 | 89 | 380 |
| 5. | June 16-30 | 45 | 35 | 104 | 69 | 253 |
| 6. | July 1-15 | 6 | 8 | 39 | 37 | 90 |
| 7. | July 16-31 | 4 | 14 | 61 | 42 | 121 |
| 8. | August 1-15 | 4 | 11 | 41 | 44 | 100 |
| 9. | $\begin{gathered} \text { August } 16- \\ \text { March } 31 \end{gathered}$ | 26 | 22 | 97 | 168 | 313 |
|  | TOTALS | 204 | 193 | 588 | 635 | 1620 . |
| Area within zone |  |  |  |  |  |  |
|  | \% | 9.26 | 19.02 | 71.72 | -- | 100.0 |

Value of Chi-Square test of the null hypothesis that use of each zone is equivalent to expected values based on the area of each zone for:

| Period | ZONE 1 |  | ZONE 2 |  | ZONE 3 |  | $\mathrm{X}^{2}$ | d.f. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | obs. | E(x) | obs. | E(x) | obs. | E(x) |  |  |
| All months | 204 | 91.2 | 193 | 187.4 | 588 | 706.4 | 160** | 2 |
| April l-June 30 | 164 | 60.4 | 138 | 124.0 | 350 | 467.6 | 209** | 2 |
| July 1-March 31 | 40 | 30.8 | 55 | 63.3 | 238 | 238.8 | 3.9 | 2 |

[^0]Table 4. Number of observations of radio-marked male brown bears (older than 2.0 years) within nested proximity zones of the Watana Impoundment (den-related activies are not included).

ZONE 1 ZONE 2 ZONE 3 ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) ( $1-5$ miles) (over 5 miles) TOTAL


Value of Chi-Square test of the null hypothesis that use of each zone is equivalent to expected values based on the area of each zone for:

| Period | ZONE 1 |  | ZONE 2 |  | ZONE 3 |  | $\mathrm{X}^{2}$ | d.f. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | obs. | $\mathrm{E}(\mathrm{x})$ | obs. | $\mathrm{E}(\mathrm{x})$ | obs. | $\mathrm{E}(\mathrm{x})$ |  |  |
| All months | 64 | 27.7 | 69 | 56.9 | 166 | 214.4 | $61.1 * *$ | 2 |
| April 1-June 30 | 50 | 20.2 | 55 | 41.5 | 113 | 156.4 | 60.4** | 2 |
| July 1-March 31 | 14 | 7.5 | 14 | 15.4 | 53 | 58.1 | 6.2** | 2 |

* Reject null hypothesis, $P$ less than 0.10 .
** Reject null hypothesis, $p$ less than 0.05 .

Table 5. Number of observations of radio-marked female brown bears (older than 2.0 years) within nested proximity zones of the Watana Impoundment (den-related activies are not included).

ZONE 1 ZONE 2 ZONE 3 ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) (1-5 miles) (over 5 miles) TOTAL

| 1. | April 1-30 | 2 | 1 | 5 | 6 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2. | May 1-15 | 6 | 5 | 13 | 42 | 66 |
| 3. | May 16-31 | 22 | 14 | 26 | 67 | 129 |
| 4. | June 1-15 | 53 | 27 | 81 | 47 | 208 |
| 5. | June 16-30 | 24 | 24 | 62 | 36 | 146 |
| 6. | Ju1y 1-15 | 4 | 4 | 23 | 20 | 51 |
| 7. | July 16-31 | 1 | 9 | 37 | 22 | 69 |
| 8. | August 1-15 | 3 | 7 | 25 | 26 | 61 |
| 9. | $\begin{gathered} \text { August } 16- \\ \text { March } 31 \end{gathered}$ | 21 | 14 | 55 | 86 | 176 |
|  | TOTALS | 136 | 105 | 327 | 352 | 920 |
| Area within zone     <br> $\left(\mathrm{km}^{2}\right)$ 159.32 327.07 1233.51 -- |  |  |  |  |  |  |
|  | \% | 9.26 | 19.02 | 71.72 | -- | 100.0 |

Value of Chi-Square test of the null hypothesis that use of each zone is equivalent to expected values based on the area of each zone for:

| Period | ZONE 1 |  | ZONE 2 |  | ZONE 3 |  | $\mathrm{X}^{2}$ | d.f. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | obs. | $E(x)$ | obs. | $\mathrm{E}(\mathrm{x})$ | obs. | $\mathrm{E}(\mathrm{x})$ |  |  |
| All months | 136 | 52.6 | 105 | 108.0 | 327 | 407.4 | 148** | 2 |
| April 1-June 30 | 107 | 33.8 | 71 | 69.4 | 187 | 261.8 | 180** | 2 |
| July 1-March 31 | 29 | 18.8 | 34 | 38.6 | 140 | 145.6 | 6.3** | 2 |

[^1]Table 6. Number of observations of radio-marked female brown bears with coy (on 15 June) within nested proximity zones of the Watana Impoundment (den-related activies are not included).

ZONE 1
ZONE 2
ZONE 3
ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) ( $1-5$ miles) (over 5 miles) TOTAL


* Reject nu11 hypothesis, $p$ less than 0.10 .
** Reject null hypothesis, p less than 0.05 .

Table 7. Chi-square test of null hypothesis that the proportion of observations in impoundment proximity zones is the same, for a group of radio-marked female brown bears, during years when they have cubs-of-the-year ("coy") as during years when they do not. (Includes both impoundments, lumps years 1980-1984, cub status is on 15 June, and observation associated with den-related activities are not included).

| Females without coy |  |
| :--- | :--- | :--- |
| No. of <br> observations | $\%$ |$\quad$| Females with coy |
| :--- |


| Proximity Zone 1 (inundation area) | 59 | 18.7 | 8 | 30.1 |
| :---: | :---: | :---: | :---: | :---: |
| Proximity Zone 2 (impoundment shoreline - 1 mile) | 58 | 18.4 | 32 | 29.4 |
| ```Proximity Zone 3 (1-5 miles from impoundment shore- 1ine)``` | 198 | 62.9 | 120 | 100.6 |
| Totals: | 315 | 100\% | 160 | 160.1 |
|  |  | Chi | $\begin{aligned} & \text { d.f } \\ & \text { sigı } \end{aligned}$ | less |

## BEARS INCLUDED:

Bear ID

283

299
'312
313
335

337
340
341
344

384
83
81

81, 84
81, 82, 83
84
82
81, 83

84

Table 8. Number of observed and expected observations of radio-marked brown bears (excluding females with coy and bears less than 2.0 years old) within nested impoundment proximity zones of the Devils Canyon Impoundment (den-related activities are not included).

ZONE 1 ZONE 2 ZONE 3 ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) ( $1-5$ miles) (over 5 miles) TOTAL

| All males $\quad 4$ | 17 | 38 | 107 | 166 |
| :---: | :---: | :---: | :---: | :---: |
| All females 10 | 76 | 165 | 174 | 425 |
| All females without cubs-of-year | 76 | 161 | 158 | 405 |
| totals |  |  |  |  |
| $\begin{aligned} & \text { Area within zone } \\ & \left(\mathrm{km}^{2}\right) \end{aligned}$ | 164.78 | 689.01 | -- | 882.71 |
| \% 3.28 | 18.67 | 78.06 | -- | 100.0 |

Value of Chi-Square test of the null hypothesis that the use of each zone is equivalent to expected values based on the area of each zone for:

|  | ZONE 1 |  | ZONE 2 |  | ZONE 3 |  | $\mathrm{X}^{2}$ | d.f. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sex group | obs. | $E(x)$ | obs. | E(x) | obs. | E(x) |  |  |
| Males and females w/o cubs (whole year) | 14 | 10.0 | 93 | 57.1 | 199 | 238.9 | 30.8** | 2 |
| Males (whole year) | 4 | 1.9 | 17 | 11.0 |  | 46.1 | 3.0 | 2 |
| Females w/o cubs | 10 | 8.1 | 76 | 46.1 | 161 | 192.8 | 25.1** | 2 |

* Reject null hypothesis, p less than 0.10.
** Reject null hypothesis, p less than 0.05 .

Table 9. Number of brown bear point locations in each of 4 impoundment proximity zones from 1 April-15 June. All years lumped and both impoundments lumped, subadult dispersers and bears from downstream study area are not included.

| Bear <br> ID | Sex | Zone 1 | Zone 2 | Zone 3 | Zone 4 | A11 Zones |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 279 | M | 1 | 1 | 8 | 26 | 36 |
| 280 | M | 13 | 8 | 23 | 7 | 51 |
| 282 | M | 1 | 2 | 13 | 4 | 20 |
| 293 | M | 1 | 0 | 1 | 7 | 9 |
| 294 | M | 1 | 3 | 1 | 1 | 6 |
| 382 | M | 11 | 12 | 3 | 5 | 31 |
| 399 | M | 2 | 4 | 15 | 11 | 32 |
| 400 | M | 0 | 1 | 14 | 13 | 28 |
| 422 | M | 0 | 11 | 13 | 1 | 25 |
| A11 | Males | 30 | 42 | 91 | 75 | 238 |
| \% |  | 12.6 | 17.6 | 38.2 | 31.5 | 100 |
| 281 | F | 25 | 12 | 21 | 9 | 67 |
| 283 | F | 1 | 5 | 17 | 30 | 53 |
| 299 | F | 29 | 5 | 8 | 9 | 51 |
| 312 | F | 1 | 2 | 5 | 18 | 26 |
| 313 | F | 2 | 9 | 0 | 43 | 54 |
| 315 | F | 0 | 5 | 6 | 0 | 11 |
| 331 | F | 1 | 2 | 2 | 6 | 11 |
| 334 | F | 0 | 0 | 10 | 11 | 21 |
| 335 | F | 0 | 0 | 12 | 32 | 44 |
| 337 | F | 0 | 0 | 1 | 27 | 28 |
| 340 | F | 9 | 19 | 28 | 10 | 66 |
| 341 | F | 7 | 5 | 6 | 0 | 18 |
| 344 | F | 0 | 2 | 9 | 8 | 19 |
| 379 | F | 0 | 0 | 0 | 9 | 9 |
| 381 | F | 5 | 8 | 15 | 4 | 32 |
| 384 | F | 0 | 1 | 1 | 5 | 7 |
| 385 | F | 0 | 0 | 0 | 14 | 14 |
| 388 | F | 0 | 0 | 12 | 17 | 29 |
| 394 | F | 2 | 6 | 7 | 0 | 15 |
| 395 | F | 2 | 0 | 3 | 1 | 6 |
| 396 | F | 0 | 1 | 9 | 1 | 11 |
| 420 | F | 0 | 18 | 11 | 0 | 29 |
| 423 | F | 0 | 0 | 5 | 0 | 5 |
| 425 | F | 2 | 4 | 7 | 0 | 13 |
| 308 | F | 2 | 5 | 4 | 0 | 11 |
| $\begin{aligned} & \text { All Females } \\ & \% \end{aligned}$ |  | 88 | 109 | 199 | 254 | 650 |
|  |  | 13.5 | 16.8 | 30.6 | 39.1 | 100 |
| ALL | BEARS | 118 | 151 | 290 | 329 | 888 |
| \% |  | 13.3 | 17.0 | 32.7 | 37.0 | 100 |

Table 10. Number of Susitna river crossings by radio-marked brown bears, 1980-1984. Includes only years with $>5$ observations.


Table 10. (cont'd)


Table 10. (cont'd)


[^2]*** excludes observations at den sites.

Table 11. Annual use of Pratrie Ck. area by radio-collared brown bears during July and August king salmon
spawning period (1980-1985). Reproductive status reflects July data for females (c=newborn cubs).

| Males (age in year first captured) | 1980 | 1981** | 1982 | 1983 | 1984*** | 1985**** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 214 @ 4 (80) | no | shed | - | - | - | no |
| 279 @ 9(80) | ND (shed) | ND | ND | yes | yes | dead |
| 280 @ 5 (80) | no | no | no | no | no | no collar |
| 282@ 4(80) | - | - | yes | yes | yes | yes |
| 293 ( 3 (80) | yes | yes | yes | no | (shed) | - |
| 294 @ 10(80) | yes | yes | - (dead) | - | - | - |
| 342a*@ 2(81) | - | no | no | no | yes (dead) | - |
| 373 @ 9(82) | - | - | yes | ND (shed) | - | - |
| 382 @ 2 (84) | - | - | - | - | - | yes |
| 386 @ 2 (83) | - | - | - | no | dead | - |
| 389 @ 2(83) | - | - | - | no | dead | - |
| 390 @ 2(83) | - | - | - | no | missing | - |
| 391 @ 2(83) | - | - | - | no | dead | - |
| 392@ ${ }^{\text {(83) }}$ | - | - | - | no | dead | - |
| 399 @ 9(83) | - | - | - | yes | yes | missing |
| 400 @ 20 (83) | - | - | - | no | no | missing |
| 422 ( ${ }^{\text {( }} 84$ ) | - | - | - | - | yes | dead |
| 427 @ A (85) | - | - | - | - | - | yes |

Subtotals for
MALES:
No. using Prairie Ck.

| (males) | 2 | 2 | 3 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Total No. of collared <br> males | 4 | 4 | 5 | 12 | 8 |

No. collared males
excluding subadult

\% males using Prairie
Ck. (excludes dis-
persers) $50 \quad 67 \quad 45$

Table 11. (cont.)

| Females (age in first captured) | 1980 | 1981** | 1982 | 1983 | 1984*** | 1985**** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 273 a 9(85) | - | - | - | - | - | no, alone |
| 277 @ 10(80) | no? | ND- (shed) | ND | ND | ND | ND |
| 281 @ 3 (80) | no, alone | no, alone | no, alone | no, alone | no, alone | no w/2c |
| 283 @ 12 (80) | yes, alone | no, w/2c | yes, alone | yes, alone | yes, alon | yes, w/2c |
| 299 @ 13 (80) | no, w/2@l | no, alone | no, alone | no, w/3c | no, w/3@1 | missing |
| 308b @ 5(80) | yes, alone | no?, alone | -dead | - | - | - |
| 312 @ 10(80) | no, alone | no, w/lc | no, w/1@1 | no, alone | dead | - |
| 313 @ 9(80) | no, alone | no, alone | no, w/2c | no, w/l@ | no, alone | missing |
| 314 @ 7(85) | - | - | - | - | - | no, alone |
| 315 @ 2 (80) | - | - | - | yes, alone | yes, alone | missing |
| 331 @ 6(81) | - | no, alone | -dead | - | - | - |
| 334 @ 10(81) | - | no, alone | -missing | - | - | - |
| 335 @ 2 (81) | - | no, alone | no, alone | no, alone | no, w/2c | no, w/2@1 |
| 337 @ 13 (81) | - | no, w/3c | no, w/l@ | no, alone | no, w/2c | no, w/2@1 |
| 340 @ 3(81) | - | no; alone | no, alone | no, alone | no, w/2c | no, w/2@1 |
| 341 @ 6(81) | - | no, alone | no, w/ 2 c | -missing | - | no, alone |
| 344 @ 5(81) | - . | no, w/2c | no, wl@l | no, alone | no, alone | missing |
| 379* @ 5(82) | - | - | no, w/2c* | no, w/2@1* | no, alone?* | no, alone* |
| 380 @ 15 (82) | - | - | yes, w/2@l | yes, alone | dead | - |
| 381 @ 3(82) | - | - | no, alone | no, alone | no, alone | no, w/2c |
| $384 @ 12(83)$ | - | - | - | - | no, w/2c | missing |
| 385 @ 2 (83) | - | - | - | no, alone | no, alone | no collar |
| 388 @ 14(83) | - | - | - | no, alone | no, alone | no, w/2c |
| 393 @ 2(83) | - | - | - | no, alone | dead | - |
| 394 @ 6(83) | - | - | - | yes, alone | yes - dead | - |
| 395 @ 3(83) | - | - | - | no, alone | dead | - |
| 396 @ 13 (83) | - | - | - | yes, alone | yes, alone | yes, alone |
| 397 @ 4 | - | - | - | - | - | yes, alone |
| 398 @ 4 | - | - | - | - | - | yes, alone |
| 403* @ 6(83) | - | - | - | no, w/2c* | no, w/l@1?* | no, alone |
| 407* @ 4 (83) | - | - | - | yes, alone* | yes, alone* | yes, alone |
| 420 @ 19 (84) | $\sim$ | - | - | - | yes, w/2@1 | yes, alone |
| 423 (1) (84) | - | - | - | - | yes, w/3c | yes, w/3/@l |
| 425 @ A (84) | - | - | - | - | no, alone | no, w/2c |
| 437 @ 2 (85) | - | - | - | - | - | no, alone |
| 447 @ A (85) | - | - | - | - | - | no, alone |

Table 11. (cont.)

| Females (age in year first captured) | 1980 | 1981** | 1982 | 1983 | 1984*** | 1985**** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Subtotals for FEMALES <br> No. using Prairie Ck. (females) |  |  |  |  |  |  |
|  | 2 | 0 | 2 | 6 | 7 | 7 |
| Total No. of collared females | 7 | 13 | 13 | 22 | 21 | 21 |
| \% females using Prairie Ck. | 29 | 0 | 15 | 27 | 33 | 33 |
| FOTALS: |  |  |  |  |  |  |
| No. bears using Prairle Ck. | 4 | 2 | 5 | 9 | 11 | 10 |
| No. bears radio-collared (excluding dispersing males) | 11 | 16 | 17 | 29 | 29 | 25 |
| \% bears using Prairie Ck. | 36 | 13** | 29 | 31 | 38 | 40 |

Bear occurs in the downstream study area
** Poor monitoring conditions in 1981
*** Intensively monitored in 1984
**** No routine monitoring, monitored only on $7 / 23-27$ and $8 / 6$ because of study termination

Table 12. Reșults of brown bear census on Prairie Creek in 1984. Flights started at 0800 hrs. and pilot Al Lee flew the plane. Bear IDs are given in parentheses. Includes only bears older than 2.0.


Table 13. Brown bear census on Prairie Creek, July-August 1985.


[^3]Table 14. Estimated average number of brown bears using Prairie Creek during the salmon run in 1985 based on bear-days estimator.

| Date | ${ }^{n} 1$ | $\mathrm{m}_{2}$ | $\mathrm{n}_{2}$ | $\begin{gathered} \mathrm{Cum} . \\ \mathrm{n}_{1} \end{gathered}$ | $\begin{gathered} \text { Cum. } \\ \mathrm{m}_{2} \end{gathered}$ | $\begin{gathered} \text { Cum. } \\ \mathrm{n}_{2} \end{gathered}$ | $N^{*}=$ Est. <br> No. bears | $\begin{aligned} & 95 \% \mathrm{CI}= \\ & +/- \text { bears } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7/23pm | 4 | 1 | 10 | 4 | 1 | 10 | 26.50 | 21.80 |
| 7/24am | 4 | 0 | 9 | 8 | 1 | 19 | 44.50 | 42.60 |
| 7/24pm | 4 | 2 | 28 | 12 | 3 | 47 | 51.67 | 36.31 |
| 7/25am | 4 | 3 | 29 | 16 | 6 | 76 | 46.50 | 23.69 |
| 7/25pm | 4 | 1 | 21 | 20 | 7 | 97 | 51.25 | 25.34 |
| 7/26am | 4 | 1 | 13 | 24 | 8 | 110 | 51.22 | 24.43 |
| 7/26pm | 3 | 0 | 26 | 27 | 8 | 136 | 60.75 | 30.05 |
| 7/27am | 4 | 1 | 13 | 31 | 9 | 149 | 59.88 | 28.40 |
| 8/06am | 15 | 5 | 20 | 46 | 14 | 169 | 59.07 | 22.85 |

Table 15. Summary of Nelchina Basin brown bear litter size data for cubs-of-the-year (based on spring observations of radio-collared bears).

| $\begin{gathered} \text { BEAR ID } \\ \text { (year-age) } \end{gathered}$ | LITTER SIZE (COY) (year) | COMMENTS | USABLE SUMMARY |
| :---: | :---: | :---: | :---: |
| 207 (1978, 11) | 3 (1978) | When last seen on $10 / 7 / 78$ had all 3 cubs on $5 / 31 / 79$, had only 1 ylg . which stayed with her until last observation on 9/12/79 | 2 of 3 lost |
| 213 (1978, 10) | 2 (1979) | Lost apparent ylg. due to 1978 capture, had newborns when transplanted in 1979, lost these 8-16 days after release, bear apparently died in study area after return | none-transplant bias |
| -231 (1979, 13) | 3 (1979) | Turgid in 1978, bred, lost 2 of 3 cubs by $6 / 11 / 79$, survivor lived at least until last observation on $8 / 3 / 79$ (no exit data in 1980) | 2 of 3 lost |
| 206 (1978, 13) | 3 (1979) | Lactating female with male in 1978, during last observation prior to shedding collar the cubs were not seen but undergrowth was thick (6/17/79) | none |
| 313 (1981, 10) | 1 (1981) | Bear had a 2-year-old offspring in 1980, lost cub (possible capture-related) | 1 of llost (capture related?) |
| 313 (1982, 11) | 2 (1982) | Both survived | 0 of 2 lost |
| 312 (1981, 11) | 2 (1981) | Had a 2-year-old in 1980, lost 1 cub by $6 / 18$, other weaned in 1983 | 1 of 2 lost |
| 312 (1984, 14) | 3 (1984) | Capture-related losses (collared) | none |
| 283 (1981, 13) | 2 (1981) | Weaned 2@2 in 1980, lost 1 cub by $9 / 1$ other lost as ylg | $2 \text { lost }$ |

Table 15. (cont'd)

| $\begin{gathered} \text { BEAR ID } \\ \text { (year-age) } \end{gathered}$ | LITTER SIZE (COY) (year) | COMMENTS | USABLE SUMMARY |
| :---: | :---: | :---: | :---: |
| 283 (1983, 15) | 1 (1983) | Killed by brown bear by 5/17/83, cub was collared | 1 of 1 lost |
| 283 (1985, 17) | 2 (1985) | Both survived to den exit | 0 of 2 lost |
| 337 (1981, 13) | 3 (1981) | Cubs and female reunited, 1 cub lost in 81/82 den, other 2 survived to exit (1 weaned in 1983, other lost as ylg.) | 1 of 3 lost |
| 337 (1984, 16) | 2 (1984) | Both survived to den exit, collared cubs | 0 of 2 lost |
| 344 (1981, 5) | 2 (1981) | Both lost in '82 as yearlings | 0 of 2 lost |
| 344 (1983, 7) | 2 (1983) | Lost 1 in early July - other survived to den exit | 1 of 2 lost |
| 379 (1982, 5) | 2 (1982) | Both survived | 0 of 2 lost |
| 341 (1981, 6) | 2 (1982) | Survived until 7/15/82 when bear was lost | none |
| 341 (1986, 11) | 1 (1986) | Survived to August at least | -- |
| 299 (1980, 13) | 1 (1982) | Bear weaned 2@2 in 1981, cub lost by 6/9/62 | 1 of 1 lost |
| 299 (1983, 16) | 3 (1983) | All cubs collared, alive to den exit | 0 of 3 lost |
| 281 (1983, 6) | 2 (1983) | Both killed by brown bear by 6/1/83, cubs collared | 2 of 2 lost |
| 281 (1984, 7) | 2 (1984) | Lost both in May, 1 suspected killed by brown bear, other unknown (accidental drowning?), collared cubs | 2 of 2 lost |

Table 15. (cont'd)


Table 16. Summary of Nelchina Basin brown bear litter size data for litters of yearlings (based on spring observation of radio-collared bears).

| $\begin{gathered} \text { BEAR ID } \\ \text { (year-age) } \\ \hline \end{gathered}$ | LITTER SIZE (y1gs.) (year) | COMMENTS | SUMMARY |
| :---: | :---: | :---: | :---: |
| 220 (1978, 5) | 1 (1978) | Y1g. entered den and was weaned in 1979, bred | 0 of 1 lost |
| 221 (1978, 8) | 2 (1978) | Survived, weaned in 1979 | 0 of 2 lost |
| $234(1978,5)$ | 2(1978) | Paxson dump bear, lost apparent ylgs. between 6/23/78 and 8/4/78, reportedly had cubs in August 1979, radio failed | none |
| 240 (1979, 5) | 2 (1979) | Bear transplanted with ylgs., not known if ylgs., survived to return to study area, bear was alone on $7 / 18 / 80$ | none |
| 244 (1979, 6) | 1 (1979) | Thin female transplanted with ylg., $y 1 g$. survived at least 21 days, female bred, but alone in July and August 1980 | none-transplant bias |
| 251 (1979, 10) | 2 (1979) | Very large ylgs. lost $10-17$ days after transplant, bear had no cubs in 1980 (August) | none-transplant bias |
| $254(1979,9)$ | 2 (1979) | Female died after transplant (ylgs.??) | none |
| $261(1979,7)$ | 2 (1979) | Lost 1 ylg. between 1 and 7 days after transplant, other survived at least until Sept., didn't return to study area | none-transplant bias |
| 269 (1979, 16) | 2 (1979) | Transplanted, returned to study area with female, no cubs on $9 / 29 / 80$, shot in fall 1981 reportedly without cubs | none, transplant bias |
| 274 (1979, 11) | , 1 (1979) | Transplanted, no radio | none |
| 207 (1978, 11) | 1 (1979) | Survived until 9/12/79 | 0. of 1 lost |
| $231(1978,12)$ | 1 (1979) | Survived until 8/79 | none |
|  |  | (continued on next page) |  |

Table 16. (cont'd)

| $\begin{gathered} \text { BEAR ID } \\ \text { (year-age) } \end{gathered}$ | LITTER SIZE (ylgs.) (year) | COMMENTS | SUMMARY |
| :---: | :---: | :---: | :---: |
| 213 (1978, 10) | 1 (1978) | Apparent ylg. was not captured, had cubs following year | 1 of 1 lost (capture related?) |
| 277 (1980, 10) | 2 (1980) | Ylgs. visually aged, not captured, survived to enter den, no exit data as bear shed collar in den | 0 of 2 lost |
| 299 (1980, 13) | 2 (1980) | Both survived, weaned next year | 0 of 2 lost |
| 299 (1984, 17) | 3 (1984) | Survived with internals to exit from den | 0 of 3 lost |
| 312 (1982, 12) | 1 (1982) | Survived, weaned next year | 0 of 1 lost |
| 281 (1986, 9) | 1 (1986) | -- . . | -- |
| 283 (1982, 14) | 1 (1982) | Lost by 5/18/82 | 1 of 1 lost |
| 283 (1986, 18) | 2 (1986) | -- | -- |
| 337 (1982, 14) | 2 (1982) | Lost 1 by 6/17/82, other survived | 1 of 2 lost |
| 337 (1985, 17) | 2 (1985) | Survived to den exit | 0 of 2 lost |
| 380 (1982, 15) | 2 (1982) | Both survived to den entrance, at least 1 exited den and was weaned | 0 of 2 lost |
| 344 (1982, 6) | 2 (1982) | Lost 1 by 6/17, other by $7 / 26 / 82$ | 2 of 2 lost |
| 344 (1984, 8) | 1 (1984) | Lost 1 in May, sibling lost year before | l. of 1 lost |
| 313 (1983, 12) | 2 (1983) | Lost 1 (surgery related?) by 6/2/83; other survived thru October | 0 of 1 lost |
| 379 (1983, 6) | 2 (1983) | Lost 1 in June-September period | 1 of 2 lost |
| 420 (1984, 19) | 2 (1984) | Survived to den exit | 0 of 2 lost |



Table 17. Summary of Nelchina Basin brown bear litter size data for litters of 2 -year-olds (based on observations of radio-collared bears).

| $\begin{gathered} \text { BEAR ID } \\ \text { (year-age) } \end{gathered}$ | 2-year-old LITTER SIZE (year) | COMMENTS |
| :---: | :---: | :---: |
| 204 (1978, 7) | 2 (1978) | weaned by 6/19/78, bred |
| 283 (1980, 12) | 2 (1980) | weaned in mid-June, bred, new litter next year |
| 312 (1980, 10) | 1 (1980) | ```weaned right after capture in May, new litter in 1981``` |
| 312 (1983, 13) | 1 (1983) | weaned by $6 / 13$, bred |
| 313 (1980, 9) | 1 (1980) | weaned by May, bred, new litter in 1981 |
| 313 (1984, 13) | 1 (1984) | weaned in May, bred |
| 220 (1978, 5) | 1 (1979) | weaned by $6 / 17$, bred |
| 221 (1978, 8) | 2 (1979) | -- |
| 269 (1979, 16) | 2? (1980) | -- |
| 299 (1980, 13) | 2 (1981) | weaned in 5/81, new litter in 1982 |
| 337 (1983, 15) | 1 (1983) | weaned by 5/15, bred |
| 337 (1986, 18) | 2 (1986) | still with mother on 9/24/86 |
| 384 (1983, 12) | 3 (1983) | weaned by $6 / 13$, one of these 3 may not have been part of this litter, bred |
| 388 (1983, 14) | 2 (1983) | weaned by 6/13; bred |
| 396 (1983, 13) | 2 (1983) | weaned by $6 / 1$, bred |
| 331 (1981, 6) | 2 (1981) | weaned by $6 / 15$, bred, no cubs in 1982 , died in 1982 (reason?) |
| 379 (1984, 7) | 1 (1984) | apparently weaned cub (time?), bred |
| 314 (1986, 8) | 1 (1986). | bear lost in May ' 86 |
| 420 (1985, 20) | 2 (1985) | weaned in May |
| 423 (1985, 23) | 3 (1986) | 3 @ 2 in June 1986 |
| Summary |  |  |
| No. of 2-year-o |  | No. of litters Mean litter size (range) |
| 34 |  | 20 1.7 (1-3) |

Table 18. Brown bear offspring survivorship and weaning, GMU 13 studies, (excludes bears transplanted in 1979).

MOTHER'S ID (age in year when first captured)

| Year | G207 (11 in 1978) | G220 (5 in 1978) | G221 (8 in 1978) | G204 (7 in 1978) | G321 (12 in 1978) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 3 cubs, April-Oct. | 1 ylg., May-Oct. in June and bred | 2 ylgs., May-Oct. | 2 @ 2 in May, weaned | bred |
| 1979 | $\begin{aligned} & 1 \text { ylg., May-Sept. } \\ & 2 \text { ylgs., lost in } \\ & 78 / 79 \text { den? } \end{aligned}$ | 1 @ 2 , weaned in June | 2 @ 2 weaned | no data <br> in May, <br> radio failure | ```2 of 3 cubs lost in June, l survived April-Sept.``` |
| 1980 | no data | no data | no data | no data | no data |

(cont inued on next page)

Table 18. (cont'd)

MOTHER'S ID (age in year when first captured)

|  | Year | G277 (10 in 1980) | G312 (10 in 1980) | G299 (13 in 1980) | G313 (9 in 1980) | G283 (13 in 1980) | G281 (3 in 1980) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1980 | 2 @ 1 survived April thru August, collar shed in den | weaned 1 @ 2 in May, breeding not observed | 2 of 2 ylgs . survived May-Oct. | weaned 1 @ 2 in May, bred | weaned 2 @ 2 in June, bred | not estrous |
|  | 1981 | no data | 1 of 2 cubs lost in June, other survived MayOct. | weaned 2 @ 2 in May and bred | 1 @ 0 lost in May (capture related?) | 1 of 2 cubs lost in Aug., other survived | estrous, bred |
|  | 1982 | no data | yearling survived | lost 1 of 100 in June | 2 @ 0 survived | $\begin{aligned} & \text { lost } 1 \text { @ } 1 \text { in } \\ & \text { May, bred } \end{aligned}$ | alone, bred |
|  | 1983 | no data | weaned 1 @ 2 in June, bred, offspring $=$ G385, transmitted | 3 @ 0 survived (w/collars) | 1 a 1 lost in June (transmitted internally), sibling survived | lost 1 @ 0 in May, bred, lost cub had transmitter | 2 @ 0 lost in May (bear predation), not seen breeding |
| $$ | 1984 | no data | w/2 @ O-bear <br> killed in May | 3 @ 1 survived (w/internals) | 1 @ 2 weaned in May, shot | alone, bred | 2 © 0 lost in May, bred |
|  | 1985 | no data $\pm$ | -- | weaned 2-yr-olds collar failed? | - | 2 @ 0, survived | 2 @ 0, I lost in June, other survived |
|  |  | no data | -- | - | -- | 2 @ 1, survived | 1 @ 1, survived |

Table 18. (cont'd)

MOTHER'S ID (age in year when first captured)

(continued on next page)

Table 18. (cont'd)

MOTHER'S ID (age in year when first captured)

| Year | G380 (15 in 1982) | G394 (6 in 1983) | G384 (12 in 1983) | G379 (5 in 1982) | G388( 14 in 1983) | G381 (3 in 1982) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 2 @ 1 survived until denning, one may have died in den | no data | no data | $2 @ 0$ survived | no data | alone |
| 1983 | at least 1 @ 2 weaned in May, possibly both shot in Sept. | lost 1 @ 0 in May (?capture-related possible?), bred | weaned 2 or 3 @ 2 in June, bred | ```I of 2 survived, lost 2 (June - Sept.)``` | weaned 2 @ 2, | alone, bred |
| 1984 | -- | alone, shot | w/2 @ 0 thru <br> Sept., missing | ```probably weaned 1 @ 2 after May }2``` | $\begin{aligned} & \text { w/2 @ 0, } \\ & \text { capture-related } \\ & \text { cub loss, bred } \end{aligned}$ | alone, bred |
| 1985 | -- | -- | -- | alone, shot | w/2 @ 0, survived | w/2 c, survived |
| $\begin{aligned} & 1986 \\ & \text { (to } \\ & \text { Sept.) } \end{aligned}$ | -- | -- | -- | -- | w/2 @ 1 , survived | w/2@ 1 , survived |

Table 18. (cont'd)


Table 19. Summary of known losses from brown bear litters of cubs and yearlings. Losses dated from emergence in year indicated to emergence the following year. IDs of females included are indicated in parentheses.


Table 20. Morphometrics of brown bear cubs-of-the-year handled in GMU 13, 1978-1986.


Totals: 17 males and 14 females

Table 21. Morphometrics of brown bear yearlings handled in GMU 13, 1978-1986.


Table 22. Summary of reproductive intervals for brown bears by bear ID. Based on data in Table 18 , this report. Year of 1 itter and reason for intervals $>2$ years are indicated in parentheses"lost" means lost complete litter.

IDS OF BEARS WITH COMPLETE INTERVALS OF:
3 YEARS
4 YEARS
5 YEARS
6 YEARS

| 220(77)** | 335 (84) | 313(82, 1 lost) | 281(85, 2 1ost) | 283* (85, |
| :---: | :---: | :---: | :---: | :---: |
| 221(77)** | 340(84) |  |  | 1 lost @ age 1) |
| 314(84)** | 312(81) |  |  |  |
| 380(81)** | 337a (81) |  |  |  |
| 420(83)** | 337b (84) |  |  |  |
| 379 (82) | 388* (85) |  |  |  |
| 423 (84) | 381*(85) |  |  |  |

INCOMPLETE INTERVALS THAT WILL BE AT LEAST THE INDICATED LENGTH:

| 4 YEARS | 5 YEARS | 6 YEARS | 7 YEARS |
| :---: | :---: | :---: | :---: |
| 420 (87,1ost 1) | 403 (1 lost | 396 (87, lost 2 | 344 (85,1ost |
|  | @ age 1) | and skipped 1) | 2 @ age 1) |
|  | 425(87, |  |  |
|  | skipped 1, |  |  |
|  | $\begin{aligned} & \text { and lost } \\ & 1 \text { @ age 1) } \end{aligned}$ |  |  |

* Will be a complete interval when 2-year-olds are weaned in 1987
** Litter was first observed when composed of 1 -year-olds
SUMMARY:
AVERAGE REPRODUCTIVE INTERVAL
COMPLETE INTERVALS ONLY $(\mathrm{N}=17) \quad 3.35$
INCOMPLETE INTERVALS ONLY $(\mathrm{N}=5) \quad 5.4$
COMPLETE AND INCOMPLETE $(\mathrm{N}=22) \quad 3.82$

Table 23. Summary of age at first reproduction for Su-Hydro area brown bears by bear ID. Based on first observed Iitter, status in previous year is given in parentheses.

FIRST REPRODUCTION AT AGE:
4 YEARS 5 YEARS 6 YEARS 7 YEARS 8 YEARS

| 220** | 379* | 394* | 281非 | 341** | 407 (alone prev. 4 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 234** | 344* | 403* | 335\# |  | litter expt. in '87) |
| 240** | 244** | 261** | 381\# |  |  |
| 331*** | 204*** | 314** | 340\# |  |  |

Mean age based on long history $(N=5)=\quad 6.4$
Mean age based on backdated litters $(\mathrm{N}=13)=5.2$
Combined data $(\mathrm{N}=18)=\quad \therefore .5$

* Backdated based on lst observation with newborn litter.
** Backdated based on lst observation with litter of ylgs. *** Backdated based on lst observation with litter of 2-year-olds. \# Accurate value as no litter was observed in preceding 3 years.

Table 24. Brown bear harvest data in 3 GMU 13 study areas, 1962-85.

| Year | Core 1979 Area * |  |  | Greater 1979 Area ** |  |  | Su-Hydro Area *** |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | No. <br> MM | No. $\mathrm{FF}$ | No. Sex Unkwn. | No. <br> MM | No. FF | No. Sex Unkwn. | No. <br> MM | $\begin{aligned} & \text { No. } \\ & \text { FF } \end{aligned}$ | No. Sex Unkwn. |
|  | $\underline{1}$ |  | Unkwn. | 4 | 1 | - 0 | 5 | F | \% 0 |
| 62 | 1 | 0 | 0 | 4 | 5 | 0 | 8 | 8 | 0 |
| 63 | 2 | 2 | 0 | 2 | 5 | 0 | 6 | 7 | 0 |
| 64 | 3 | 2 | 0 | 1 | 2 | 0 | 8 | 9 | 0 |
| 65 | 3 | 3 | 0 | 5 | 3 | 1 | 6 | 7 | 0 |
| 66 | 2 | 1 | 1 | 2 |  | 0 |  |  | 0 |
| 67 | 1 | 3 | 0 | 2 | 4 | 0 | 6 | 5 | 0 |
| 68 | 0 | 4 | 0 | 3 | 5 | 0 | 6 | 4 | 0 |
| 69 | 1 | 1 | 0 | 5 | 1 | 0 | 5 | 0 | 0 |
| 70 | 1 | 0 | 0 | 3 | 1 | 1 | 3 | 3 | 0 |
| 71 | 4 | 0 | 0 | 5 | 2 | 0 | 5 | 11 | 0 |
| 72 | 3 | 1 | 0 | 4 | 3 | 0 | 8 | 1 | 0 |
| 73 | 1 | 0 | 0 | 5 | 1 | 0 | 6 | 2 | 0 |
| 74 | 4 | 2 | 0 | 5 | 7 | 0 | 5 | 6 | 0 |
| 75 | 7 | 5 | 1 | 12 | 10 | 1 | 3 | 8 | 2 |
| 76 | 1 | 4 | 1 | 4 | 6 | 2 | 8 | 12 | 1 |
| 77 | 4 | 0 | 0 | 6 | 1 | 0 | 5 | 1 | 0 |
| 78 | 7 | 2 | 0 | 8 | 5 | 0 | 10 | 4 | 1 |
| 79 | 7. | 3 | 0 | 10 | 3 | 0 | 7 | 6 | 0 |
| 80 | 4 | 3 | 1 | 8 | 7 | 2 | 9 | 4 | 2 |
| 81 | 7 | 1 | 0 | 7* | 3 | 0 | 13 | 9 | 0 |
| 82 | 2 | 7 | 0 | 8 | 12 | 0 | 15. | 6 | 0 |
| 83 | 3 | 3 | 0 | 11 | 5 | 0 | 12 | 13 | 1 |
| 84 | 6 | 5 | 0 | 14 | 11 | 2 | 15 | 14 | 1 |
| 85 | 7 | 3 | 0 | 14 | 6 | 0 | 19 | 19 | 1 |
| Total | 78 | 55 | 4 | $\overline{149}$ | $\overline{107}$ | 9 | 193 | 162 | 9 |

* Includes Uniform Coding Areas 2500-2900 and 3100-3200 in 13E, 0500-0800 in 13B, plus dump codes for: Susitna R. 13B unknown, Susitna R. (N. of Forks 13B), Nenana R. 13E unknown, Denali Hwy. unknown 13E, Susitna R. (Butte Ck. to the Forks 13), Susitna R. (N. of Forks 13).
** Includes Uniform Coding Areas 2500-2900 and 3100-3200 in 13E, 0300-1300 and 1600 in 13B, plus above-listed dump codes and: Denali Hwy. unknown 13B, Denali Hwy. unknown 13.
*** Includes Uniform Coding Areas 1300-1400 and 1600-2500 in 13E, 1500-1800 and 2100 in 13A, 0100-0200 in 13B, and 0200-0300 in 14B, plus dump codes for: Susitna R. 13A unknown, Susitna R. Jay Ck-Butte Ck. 13A, Tyone R./Ck. 13A unknown, Susitna R. 13E unknown, Talkeetna R. 13E unknown, Kosina Ck. 13E unknown, Kosina Ck. 13 unknown, Susitna R. (Jay Ck.-Butte Ck., 13), Talkeetna R. 13 unknown, Talkeetna R. Unit 14B unknown.
 fall season, Sp=shot in spring season).


Not included:
Subadults @2 in 1978, $=203,223$ (all ND).
Subadults @1 in $1978=232$ (ND)

* suspected mortality of 213 in 1979, not included as alive in 1979 or subsequently.
** recaptured $4 / 80$ and $6 / 85$ in Su-Hydro area.

Table 26. Status of brown bears first captured in 1979 (all were transplanted from upper Susitna drainage). (A-alive, $N R=n o$ return, $R=r e t u r n e d$, $N D=$ no data available, $F=s h o t$ in fall season, $S P=s h o t$ in spring season). Does not include transplanted bears first captured in 1978 (see Table 13). ND in year of capture indicated bear was not collared or soon shed its collar and no subsequent data were collected.

| Bear ID | Sex/age | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 246 | M/3 in '79 | Shot-F | - | - | - | - | - | - | - |
| 247 | M/8 in '79 | A | A | A | A | Shot-F | - | - | - |
| 243 | M/2 in /79 | A | A | Shot-F | - | - | - | - | - |
| 265 | M/4 in '79 | A | Shot-Sp | - | - | - | - | - | - |
| 268 | M/4 in '79 | A | Shot-Sp | - | - | - | - | - | - |
| 269 | F/18 in ${ }^{1} 79$ | A | A | Shot-F | - | - | - | - | - |
| 270 | F/l in '79 | A | Shot-F | - | - | - | - | - | - |
| 272 | M/9 in ' 79 | A | A | A | Shot-F | - | - | - | - |
| 260 | M/4 in '79 | A | A | A | A | Shot-F | - | - | - |
| 240 | F/5 in '79 | A, R | A | A | A | A | Shot-Sp | - | - |
| 273** | F/3 in '79 | A, R | A | A | A | A | A | A | A |
| 241 | M/3 in '79 | A, ND | ND | ND | ND | ND | ND | ND | ND |
| 249 | M/5 in '79 | $\mathrm{A}, \mathrm{ND}$ | ND | ND | ND | ND | ND | ND | ND |
| 258 | M/21 in '79 | A,ND | ND | ND | ND | ND | ND | ND | ND |
| 264 | F/4 in '79 | A, ND | ND | ND | ND | ND | ND | ND | ND |
| 267 | F/4 in ' 79 | A, ND | ND | ND | ND | ND | ND | ND | ND |
| 274 | F/11 in ${ }^{\text {c }} 79$ | A, ND | ND | ND | ND | ND | ND | ND | ND |
| 276 | M/4 in ' 79 | A, ND | ND | ND | ND | ND | ND | ND | ND |
| 236 | F/5 in '79 | A, R | ND | ND | ND | ND | ND | ND | ND |
| - 237 | M/10 in ' 79 | A, R | ND | ND | ND | ND | ND | ND | ND |
| 244 | F/6 in '79 | A, R | A | ND | ND | ND | ND | ND | ND |
| 251 | F/10 in '79 | A, R | A | ND | ND | ND | ND | ND | ND |
| 248 | F/4 in '79 | A, NR | ND | ND | ND | ND | ND | ND | ND |
| 261 | F/7 in ' 79 | A, NR | ND | ND | ND | ND | ND | ND | ND |

Max. no. Bears

| potentially alive <br> in year includes ND ( $M: F$ ) | 24 (12:12) | 23(11:12) | 20(9:11) | 18(8:10) | 17(7:10) | 14(4:10) | 13(4:9) | 13(4:9) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. marked bears <br> known shot in year (M:F) | 1(1:0) | 3(2:1) | 2(1;1) | 1(1:0) | 2(2:0) | 1(0:1) | 0 | 0 |
| Known of of potentially alive bears shot in year | 4\% | 13\% | 10\% | 68 | 12\% | 7\% | 0 | 0 |
| Cumulative \% (min.) of marked bears shot ( $\mathrm{N}=24$ ) | 4\% | 17\% | 25\% | 29\% | 38\% | 42\% | 42\% | 42\% |

[^4] in spring season). ND in year of capture indicates bear was not collared or soon shed its collar and no subsequent data were collected.

| Bear ID | Sex/age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 captures |  |  |  |  |  |  |  |  |
| 277 | F/10 in '80 | A | ND | ND | ND | ND | ND | ND |
| 279 | M/9 in 180 | A | A | A | A | Shot-F | - | - |
| 280 | M/5 in '80 | A | A | A | A | A | A | A |
| 281 | F/3 in ' 80 | A | A | A | A | A | A | A |
| 282 | M/4 in '80 | A | A | A | A | A | A | A |
| 283 | F/12 in ${ }^{1} 80$ | A | A | A | A | A | A | A |
| 284 | M/2 in 180 | A | Shot-Sp | - | - | - | - | - |
| 286 | M/3 in 180 | A | A | A | A | Shot-F | - | - |
| 292 | F/3 in ${ }^{180}$ | ND | ND | ND | ND | ND | ND | ND |
| 293 | M/3 in '80 | A | A | A | A | ND | Shot-Sp | - |
| 294 | M/10 10180 | A | Died in Aug. | - | - | - | - | - |
| 295 | M/12 in ${ }^{\prime} 80$ | ND | ND | ND | ND | ND | ND | ND |
| 299 | F/13 in ${ }^{1} 80$ | A | A | A | A | A | ND | ND |
| 297 | M/1 in 180 | A | Shot-F |  |  | A | - |  |
| 306 | F/3 in ${ }^{1} 80$ | ND | ND | ND | ND | ND | ND | ND |
| 308a | M/6 in 180 | A | A | A | Shot-F | - | - | - |
| 308b | F/5 in ' 80 | A | Died in Aug. |  | - | - | - | - |
| 309 | M/12 in ' 80 | A | A | A | A | A | A | ND |
| 311 | M/2 2 n ' 80 | Shot-F | - | - | - | - | - | - |
| 312 | $F / 10$ in ' 80 | A | A | A | A | Died-NS | - | - |
| 313 | F/9 in 180 | A | A | A | A | A | Shot-F | - |
| 314 | F/2 in ${ }^{\prime} 80$ | A | A | A | A | A | A | A |
| 315 | F/2 in ' 80 | A | A | A | A | A | A | Shot-Sp |
| 1981 captures |  |  |  |  |  |  |  |  |
| 331 | F/6 in 181 | - | A | Died in Aug. | - | - | - | - |
| 332 | M/2 in '81 | - | A | Shot-F | - | - | - | - |
| 333 | M/2 in ' 81 | - | Shot-F | - | - | - | - | - |
| 334 | F/10 in '81 | - | Lost in Sept.shot? | - | - | - | - | - |
| 335 | F/2 in '81 | - | A | A | A | A | A | A |
| 337 | F/13 in '81 | - | A | A | A | A | A | A |
| 339 | M/0 in ' 81 | - | Cub | Y1g | A | A | Shot-F | - |
| 340 | F/3 in '81 | - | A | A | A | A | A | A |
| 341 | F/6 in '81 | - | A | A | A | A | A | A |
| 342a | M/2 in '81 | - | A | A | A | Died-NS | - | - |
| 344 | F/5 in '81 | - | A | A | A | Lost Sept.shot? | - | - |
| 347 | M/14 in ${ }^{\text {' }} 81$ | - | A | A | A | A | A |  |
| 214*** | M/2 in ${ }^{\prime} 78$ | A | A | A | A | A | A | A |
| 273*** | F/3 in 179 | A | A | A | A | A | A | A |

Table 27. (cont'd)

Bear ID
Sex/age
1982
1983
1984
1985
1986
1982 captures


Table 27. (cont'd)


## Not Included:

Subadults @2 in1980: 285;
1983: 397 \& 398 both recaptured in 1985
Subadults @1 in 1980: 298;
1983: 383;
1984: 421, 418, 419

* G373 (M@9 in 1982) not included as it shed collar and had no ear tags or tattoo, so was not recogntzable as a marked bear subsequently.
** Downstream study area.
*** Captured earlier as part of studies outside of Su -Hydro area.

Table 28. Summary of Tables 25-27, marked hunter-killed brown bears in GMU 13.

|  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 (prelim.) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Maximum no. of marked bears potentially alive in year (Includes N.D.) (M:F) | 28(15:13) | 51(28:33) | 74 (40:34) | 77(37:39) | 70 (31:39) | 82(37:45) | 78(29:49) | 73(28:45) | 68(23:45) |
| No. marked bears shot in year* (M:F) | 1(0:1) | 2(2:0) | 6(5;1) | 7(5:2) | 5(3:2) | 8(6:2) | 11(7:4) | 5(2:3) | 6(3:3) |
| Min. \% of marked bears shot in year | 4\% | 4\% | 8\% | 9\% | 78 | 10\% | 14\% | 7\% | 9\% |
| \% males in population of marked bears | 54\% | 55\% | 54\% | 48\% | 44\% | 45\% | 37\% | 38\% | 34\% |
| \% males in harvest |  |  |  |  |  |  |  | 1978-1984 19 | 1978-1986 |
| - of marked bears | 0 | 100\% | 83\% | 71\% | 60\% | .75\% | 64\% | 70\% | $65 \%$ |

Table 29. Sumary of apparent natural mortalities of radio-collared adult bears. Susitna Hydro

| project. Includes black bears $\geq 1$ year of age and brown bears $\geq 2$ year of age. |
| :---: |

Bear ID \begin{tabular}{c}

Sex/age (at death), | reprod. status |
| :--- | <br>

\hline
\end{tabular}$\quad$ Conments

## Black bears

| B291 | M/3 | Died 2-28 July 1980, 2 months after capture, cause of death unknown. |
| :---: | :---: | :---: |
| B300 | M/7 | Died 6-14 May 1980, 2-10 days after capture, cause of death unknown but capture myopathy possible (M99/Rompun used, immobilization, and recovery were apparently normal). |
| B288 | F/10 with 3c | Not sure bear died but suspect that it did and collar was moved away from carcass by predator. Probably died 22-27 August 1980;' 6 months after capture. |
| B319 | M/4 | Died 29 July-4 August 1981, 11 months after capture, cause unknown. |
| B330 | M/1 | Died 17-24 August 1981, 5 months after capture in den with mother and sibling, apparently killed and eaten by predator. Radio-collared female sibling survived (B329). |
| B357 | M/4 | Died winter of 1981, 6 months after capture, apparently killed by another bear (species?) at or near its den and eaten. |
| B322 | M/6 | Died 24-29 June 1982, 4 weeks after recapture (was very skinny and weighed an est. 90 lbs.), cause unknown. |
| B327 | F/8 with 2c | Died 20 June-1 July 1983, 4 months after recapture in den, killed by predator (probably bear), but not eaten (cub defense?). |
| B379 | F/9 with 3c | Died early July 1983 (?), 3 months after recapture in den, canine punctures in scapula, in brown bear habitat, lost cubs earlier. Suspect was killed by brown bear. |
| B365 | M/6 | Died October 19839 months after recapture in den. Scavenged (killed?) by wolves. Guess may have been wounded by hunter (no evidence). Good condition. |
| B346 | M/12 | Died in May 1984, eaten by unknown predator-suspect a brown bear. |
| B343 | M/8 | Died in fall '84. Suspect may have been wounded by hunter, but have no evidence. |
| B358 | M/4 | Died summer '84, cause unknown, not disturbed. |

## Brown bear

| G331 | F/7 | Died l-31 July 1982, 14 months after capture, cause of death unknown, <br> had no cubs in 1982, but should have (weaned 2@2 in 1981). Bones not <br> scattered. We lghed 284 lbs. on 5/81 (large). |
| :--- | :--- | :--- |
| G389 G422 | M/2 | Died early October 1983. Cause undeternined. |
|  | M/7 | Died June 1985. Cause undetermined, but suspect injury from moose <br> or another bear. Bear moved suddenly miles from home range <br> and was found dead 2 weeks later. |

Table 30. Brown bear home range sizes. Code 99 in year or age column indicates lumping of all years. Area $1=$ upstream, area $2=$ downstream, sex $1=$ male, sex $2=$ female, code 1 for COY indicates bear had iitter of newborn cubs.


Table 30 . (cont'd)


Table 30. (cont'd)

| $\begin{aligned} & \overline{\mathrm{ID}} \\ & \text { No. } \end{aligned}$ | Area | Sex | Year | Age | $\begin{aligned} & \text { No. } \\ & \text { Pts. } \end{aligned}$ | $\begin{gathered} \text { Size } \\ \text { Sq. Km. } \end{gathered}$ | Period | Comments | COY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 299 | 1 | 2 | 82 | 15 | 21 | 191.3 | Apr-Oct | w/coy (lost 6/82) | 0 |
| 299 | 1 | 2 | 83 | 16 | 24 | 223.9 | Apr-Oct | w/coy, survived | 1 |
| 299 | 1 | 2 | 84 | 17 | 60 | 466.7 | Apr-Oct | w/ylgs,failed 4/85 | 0 |
| 299 | 1 | 2 | 99 | 99 | 141 | 949.4 | 1980-1985 |  | 0 |
| 312 | 1 | 2 | 80 | 10 | 13 | 157.0 | May-0ct | w/@2 | 0 |
| 312 | 1 | 2 | 81 | 11 | 24 | 181.7 | Apr-Oct | w/coy, survived | 1 |
| 312 | 1 | 2 | 82 | 12 | 20 | 251.6 | Apr-Oct | w/ylgs | 0 |
| 312 | 1 | 2 | 83 | 13 | 15 | 191.0 | Apr-Sept | w/@2, no den | 0 |
| 312 | 1 | 2 | 99 | 99 | 74 | 457.9 | 1980-85 | died 5/84 CM | 0 |
| 313 | 1 | 2 | 80 | 9 | 14 | 81.5 | May-Oct | w/1@2 (g314) | 0 |
| 313 | 1 | 2 | 81 | 10 | 25 | 210.9 | Apr-0ct | w/coy (lost 5/81) | 0 |
| 313 | 1 | 2 | 82 | 11 | 22 | 128.3 | Apr-Oct | w/coy, survived | 1 |
| 313 | 1 | 2 | 83 | 12 | 20 | 271.5 | Apr-Oct | w/ylg, survived | 0 |
| 313 | 1 | 2 | 84 | 13 | 60 | 187.7 | Apr-Sept | shot 9/84 | 0 |
| 313 | 1 | 2 | 99 | 99 | 141 | 455.0 | 1980-84 |  | 0 |
| 315 | 1 | 2 | 83 | 5 | 18 | 280.4 | May-Oct | 1 st @ 2 in 80 | 0 |
| 315 | 1 | 2 | 84 | 6 | 24 | 222.7 | May-Oct | No den, no cubs | 0 |
| 315 | 1 | 2 | 99 | 99 | 43 | 351.2 | 1983-84 | failed 10/84 | 0 |
| 331 | 1 | 2 | 81 | 6 | 24 | 1281.7 | May-Oct | w/@2, died 7/82 | 0 |
| 331 | 1 | 2 | 99 | 99 | 34 | 1280.7 | 1981-82 | Natural mort. 7/82 | 0 |
| 334 | 1 | 2 | 81 | 10 | 31 | 110.9 | May-Sept | w/@2, failed 9/81 | 0 |
| 334 | 1 | 2 | 99 | 10 | 31 | 110:9 | 1981 |  | 0 |
| 335 | 1 | 2 | 81 | 3 | 34 | 179.8 | May-Oct | alone | 0 |
| 335 | 1 | 2 | 82 | 4 | 20 | 131.2 | Apr-Oct |  | 0 |
| 335 | 1 | 2 | 83 | 5 | 19 | 183.3 | Apr-Oct |  | 0 |
| 335 | 1 | 2 | 84 | 6 | 36 | 123.8 | Apr-Oct | w/2@0 | 1 |
| 335 | 1 | 2 | 99 | 99 | 118 | 431.3 | 1982-85 | w/ylgs. in ' 85 | 0 |
| 337 | 1 | 2 | 81 | 13 | 19 | 269.6 | May-Oct | w/coy, survived | 1 |
| 337 | 1 | 2 | 82 | 14 | 20 | 356.3 | Apr-Oct | w/ylg, survived | 0 |
| 337 | 1 | 2 | 83 | 15 | 20 | 245.9 | Apr-Oct | w/@2 | 0 |
| 337 | 1 | 2 | 84 | 16 | 26 | 195.7 | Apr-Oct | w/coy, survived | 1 |
| 337 | 1 | 2 | 99 | 99 | 94 | 545.4 | 1981-85 |  | 0 |
| 340 | 1 | 2 | 81 | 3 | 39 | 613.3 | May-Oct | alone | 0 |
| 340 | 1 | 2 | 82 | 4 | 23 | 712.0 | Apr-Oct | alone | 0 |
| 340 | 1 | 2 | 83 | 5 | 18 | 538.7 | Apr-Oct | alone | 0 |

Table 30. (cont'd)

|  | $\begin{aligned} & \hline \text { ID } \\ & \text { No. } \end{aligned}$ | Area | Sex | Year | Age | No. Pts. | $\begin{aligned} & \text { Size } \\ & \text { Sq. Km. } \end{aligned}$ | Period | Cormments | COY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 340 | 1 | 2 | 84 | 6 | 60 | 168.9 | Apr-0ct | w/2@0, survived | 1 |
|  | 340 | 1 | 2 | 99 | 99 | 152 | 1040.0 | 1981-85 | w/2@1 thru 85 | 0 |
|  | 341 | 1 | 2 | 81 | 6 | 28 | 888.7 | May-0ct | alone | 0 |
|  | 341 | 1 | 2 | 99 | 99 | 44 | 903.9 | 1981-82,85 | recaptured in '85 | 0 |
|  | 344 | 1 | 2 | 81 | 5 | 21 | 270.4 | May-Oct | w/coy, survived | 1 |
|  | 344 | 1 | 2 | 82 | 6 | 22 | 400.9 | Apr-Oct | w/ylg(lost 7/82) | 0 |
|  | 344 | 1 | 2 | 83 | 7 | 18 | 287.0 | Apr-Oct | w/coy, survived | 1 |
|  | 344 | 1 | 2 | 84 | 8 | 13 | 246.9 | Apr-Sept | w/ylg(lost 5/84) | 0 |
|  | 344 | 1 | 2 | 99 | 99 | 74 | . 615.4 | 1981-1984 | missing 9/84 | 0 |
|  | 380 | 1 | 2 | 82 | 15 | 9 | 493.1 | Jun-Oct | w/ylg | 0 |
|  | 380 | 1 | 2 | 83 | 16 | 12 | 450.0 | Apr-Sept | Shot 9/83 | 0 |
|  | 380 | 1 | 2 | 99 | 99 | 21 | 548.6 | 1982-83 | shot 9/83 | 0 |
|  | 381 | 1 | 2 | 82 | 3 | 17 | 264.9 | Jun-0ct | alone | 0 |
| $\omega$ | 381 | 1 | 2 | 83 | 4 | 18 | 250.6 | Apr-Oct | alone | 0 |
| $u$ | 381 | 1 | 2 | 84 | 5 | 43 | 325.8 | Apr-0ct | alone | 0 |
|  | 381 | 1 | 2 | 99 | 99 | 84 | 489.5 | 1982-85 | coy survived ' 85 | 1 |
|  | 384 | 1 | 2 | 83 | 12 | 16 | 198.9 | May-0ct | w/@2 | 0 |
|  | 384 | 1 | 2 | 99 | 99 | 25 | 350.6 | 1983-84 | failed 6/84 w/coy | 0 |
|  | 385 | 1 | 2 | 83 | 2 | 16 | 253.3 | May-Oct | w/g337 | 0 |
|  | 385 | 1 | 2 | 84 | 3 | 19 | 196.8 | Apr-Oct | no den, failed 10/84 | 0 |
|  | 385 | 1 | 2 | 99 | 99 | 37 | 464.9 | 1983-85 | spotted in 85 | 0 |
|  | 388 | 1 | 2 | 83 | 14 | 16 | 146.1 | May-0ct | w/@2 | 0 |
|  | 388 | 1 | 2 | 84 | 15 | 47 | 329.6 | Apr-Oct | w/coy (lost 5/84) | 0 |
|  | 388 | 1 | 2 | 99 | 99 | 73 | 403.6 | 1983-85 | coy in ' 85 , survived | 0 |
|  | 393 | 1 | 2 | 83 | 2 | 14 | 155.7 | May-Sept | no den, lost 9/83 | 0 |
|  | 393 | 1 | 2 | 99 | 2 | 14 | 155.7 | 1983 only | w/g384 \& sibs | 0 |
|  | 394 | 1 | 2 | 83 | 6 | 20 | 201.0 | May -Oct | w/coy (lost 5/83) | 0 |
|  | 394 | 1 | 2 | 84 | 7 | 25 | 151.2 | Apr-Sept | shot 9/84 | 0 |
|  | 394 | 1 | 2 | 99 | 99 | 45 | 249.3 | 1983-84 | shot 9/84 | 0 |
|  | 395 | 1 | 2 | 83 | 3 | 11 | 457.6 | May-Aug | no den, shot 8/83 | 0 |
|  | 395 | 1 | 2 | 99 | 99 | 11 | 457.6 | 1983 only | no den, shot 8/83 | 0 |
|  | 396 | 1 | 2 | 83 | 13 | 16 | 253.6 | May-Oct | w/@2 | 0 |
|  | 396 | 1 | 2 | 84 | 14 | 23 | 252.9 | Apr-Oct | coy (lost 5/84) | 0 |
|  | 396 | 1 | 2 | 99 | 99 | 59 | 377.4 | 1983-84 |  |  |
|  | 420 | 1 | 2 | 84 | 19 | 61 | 737.9 | May-0ct | w/ylgs, survived | 0 |

[^5]Table 30. (cont'd)

|  | $\begin{aligned} & \overline{\mathrm{ID}} \\ & \text { No. } \end{aligned}$ | Area | Sex | Year | Age | No. Pts. | $\begin{aligned} & \text { Size } \\ & \text { Sq. Km. } \end{aligned}$ | Period | Comments | COY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 420 | 1 | 2 | 99 | 99 | 79 | 772.4 | 1984-85 | coy in ' 86 | 0 |
|  | 423 | 1 | 2 | 84 | 99 | 27 | 151.8 | May-0ct | coy, unaged adult | 1 |
|  | 423 | 1 | 2 | 99 | 99 | 43 | 288.3 | 1984-85 | had @2 in '86 | 0 |
|  | 425 | 1 | 2 | 84 | 8 | 39 | 234.0 | May-Oct | alone | 0 |
|  | 425 | 1 | 2 | 99 | 99 | 48 | 318.9 | 1984-85 | coy in '85 (survived) | 1 |
|  | 308b | 1 | 2 | 80 | 5 | 15 | 142.0 | May-Oct | alone | 0 |
|  | 308b | 1 | 2 | 81 | 6 | 14 | 110.1 | Apr-Aug | died 8/81 | 0 |
|  | 308b | 1 | 2 | 99 | 99 | 29 | 190.9 | 1980-81 | Died 8/81, CM | 0 |
|  | 379 | 2 | 2 | 82 | 5 | 19 | 226.7 | Jun-Oct | w/coy, survived | 1 |
|  | 379 | 2 | 2 | 83 | 6 | 20 | 72.7 | Apr-Oct | w/ylg., survived | 0 |
|  | 379 | 2 | 2 | 84 | 7 | 13 | 104.3 | Apr-Oct | alone, shot 9/85 | 0 |
|  | 379 | 2 | 2 | 99 | 99 | 59 | 520.6 | 1982-85 |  | 0 |
|  | 403 | 2 | 2 | 83 | 6 | 19 | 135.4 | May-Oct | w/coy (survived) | 1 |
| $\stackrel{\rightharpoonup}{\omega}$ | 403 | 2 | 2 | 84 | 7 | 18 | 338.2 | Apr-Oct | w/ylg(survived) | 0 |
| の | 403 | 2 | 2 | 99 | 99 | 43 | 507.5 | 1983-85 | w/coy in' 86 | 0 |
|  | 407 | 2 | 2 | 83 | 4 | 17 | 185.7 | May-Oct | alone, downstream | 0 |
|  | 407 | 2 | 2 | 84 | 5 | 19 | 195.3 | Apr-Oct | alone | 0 |
|  | 407 | 2 | 2 | 99 | 99 | 45 | 250.8 | 1983-85 | alone in '85 too | 0 |

Table 31. Mean brown bear home range size in the Su-Hydro study area by sex and reproductive status categories, 1980-1984.

| Category | No.Individuals | Number of Points |  |  | Home Range Size ( $\mathrm{km}^{2}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean | Max. | Min. | Mean | S.D. | Max. | Min. |

TOTAL HOME RANGE (Summation all years)

| A11 bears | 47 | 59.1 | 162 | 6 | 1021.6 | 1167.9 | 5923.5 | 87.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| All males | 17 | 47.3 | 115 | 13 | 1941.0 | '1541.5 | 5923.5 | 87.5 |
| All females | 30 | 65.8 | 162 | 6 | 500.6 | 275.8 | 1280.7 | 110.9 |
| ANNUAL HOME R | (all | lenda | ear) |  |  |  |  |  |
| All bears | 106 | 23.7 | 84 | 6 | 580.5 | 635.0 | 3129.5 | 72.7 |
| All males | 32 | 22.8 | 84 | 8 | 1271.7 | 755.0 | 3129.5 | 87.5 |
| All females | 74 | 24.0 | 61 | 6 | 281.6 | 194.5 | 1281.7 | 72.7 |
| $\begin{aligned} & \text { Females } 5.0+, \\ & \text { without coy } \end{aligned}$ | 48 | . 24.2 | 61 | 6 | 300.5 | 215.2 | 1281.7 | 72.7 |
| ```Females 5.0+, with coy``` | 13 | 25.8 | 60 | 18 | 189.0 | 62.6 | 94.3 | 287.0 |

[^6]Table 32. Brown bear predation rates, by bear ID based on intensive monitoring in spring in the Su-Hydro study area. Only kills made on a consecutive observation day are listed. Area $1=$ upstream, $2=$ downsteam, $3=178$ studies (Ballard et al. in prep). Sex $1=$ male, $2=$ female, Status $1=a l o n e$ or $w / @ 2,2=w / c o y, 3=w /(0 i$, based on status on is june. if another bear or wolves also on kill, each credited with 0.5 kills. Observation day $=$ a day in which at least 1 visual observation was made. Consecutive observation day sums all days, for periods of $>2$ consecutive days. Misc. kills Lnclude suspected and probable kills.

| $\begin{aligned} & \text { Bear } \\ & \text { ID } \end{aligned}$ | Area | Sex | Age | Year | Repro. status |  | Period | Missing perlod | No. moose calves | No. adiult moose | Unident. moose | No. adult caribou | ```Ungulate age/ species unk.``` | $\begin{aligned} & \text { Misc. } \\ & \text { kills } \end{aligned}$ | Total <br> kills | No. Kills/ 100 con ob_day | No. con ob days per kill |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 207 | 3 | 2 | 11 | 78 | 2 | 7 | 5/28-6/22 |  |  |  |  |  | 1 |  | 1 | 14.29 | 7.00 |
| 220 | 3 | 2 | 5 | 78 | 3 | 16 | 5/28-6/22 |  | 1 |  | 1 |  |  |  | 2 | 12.50 | 8.00 |
| 221 | 3 | 2 | 8 | 78 | 3 | 15 | 5/28-6/22 |  | 5 | 1 |  |  | . |  | 6 | 40.00 | 2.50 |
| 204 | 3 | 2 | 7 | 78 | 1 | 13 | 5/28-6/22 |  | 2 | 1 |  |  |  |  | 3 | 23.08 | 4.33 |
| 202 | 3 | 2 | 8 | 78 | 1 | 18 | 5/28-6/22 |  | 5 | 1 |  |  |  |  | 6 | 33.33 | 3.00 |
| 206 | 3 | 2 | 13 | 78 | 1 | 18 | 5/28-6/22 |  | 1.5 | 0.5 |  |  |  |  | 2 | 11.11 | 9.00 |
| 208 | 3 | 2 | 12 | 78 | 1 | 21 | 5/28-6/22 |  | 8 | 2 |  |  | 1 |  | 11 | 52.38 | 1.91 |
| 209 | 3 | 2 | 4 | 78 | 1 | 14 | 5/28-6/22 |  |  |  | 1 |  |  |  | 1 | 7.14 | 14.00 |
| 212 | 3 | 2 | 10 | 78 | 1 | 6 | 5/28-6/22 |  |  |  |  |  |  |  | 0 | 0.00 | - |
| 213 | 3 | 2 | 10 | 78 | 1 | 8 | 5/28-6/22 |  | 1 |  |  |  |  |  | 1 | 12.50 | 8.00 |
| 219 | 3 | 2 | 4 | 78 | 1 | 5 | 5/28-6/22 |  | . |  |  |  |  |  | 0 | 0.00 | - |
| 231 | 3 | 2 | 12 | 78 | 1 | 11 | 5/28-6/22 |  |  | . |  |  |  |  | 0 | 0.00 | - |
| 201 | 3 | 1 | 10 | 78 | 1 | 11 | 5/28-6/22 |  |  |  |  |  |  |  | 0 | 0.00 | - |
| -205 | 3 | 1 | 4 | 78 | 1 | 22 | 5/28-6/22 |  | 2.5 | 2.5 |  |  | 0.5 |  | 5.5 | 25.00 | 4.00 |
| - 11 | 3 | 1 | 4 | 78 | 1 | 6 | 5/28-6/22 |  |  | 0.5 |  |  |  |  | 0.5 | 8.33 | 12.00 |
| 817 | 3 | 1 | 3 | 78 | 1 | 11 | 5/28-6/22 |  | 1 | 1 | . |  | 1 |  | 3 | 27.27 | 3.67 |
| 222 | 3 | 1 | 11 | 78 | 1 | 9 | 5/28-6/22 |  | 0.5 | 0.5 |  |  | 0.5 |  | 1.5 | 16.67 | 6.00 |
| 225 | 3 | 1 | 4 | 78 | 1 | 16 | 5/28-6/22 |  |  | 2 |  |  | 1 |  | 3 | 18.75 | 5.33 |
| 227 | 3 | 1 | 9 | 78 | 1 | 5 | 5/28-6/22 |  |  |  |  | 1 |  |  | 1 | 20.00 | 5.00 |
| 281 | 1 | 2 | 8 | 81 | 1 | 8 | 5/21-6/22 |  |  |  |  |  |  |  | 0 | 0.00 | - |
| 340 | 1 | 2 | 3 | 81 | 1 | 15 | 5/21-6/22 |  | 3 |  |  |  |  | 1 | 4 | 26.67 | 3.75 |
| 334 | 1 | 2 | 18 | 81 | 1 | 18 | 5/22-6/22 |  |  |  |  |  |  |  | 0 | 0.00 | - |
| 341 | 1 | 2 | 5 | 81 | 1 | 5 | 5/21-6/22 |  |  |  |  |  |  |  | 0 | 0.00 | - |
| 355 | 1 | 2 | 10 | 81 | 1 | 10 | 5/22-6/22 |  |  |  |  |  | 1 |  | 1 | 10.00 | 10.00 |
| 340 | 1 | 2 | 6 | 84 | 2 | 28 | 5/28-7/1 |  | 0.5 | 2 |  |  |  |  | 2.5 | 8.93 | 11.20 |
| 299 | 1 | 2 | 17 | 84 | 3 | 22 | 5/28-7/1 |  | 2 |  |  | . |  |  | 2 | 9.09 | 11.00 |
| 420 | 1 | 2 | 19 | 84 | 3 | 18 | 5/31-7/1 |  | 4 |  |  |  |  | 1 | 5 | 27.78 | 3.60 |
| 281 | 1 | 2 | 7 | 84 | 1 | 17 | 5/28-7/1 |  | 1 |  |  |  |  | 1 | 2 | 11.76 | 8.50 |
| 283 | 1 | 2 | 16 | 84 | 1 | 19 | 5/28-7/1 |  |  | 1 |  |  |  |  | 1 | 5.26 | 19.00 |
| 313 | 1 | 2 | 13 | 84 | 1 | 23 | 5/28-7/1 |  | 6.5 |  |  |  |  |  | 6.5 | 28.26 | 3.54 |
| 381 | 1 | 2 | 5 | 84 | 1 | 11 | 5/28-7/1 | 6/11-6/23 | 1 |  |  |  |  |  | 1 | 9.09 | 11.00 |
| 388 | 1 | 2 | 15 | 84 | 1 | 13 | 5/28-7/1 |  |  |  |  |  |  |  | 0 | 0.00 | ERR |
| 425 | 1 | 2 | 8 | 84 | 1 | 6 | 6/1-7/1 | 6/9-6/15 |  | 0.5 |  |  |  |  | 0.5 | 8.33 | 12.00 |
| 279 | 1 | 1 | 13 | 84 | 1 | 12 | 5/28-6/12 |  | 0.5 |  |  |  |  |  | 0.5 | 4.17 | 24.00 |
| 280 | 1 | 1 | 9 | 84 | 1 | 11 | 5/28-7/1 | 6/i1-6/22 |  |  |  |  |  | 2 | 2 | 18.18 | 5.50 |
| 282 | 1 | 1 | 8 | 84 | 1 | 11 | 6/1-7/1 | 6/9-6/14 | 1 | 0.5 |  |  |  | 2 | 3.5 | 31.82 | 3.14 |
| 382 | 1 | 1 | 2 | 84 | 1 | 16 | 5/28-7/1 |  |  |  |  |  | 2 |  | 2 | 12.50 | 8.00 |
| 399 | 1 | 1 | 10 | 84 | 1 | 15 | 5/28-6/25 |  | 2 |  |  | : |  |  | 2 | 13.33 | 7.50 |
| 400 | 1 | 1 | 21 | 84 | 1 | 9 | 5/30-7/1 | 6/19-6/22 | 1 |  |  |  |  |  | 1 | 11.11 | 9.00 |
| 422 | 1 | 1 | 4 | 84 | 1 | 15 | 5/28-7/1 | 6/20-6/24 | 3 |  |  |  |  | 1 | 4 | 26.67 | 3.75 |

Table 32. (cont'd)

| SUMMARY |  | No. moose calves | No. adult moose | Unident. moose | $\begin{aligned} & \text { No. } \\ & \text { adult } \end{aligned}$ caribou | $\begin{gathered} \text { Ungulate } \\ \text { age/ } \\ \text { species } \\ \text { unk. } \end{gathered}$ | Misc. <br> kills | Total kills | No. Kilis/ 100 con ob_day | No. con ob days ${ }^{-}$ per kill |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TOTALS, all bears = No. of bear-years $=$ | $\begin{array}{r} 534 \\ 40 \end{array}$ | 53 | 16 | 2 | 1 | 8 | 8 | 88 | 16.48 | 6.07 |
| Totals, males only = | 169 | 11.5 | 7 | 0 | 1 | 5 | 5 | 29.5 | 17.46 | 5.73 |
| No. of bear-years $=$ | 14 |  |  |  |  |  |  |  |  |  |
| Totals, females only $=$ No. of bear-years = | $\begin{array}{r} 365 \\ 26 \end{array}$ | 41.5 | 9 | 2 | 0 | 3 | 3 | 58.5 | 16.03 | 6.24 |
| Totals, females status l= No. bear-years = | $\begin{array}{r} 259 \\ 20 \end{array}$ | 29 | 6 | 1 | 0 | 2 | 2 | 40 | 15.44 | 6.48 |
| Totals, females status $2=$ No. of bear-years = | $\begin{array}{r} 35 \\ 2 \end{array}$ | 0.5 | 2 | 0 | 0 | 1 | 0 | 3.5 | 10.00 | 10.00 |
| Gotals, females status $3=$ No. of bear-years = | $\begin{array}{r} 71 \\ 4 \end{array}$ | 12 | 1 | 1 | 0 | 0 | 1 | 15 | 21.13 | 4.73 |
| Totals, all bears area $1=$ No. of bear-years = | $\begin{array}{r} 302 \\ 21 \end{array}$ | 25.5 | 4 | 0 | 0 | 3 | 8 | 40.8 | 14.41 | 7.46 |
| Totals, males area $1=$ No. bear-years = | $\begin{array}{r} 89 \\ 7 \end{array}$ | 7.5 | 0.5 | 0 | 0 | 2 | 5 | 15 | 16.85 | 5.93 |
| Totals, females area $1=$ No. bear-years = | $\begin{array}{r} 213 \\ 14 \end{array}$ | 18 | 3.5 | 0 | 0 | 1 | 3 | 25.5 | 11.97 | 8.35 |
| Totals, females area l \& status l= No. bear-years = | $\begin{array}{r} 145 \\ 11 \end{array}$ | 11.5 | 1.5 | 0 | 0 | 1 | 2 | 16 | 11.03 | 9.06 |
| Totals, females area $1 \&$ status $2=$ No. of bear-years = | $\begin{array}{r} 28 \\ 1 \end{array}$ | 0.5 | 2 | 0 | 0 | 0 | 0 | 2.5 | 8.93 | 11.20 |
| Totals, females area $1 \&$ status $3=$ No. of bear-years = | $\begin{array}{r} 40 \\ 2 \end{array}$ | 6 | 0 | 0 | 0 | 0 | 1 | 7 | 17.50 | 5.71 |

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Table 32. (cont'd)

| SUMMARY | No. consec. obsv.- days | No. moose Calves | No, adult moose | Unident. moose | No. adult caribou | ```Ungulate age/ species unk.``` | Misc. kills | Total <br> kills | $\begin{aligned} & \text { No. } \\ & \text { Kilis/ } \\ & 100 \text { con_ } \\ & \text { ob_day } \end{aligned}$ | No. Con ob_days ${ }^{-}$ per $k i l l$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Totals, all bears area $3=$ No. of bear-years = | $\begin{array}{r} 232 \\ 40 \end{array}$ | 27.5 | 12 | 2 | 1 | 5 | 0 | 47.5 | 20.47 | 4.88 |
| Totals, males area $3=$ No. bear-years = | $\begin{array}{r} 80 \\ 7 \end{array}$ | 4 | 6.5 | 0 | 1 | 3 | 0 | 14.5 | 18.13 | 5.52 |
| Totals, females area $3=$ No. bear-years = | $\begin{array}{r} 152 \\ 12 \end{array}$ | 23.5 | 5.5 | 2 | 0 | 2 | 0 | 33 | 21.71 | 4.61 |
| Cotals, females area $3 \&$ status $1=$ No. bear-years $=$ | $\begin{array}{r} 114 \\ 9 \end{array}$ | 17.5 | 4.5 | 1 | 0 | . 1 | 0 | 24 | 21.05 | 4.75 |
| Totals, females area $3 \&$ status $2=$ No. of bear-years = | 7 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 14.29 | 7.00 |
| Totals, females area $3 \&$ status $3=$ No. of bear-years = | $\begin{array}{r} 31 \\ 2 \end{array}$ | 6 | 1 | 1 | 0 | 0 | 0 | 8 | 20.00 | 5.00 |

Table 32. (cont'd)

| SUMMARY | ```No. consec. obsv.- days``` | No. moose calves | No. adult moose | Unident. moose | No. adult caribou | ```Ungulate age/ species unk.``` | Misc. <br> kills | Total kills | $\begin{aligned} & \text { No. }_{\text {N }} \\ & \text { Kilis/ } \\ & 100 \text { con_ } \\ & \text { ob_day } \end{aligned}$ | No. con ob days per̄ kill |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Totals, in $1981=$ No. of bear-years = | $\begin{array}{r} 56 \\ 5 \end{array}$ | 3 | 0 | 0 | 0 | 1 | 1 | 5 | 8.93 | 11.20 |
| Totals, males in $1981=$ No. bear-years $=$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | - |
| Totals, females in 1981 = No. bear-years = | $\begin{array}{r} 56 \\ 5 \end{array}$ | 3 | 0 | 0 | 0 | ${ }^{1}$ | 1 | 5 | 8.93 | 11.20 |
| Totals, FF in ' $81 \mathrm{w} / \mathrm{status} 1$ No. bear-years = | $\begin{array}{r} 56 \\ 5 \end{array}$ | 3 | 0 | 0 | 0 | 1 | 1 | 5 | 8.93 | 11.20 |
| Totals, FF in ' $81 \mathrm{w} / \mathrm{status} 2$ <br> No. of bear-years $=$ $\Delta$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | - |
| Totals, $\mathbf{F F}$ in ' $81 \mathrm{w} /$ status 3 No. of bear-years $=$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | - |
| Totals, all bears in $1984=$ No. of bear-years = | $\begin{array}{r} 246 \\ 16 \end{array}$ | 22.5 | 4 | 0 | 0 | 2 | 7 | 35.5 | 14.43 | 6.93 |
| Totals, males in $1984=$ No. bear-years $=$ | $\begin{array}{r} 89 \\ 7 \end{array}$ | 7.5 | 0.5 | 0 | 0 | 2 | 5 | 15 | 16.85 | 5.93 |
| Totals, females in $1984=$ No. bear-years = | $\begin{array}{r} 157 \\ 9 \end{array}$ | 15 | 3.5 | 0 | 0 | 0 | 2 | 20.5 | 13.06 | 7.66 |
| Totals, FF in ' $84 \mathrm{w} /$ status 1 No. bear-years $=$ | $\begin{array}{r} 28 \\ 6 \end{array}$ | 8.5 | 1.5 | 0 | 0 | 0 | 1 | 11 | 39.29 | 2.55 |
| Totals, FF in ' $84 \mathrm{w} /$ status 2 No. of bear-years = | $\begin{array}{r} 28 \\ 1 \end{array}$ | 0.5 | 2 | 0 | 0 | 0 | 0 | 2.5 | 8.93 | 11.20 |
| Totals, FF in ' $84 \mathrm{w} / \mathrm{status} 3$ No. of bear-years = | $\begin{array}{r} 40 \\ 2 \end{array}$ | 6 | 0 | 0 | 0 | 0 | 1 | 7 | 17.50 | 5.71 |

Table 33. Results of intensive monitoring of brown bear predation rates during summer 1984. Bears were 1ocated once/day from 23 July through 1 Äugust, conditions permitting.

| Bear ID | Sex | Age | Repro. status | No. of locations | $\begin{aligned} & \text { No. of } \\ & \text { visuals (\%) } \end{aligned}$ | No. of locations at salmon streams | No. of visuals <br> at salmon streams | Total known or suspected kills of ungulates |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MALES |  |  |  |  |  |  |  |  |
| 282 | M | 8 | -- | 9 | 4 | 9 | 4 | 0 |
| 382 | M | 2 | -- | 5 | 1 | 0 | 0 | 0 |
| 280 | M | 9 | - | 4 | 1 | 0 | 0 | 0 |
| 399 | M | 10 | -- | 9 | 5 | 9 | 5 | 0 |
| 279 | M | 13 | -- | 6 | 3 | 6 | 3 | 0 |
| 400 | M | 21 | -- | 6 | 0 | 0 | 0 | 0 |
| 422 | M | A | -- | 6 | 5 | 0 | 0 | 1 |
| 官342 | M | 5 | -- | 5 | 1 | 5 | 1 | 0 |
| N | Sub | tals | males | 50 | 20(40.08) | 29 | 13 (44.8\%) | 1 |
| FEMALES |  |  |  |  |  |  |  |  |
| 381 | F | 5 | alone | 4 | 0 | 0 | 0 | 0 |
| 281 | F | 7 | alone | 6 | 0 | 0 | 0 | 0 |
| 313 | F | 13 | alone | 6 | 2 | 0 | 0 | 0 |
| 388 | F | 15 | alone | 4 | 1 | 0 | 0 | 0 |
| 283 | F | 16 | alone | 8 | 2 | 1 | 1 | 0 |
| 425 | F | A | alone | 6 | 2 | 0 | 0 | 0 |
| 315 | F | 6 | alone | 8 | 5 | 8 | 5 | 0 |
| 394 | F | 7 | alone | 8 | . 1 | 8 | 1 | 0 |
| 396 | F | 15 | alone | 6 | 2 | 5 | 1 | 0 |

Table 33. (cont'd)

| Bear ID | Sex | Age | Repro. status | $\begin{aligned} & \text { No. of } \\ & \text { locations (\%) } \end{aligned}$ | $\begin{aligned} & \text { No of } \\ & \text { vtsuals (\%) } \end{aligned}$ | No. of locations at salmon streams | No. of locations at salmon streams (\%) | Total known or suspected kills of ungulates |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 407 | F | 6 | alone | 6 | 5 | 6 | 5 | 0 |
| 344 \& 385 | F | -- | alone | 2 | 2 | 0 | 0 | 0 |
| 340 | F | 6 | w/2@0 | 6 | 6 | 0 | 0 | 0 |
| 423 | F | A | 2/3@0 | 9 | 7 | 7 | 5 | 0 |
| 335 | F | 6 | w/2@0 | 5 | 3 | 0 | 0 | 0 |
| 337 | F | 10 | w/2@0 | 2 | 2 | 0 | 0 | 0 |
| 299 | F | 18 | w/3@1 | 6 | 6 | 0 | 0 | 0 |
| 420 | F | A | w/2@1 | 9 | 5 | 9 | 5 | 0 |
|  | Subtotals for females |  |  | 101 | 51 (50.5\%) | 44 | 23 (52.3\%) | 0 |
| $\stackrel{\sim}{\omega}$ TOTALS FOR | ALL B |  |  | 161 | 71 (44.1\%) | 73 | 36 (49.3\%) | 1 |

* Note that if the same ratio of kills to visuals observed in the spring (48:475) were present in the summer, then 7.2 kills would have been observed during the. 71 visual observations made. Excluding the observations at salmon streams leaves only 35 visual observations and 3.5 kills would have been expected with this number of observations using the ratio of kills:visual observations observed in the spring.

Table 34. Brown bear predation rates by different sex and age categories. All data, 1978-1984, are included. Status $1=$ alone or with 2 yearmolds status $2=$ with cubs, and status $3=$ with yearlings. Area $1=$ Su-hydro studies and Area $3=$ work in 1978 based on 2 yearmolds status $2=$ with cubs, and status $3=$ with yearlings.
Spraker et al. (1981). Den site observations are not included.

| all bears | No. Visuals | No. w/o Visuals | Visuals | $\begin{aligned} & \text { No. } \\ & \text { moose } \\ & \text { calves } \end{aligned}$ | No. adult moose | $\begin{aligned} & \text { No. } \\ & \text { adit. } \\ & \text { caribou } \end{aligned}$ | $\begin{aligned} & \text { Age/ } \\ & \text { spec. } \\ & \text { Unknown } \end{aligned}$ | $\underset{\text { kill }}{\text { Probable }}$ | $\underset{\text { kill }}{\text { Suspected }}$ | $\begin{aligned} & \text { Total } \\ & \text { Kills } \end{aligned}$ | $\begin{gathered} \text { Kills/100 } \\ \text { visuals } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Thals, all bears = | 2188 | 852 | 72.0 | 68 | 42 | 9 | 26.5 | 10.5 | 12.5 | 168.5 | 7.70 |
| No. of bear-years = | 156 |  |  |  |  |  |  |  |  |  |  |
| Totals, males only $=$ | 582 | 269 | 68.4 | 17.5 | 15 | 0 | 8 | 5 | 5 | 50.5 | 8.68 |
| No. of bear-years = | 46 |  |  |  |  |  |  |  |  |  |  |
| Totals, females only $=$ | 1606 | 583 | 73.4 | 50.5 | 27 | 9 | 18.5 | 5.5 | 7.5 | 118 | 7.35 |
| No. of bear years = | 110 |  |  |  |  |  |  |  |  |  |  |
| Totals, females status $1=$ | 978 | 424 | 69.8 | 32 | 18 | 7 | 9.5 | 2.5 | 6.5 | 75.5 | 7.72 |
| No. bear-years = | 68 |  |  |  |  |  |  |  |  |  |  |
| Totals, females status $2=$ | 334 | 90 | 78.8 | 2.5 | 4 | 1 | 2 | 3 | 0 | 12.5 | 3.74 |
| No. of bear-years = | 23 |  |  |  |  |  |  |  |  |  |  |
| Totals, females status $3=$ | 294 | 69 | 81.0 | 16 | 5 | 1 | 7 | 0 | 1 | 30 | 10.20 |
| No. of bear-years = | 19 |  |  |  |  |  |  |  |  |  |  |

Table 34. (cont'd)

|  | SU HYDRO ONLY | $\begin{gathered} \text { No. } \\ \text { Visuals } \end{gathered}$ | No. w/o <br> Visuals | $\stackrel{\%}{\text { Visuals }}$ | moose calves | No. adult moose | $\begin{aligned} & \text { No. } \\ & \text { adit. } \\ & \text { caribou } \end{aligned}$ |  | $\begin{gathered} \text { Probable } \\ \text { kill } \end{gathered}$ | $\begin{aligned} & \text { Suspected } \\ & \text { kill } \end{aligned}$ | $\begin{aligned} & \text { Total } \\ & \text { Kills } \end{aligned}$ | $\begin{aligned} & \text { Kills/100 } \\ & \text { visuals } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Totals, all bears area 1 = No. of bear-years = | $\begin{array}{r} 1632 \\ 118 \end{array}$ | 691 | 70.3 | 40 | 18.5 | 6 | 17.5 | 6 | 8.5 | 96.5 | 5.91 |
|  | Totals, males area $1=$ No. bear-years = | $\begin{array}{r} 404 \\ 32 \end{array}$ | 218 | 65.0 | 11 | 3 | 0 | 5 | 3 | 3 | 25 | 6.19 |
|  | Totals, females area $1=$ No. bear-years $=$ | $\begin{array}{r} 1228 \\ 86 \end{array}$ | 473 | 72.2 | 29 | 15.5 | 6 | 12.5 | 3 | 5.5 | 71.5 | 5.82 |
|  | Totals, females area $1 \&$ status $1=$ No. bear-years = | $\begin{array}{r} 716 \\ 53 \end{array}$ | 383 | 65.2 | 17.5 | 9.5 | 5 | 6.5 | 0 | 4.5 | 43 | 6.01 |
|  | Totals, females area $1 \&$ status $2=$ No. of bear-years = | $\begin{array}{r} 289 \\ 19 \end{array}$ | 51 | 85.0 | 1.5 | 3 | 1 | 1 | 3 | 0 | 9.5 | 3.29 |
| $\stackrel{\text { - }}{\text { i }}$ | Totals, females area $1 \&$ status $3=$ No. of bear-years = | $\begin{array}{r} 223 \\ 14 \end{array}$ | 39 | 85.1 | 10 | 5 | 0 | 5 | 0 | 1 | 19 | 8.52 |

(cont inued on next page)

Table 34. (cont'd)

| 1978 ONLY | $\begin{aligned} & \text { No. } \\ & \text { Visuals } \end{aligned}$ | No. w/o Visuals | Visuals | No. moose calves | No. adult moose | No. adlt. caribou | Age/ spec. Unknown | $\underset{\text { kill }}{\text { Probable }}$ | Suspected $k 111$ | Total <br> Kills | $\begin{aligned} & \text { Kills/100 } \\ & \text { visuals } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Totals, all bears area $3=$ No. of bear-years = | $\begin{array}{r} 483 \\ 26 \end{array}$ | 67 | 87.8 | 28 | 23.5 | 3 | 9 | 4.5 | 4 | 72 | 14.91 |
| Totals, males area $3=$ No. bear-years = | $\begin{array}{r} 160 \\ 10 \end{array}$ | 23 | 87.4 | 6.5 | 12 | 0 | 3 | 2 | 2 | 25.5 | 15.94 |
| Totals, females area $3=$ No. bear-years = | $\begin{array}{r} 323 \\ 16 \end{array}$ | 44 | 88.0 | 21.5 | 11.5 | 3 | 6 | 2.5 | 2 | 46.5 | 14.40 |
| Totals, females area $3 \&$ status $1=$ No. bear-years = | $\begin{array}{r} 226 \\ 11 \end{array}$ | 25 | 90.0 | 14.5 | 8.5 | 2 | 3 | 2.5 | 2 | 32.5 | 14.38 |
| Totals, females area $3 \&$ status $2=$ No. of bear-years = | $\begin{array}{r} 32 \\ 2 \end{array}$ | 16 | 66.7 | 1 | 1 | 0 | 1 | 0 | 0 | 3 | 9.38 |
| Totals, females area 3 \& status $3=$ No. of bear-years = | $\begin{array}{r} 65 \\ 3 \end{array}$ | 3 | 95.6 | 6 | 2 | 1 | 2 | 0 | 0 | 11 | 16.92 |

Table 35. Den entrance and emergence dates of radio-collared brown bears for the winter of 1980-81 ("S" is the standard deviation, but it includes variability from the fluctuating time between observations, as well as variability in denning times).


Table 36. Den entrance and emergence dates of radio-collared brown bears for the winter of 1981-82 ("S"' is the standard deviation, but it includes variability from the fluctuating time between observations, as well as variability in denning times)


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Table 37. Den entrance and emergence dates of radio-collared brown bears for the winter of 1982-83 ("S" is the standard deviation, but it included variability from the fluctuating time between observations, as well as variability in denning times).

| Bear ID | Sex | Reproductive status at exit | 1982 Entrance |  |  | 1983 Emergence |  |  | Days in Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Min. | Max. | Mid. | Min. | Max. | Mid. | Min. | Max. | Mid. |
| 280 | M |  | 6 Oct | 15 Oct | 10 Oct | 17 Apr | 25 Apr | 21 Apr | 184 | 201 | 193 |
| 281 | F | $2 @ 0$ | 6 Oct | 20 Oct | 13 Oct | 14 May | 16 May | 15 May | 206 | 222 | 214 |
| 283 | F | $1 @ 0$ | 6 Oct | 15 Oct | 10 Oct | 14 May | 15 May | 15 May | 211 | 221 | 217 |
| 299 | F | $3 @ 0$ | 6 Oct | 15 Oct | 10 Oct | 23 May | 1 Jun | 28 May | 220 | 238 | 230 |
| 312 | F | $1 @ 2$ | 6 Oct | 20 Oct | 13 Oct | 25 Apr | 4 May | 30 Apr | 187 | 210 | 199 |
| 313 | F | $2 @ 1$ | 15 Oct | 20 Oct | 18 Oct | 14 May | 15 May | 15 May | 206 | 212 | 209 |
| 335 | F | w/o | 20 Sep | 6 Oct | 28 Sep | 17 Apr | 25 Apr | 21 Apr | 193 | 217 | 205 |
| 337 | F | $1 @ 2$ | 20 Oct | 15 Nov | 2 Nov | 10 May | 14 Máy | 12 May | 176 | 206 | 191 |
| 340 | F | w/o | 6 Oct | 15 Nov | 26 Oct | 25 Apr | 4 May | 30 Apr | 161 | 210 | 186 |
| 344 | F | 2@0 | 20 Oct | 15 Nov | 2 Nov | 14 May | 15 May | 15 May | 180 | 207 | 194 |
| 282 | M |  | 20 Oct | 15 nov | 2 Nov | 17 Apr | 25 Apr | 21 Apr | 153 | 187 | 170 |
| 379 | F | 2@1 | 20 Oct | 17 Nov | 4 Nov | 25 Apr | 4 May | 30 Apr | 159 | 196 | 177 |
| 381 | F | w/o | 6 Oct | 15 Oct | 10 Oct | 17 Apr | 25 Apr | 21 Apr | 184 | 201 | 193 |
| 380 | F | w/o | N. D. | N. D. | N. D. | 10 May | 19 May | 15 May | - | - | - |
| 342 | M |  | N. D. | N. D. | N. D. | 17 Apr | 25 Apr | 21 Apr | - | - | - |
|  | MEAN |  | $\overline{120 c t}$ | $\overline{280 c t}$ | $\overline{190 c t}$ | 1 May | 8 May | 5 May | $\overline{186}$ | $\overline{210}$ | $\overline{198}$ |
|  | "S" |  | 7 | 16 | 12 | 13 | 11 | 12 | 21 | 13 | 17 |
|  | n |  | 13 | 13 | 13 | 15 | 15 | 15 | 13 | 13 | 13 |

Table 38. Brown bear den entrance and emergence dates, winter of 1983/84.

|  |  | Reproductive status | 1983 Entrance |  |  | 1984 Emergence |  |  | Days in Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bear ID | Sex | at exit | earliest | latest | mid. | earliest | latest | mid. | Min. | Max. | MId. |
| G279 | M |  | 26 Sep | 24 Oct | 10 Oct | 3 Apr | 18 Apr | 11 Apr | 162 | 205 | 184 |
| G280 | M |  | 50 ct | 25 Oct | 15 Oct | 18 Apr | 30 Apr | 24 Apr | 176 | 208 | 192 |
| G281 | F | 2@0 | 26 Sep | 24 Oct | 10 Oct | 30 Apr | 10 May | 5 May | 189 | 227 | 208 |
| G282 | M |  | 5 Oct | 24 Oct | 15 Oct | 3 Apr | 7 Apr | 5 Apr | 162 | 215 | 189 |
| G283 | F | w/o | 26 Sep | 5 Oct | 1 Oct | 18 Apr | 10 May | 29 Apr | 196 | 227 | 212 |
| G293 | M |  | 27 Sep * | - | - | - | - | - | - | - | - |
| G299 | F | 301 | 27 Sep* | 24 Oct* | 11 Oct* | 8 Apr | 18 Apr | 13 Apr | 167 | 204 | 186 |
| 6313 | F | $1 @ 2$ | 5 Oct | 24 Oct | 150 ct | 30 Apr | 10 May | 5 May | 189 | 218 | 204 |
| G315 | F | w/o | - 26 Sep | 24 Oct | 10 Oct | 18 Apr | 30 Apr | 24 Apr | 177 | 217 | 197 |
| G335 | F | 2@0 | 15 Sep | 26 Sep | 6 Oct | 30 Apr | 10 May | 5 May | 217 | 238 | 228 |
| G337 | F | 2@0 | 5 Oct | 24 Oct | 15 Oct | 30 Apr | 10 May | 5 May | 189 | 218 | 204 |
| G340 | F | 2@0 | 5 Oct | 24 Oct | 150 ct | 10 May | 17 May | 14 May | 199 | 225 | 212 |
| G342 | M |  | 26 Sep* | 14 Nov* | 21 Oct* | 30 Apr | 10 May | 5 May | 168 | 227 | 197 |
| G344 | F | $1 @ 1$ | 27 Sep* | 14 Nov* | 25 Oct* | 30 Apr | 10 May | 5 May | 168 | 226 | 196 |
| G379 | F | $1 @ 2$ | 24 Oct | 14 Nov | 25 Oct | 3 Apr | 18 Apr | 11 Apr | 141 | 177 | 159 |
| G381 | F | w/o | 25 Oct* | -- | - | 18 Apr | 30 Apr | 24 Apr | - | 188 | - |
| G384 | F | 2 ¢0 | 5 Oct | 25 Oct | 15 Oct | 10 May | 28 May | 19 May | 198 | 236 | 217 |
| G385 | F | w/o | 26 Sep* | 24 Oct* | 10 Oct* | 30 Apr | 10 May | 5 May | 189 | 227 | 208 |
| G386 | M |  | 5 Oct | 24 Oct | 15 Oct | - | - | - |  | - | - |
| G388 | F | $2 @ 0$ | 26 Sep* | 15 Nov* | 21 Oct* | 30 Apr | 10 May | 5 May | 167 | 227 | 197 |
| G390 | M |  | 50 ct | 24 Oct | 150 ct | 30 Apr | 3 May | 1 May | 189 | 211 | 200 |
| G391 | M |  | 5 Oct | 24 Oct | 15 Oct | - | - | - | - | - | - |
| G393 | F | ? | 27 Sep* | , | , | - | - | - | - | - | - |
| G394 | F | w/o | 5 Oct | 24 Oct | 15 Oct | 30 Apr | 10 May | 5 May | 189 | 218 | 204 |
| G396 | F | $1 @ 0$ | 27 Sep * | 25 Oct* | 11 Oct* | 18 Apr | 30 Apr | 24 Apr | 176 | 216 | 196 |
| G399 | M |  | 50 ct | 25 Oct | 15 Oct | 18 Apr | 30 Apr | 24 Apr | 176 | 208 | 196 |
| G400 | M |  | 27 Sep* | 24 Oct | 11 Oct* | 18 Apr | 10 May | 24 Apr | 177 | 226 | 202 |
| G403 | F |  | 24 Oct | 14 Nov | 4 Nov |  |  |  | 141 | 177 | 159 |
| G407 | F | W/o | - | - | - | 18 Apr | 30 Apr | 24 Apr | -- | - | - |
| G423 | F | 4@0 | - | - | - | 16 May | 17 May | 17 May | -- | - | - |
|  | Mean |  | 30 OEt | 23 Oct | 150 ct | 23 Apr | 4 May | 29 Apr | 178 | 215 | 198 |
|  | "S" |  | 7.8 | 10.9 | 7.1 | 12.0 | 11.2 | 11.4 | 18.0 | 16.2 | 15.7 |
|  | n |  | 18 | 18 | 18 | 26 | 26 | 26 | 23 | 24 | 23 |

* Not included in calculation of means

Table 39. Brown bear den entrance and emergence dates, winter of 1984/85.

| Bear ID | Sex | Repro. status at exit | 1984 Entrance |  |  | 1985 Emergence |  |  | Days in Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | earliest | 1atest | Mid. | earliest | latest | Mid. | M1n. | Max. | Mid. |
| G280 | M |  | 11 Oct |  | (missing) | - | - | - |  |  |  |
| G281 | F | 200 | 11 Oct | 24 Oct | 18 Oct | 23 May | 1 June | 28 May | 211 | 233 | 222 |
| G282 | M |  | 7 Nov | 13 Nov | 10 Nov (unconfirmed) | 11 April | 18 April | 14 April | 149 | 162 | 156 |
| G283 | F | 200 | 11 Oct | 24 Oct | 18 Oct | 23 May | 1 June | 28 May | 211 | 233 | 222 |
| G299 | F | $3 @ 2$ ? | 10 ct | 11 Oct | 6 Oct | 18 April | 30 April | 24 April | 189 | 211 | 200 |
| G315 | F | ? | 11 Oct | 24 Oct | 18 0ct | is (missing) |  | - | - | - | - |
| G335 | F | 2 l | 11 Oct | 24 Oct | 18 Oct | 30 April | 9 May | 5 May | 188 | 210 | 199 |
| G337 | F | $2 @ 1$ | 11 Oct | 24 Oct | 18 Oct | 16 May | 23 May | 20 May | 204 | 224 | 214 |
| G340 | F | 201 | 11 Oct | 24 Oct | 18 Oct | 18 April | 30 April | 24 April | 176 | 201 | 189 |
| G379 | F | alone? | 1 Oct | 11 Oct | 6 Oct | 9 May | 16 May | 13 May | 210 | 227 | 219 |
| G381 | F | 2 @ | 11 Oct | 24 Oct | 18 Oct | 16 May | 23 May | 20 May | 204 | 224 | 214 |
| G388 | F | $2 @ 0$ | 11 Oct | 24 Oct | 18 Oct | 23 May | 1 June | 28 May | 211 | 233 | 222 |
| G396 | F | 2@o | 21 Sep | 11 Oct | 1 Oct (shed?) | 16 May | 23 May | 20 May | 217 | 244 | 231 |
| G399 | M |  | 11 Oct | 24 Oct | 18 Oct | 18 April | 30 April | 24 April | 176 | 201 | 189 |
| G400 | M |  | 11 Oct | 24 Oct | 18 Oct | 30 April | 9 May | 5 May | 188 | 210 | 199 |
| G403 | F | 1@2? | 7 Nov | 13 Nov | 10 Nov | 9 May | 16 May | 13 May | 177 | 190 | 199 |
| G382 | M |  | 11 Oct | 24 Oct | 18 Oct | 30 April | 9 May | 5 May | 188 | 210 | 199 |
| G407 | F | alone | 11 Oct | 24 Oct | 18 Oct | 18 April | 30 April | 24 April | 176 | 201 | 189 |
| G420 | F | $2 @ 2$ | 11 Oct | 24 Oct | 180 ct | 30 April | 9 May | 5 May | 188 | 210 | 199 |
| G422 | M |  | 11 Oct | 24 Oct | 18 Oct | 18 April | 30 April | 24 April | 176 | 201 | 189 |
| G423 | F | 3 l | 11 Oct | 240 ct | 18 Oct | 30 April | 9 May | 5 May | 188 | 210 | 199 |
| G425 | F | 2@0 | 11 Oct | 24 Oct | 18 Oct | 23 May | 2 June | 28 May | 211 | 233 | 222 |
|  | Mean |  | 11 Oct | 24 Oct | $\overline{180 c t}$ | 4 May | 13 May | 10 May | 192 | 213 | 204 |
|  | "S" |  | 9.7 | 8.1 | 9.0 | 14.2 | 13.3 | 13.8 | 17.6 | 18.9 | 17.5 |
|  | $n$ |  | 24 | 21 | 21 | 20 | 20 | 20 | 20 | 20 | 20 |




Table 40. (continued)


(continued on next page)

Table 40. (continued)


Table 40. (continued)

|  | $\begin{aligned} & \text { Deñ } \\ & \text { No. } \end{aligned}$ | $\begin{gathered} \text { Beă̈ } \\ \text { ID No. } \end{gathered}$ | Age at Exit | $\begin{gathered} \text { Elevation } \\ \text { (Feet) } \end{gathered}$ | $\begin{aligned} & \text { Slope } \\ & \text { (Degrees) } \end{aligned}$ | $\begin{aligned} & \text { Aspect } \\ & \text { (True N.) } \\ & \hline \end{aligned}$ | Vegetation | ENTRANCE |  | CHAMBER |  |  | Total Length (cm.) | $\begin{aligned} & \text { Previously } \\ & \text { iusea? } \\ & \text { (Yes/No) } \\ & \hline \end{aligned}$ |  | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  | $\begin{aligned} & \mathrm{Ht} .) \\ & \left(\mathrm{cm}_{\mathrm{o}}\right) \end{aligned}$ | $\begin{aligned} & \text { Wiath } \\ & (\mathrm{cm} .) \end{aligned}$ | $\begin{aligned} & \left.\mathrm{mn}_{-}\right) \\ & (\mathrm{cm}) \end{aligned}$ | $\begin{aligned} & \text { width } \\ & (\mathrm{cm} .) \end{aligned}$ | $\begin{aligned} & \mathrm{Ht} . \\ & \text { (cm.) } \end{aligned}$ |  |  |  |  |
| MALES |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 136 | G399 | 10 | -- | -- | -- | -- | - | - | - | - | $-$ | - | - | Not | located |
|  | 151 | G342 | 7 | -- | -- | -- | -- | - | - | - | - | - | - | - | Not | located |
|  | 176 | G282 | 9 | 3400** | 30** | 301** | Alders | - | - | - | - | - | - | - |  | visited |
|  | 197 | G399 | 11 | 4250** | -- | 125** | Tundra | - | - | - | - | - | - | - | Not | visited |

* Entered den with 2 yearlings, shed collar in den so exit not observed.
** Approximate value
Downstream

Dens No. 14, 16, 22, 24, 30, 31, 25, 28, 23, 5, 1, 15, 29, 17, 26 27 are 1980/1981


Dens No. $42,44,47,52,54,59,37,46,56,36,60,53,41,48$, 45 are 1981/1982

Dens No. 76, 78, 87, 89, 101, 102, 102; 103, 105, 107, 108, 109, 79, $106,111,94,86,110,77$ are 1982/1983

Dens No. 112, 117, 118, 119, 120, 121, 124, 125, 133, 134, 135, 153, $122,131,123,132,149,155,137,139,148,150,136,151$ are 1983/84

Dens No. 179, 194, 161, 164, 193, 162, 182, 192, 195, 163, 189, 166, 175, 165, 177, 196, 199, 170, 178, 183, 176, 197 are 1984/1985

Table 41. Brown bear den elevations by sex and reproductive status. Includes some bears of unknown sex and reproductive status in totals for all bears.

|  | Mean Elevation (feet) | N | Maximum | Minimum | Std. Dev. |
| :---: | :---: | :---: | :---: | :---: | :---: |
| UPSTREAM STUDY AREA |  |  |  |  |  |
| Females w/COY | 4221 | 29 | 5330 | 2010 | 695.3 |
| Females w/o COY | 4181 | 33 | 5240 | 2330 | 805.8 |
| Females w/COY or YLG | 4261 | 41 | 5330 | 2010 | 662.4 |
| Females w/YLG or @2 | 4465 | 19 | 5240 | 3350 | 541.1 |
| Single females | 3879 | 13 | 5150 | 2330 | 939.7 |
| All females | 4200 | 62 | 5330 | 2010 | 750.3 |
| All males | 3674 | 12 | 4950 | 2650 | 652.7 |
| All bears | 4128 | 80 | 5330 | 2010 | 738.6 |
| DOWNSTREAM STUDY AREA |  |  |  |  |  |
| All bears | 2100 | 10 | 3900 | 1050 | 817.2 |

Table 42. Distances between den sites (miles) used in different years by radio-collared brown bears. Based on principal winter den, early spring dens not considered.

| $\begin{aligned} & \text { Bear } \\ & \text { ID } \\ & \hline \end{aligned}$ | Age | $\begin{array}{r} 80 / 81 \\ \text { to } \\ 81 / 82 \\ \hline \end{array}$ | $\begin{array}{r} 80 / 81 \\ \text { to } \\ 82 / 83 \\ \hline \end{array}$ | $\begin{array}{r} 80 / 81 \\ \text { to } \\ 83 / 84 \\ \hline \end{array}$ | $\begin{gathered} 81 / 82 \\ \text { top } \\ 82 / 83 \\ \hline \end{gathered}$ | $\begin{gathered} 81 / 82 \\ \text { to } \\ 83 / 84 \\ \hline \end{gathered}$ | $\begin{gathered} 82 / 83 \\ \text { to } \\ 83 / 84 \\ \hline \end{gathered}$ | $\begin{array}{r} 80 / 81 \\ \text { to } \\ 84 / 85 \\ \hline \end{array}$ | $\begin{aligned} & 81 / 82 \\ & \text { to } \\ & 84 / 85 \\ & \hline \end{aligned}$ | $\begin{array}{r} 82 / 83 \\ \text { to } \\ 84 / 85 \\ \hline \end{array}$ | $\begin{gathered} 83 / 84 \\ \text { to } \\ 84 / 85 \\ \hline \end{gathered}$ | $\overline{\mathrm{x}}$ | $s$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FEMALES |  |  |  |  |  |  |  |  |  |  |  |  |  |
| G283 | 13 in' 81 | 3.2 | 2.4 | 1.6 | 5.3 | 4.9 | 1.7 | 3.4 | 3.5 | 5.8 | 4.4 | 3.6 | 1.5 |
| G313 | 10 in' 81 | 4.1 | 4.4 | 3.4 | 6.7 | 1.0 | 5.7 | - | - | - | - | 4.2 | 2.0 |
| G337 | 13 in' 81 | 3.3 | 2.4 | 1.9 | 3.7 | 3.1 | 0.6 | 4.2 | 1.0 | 4.7 | 4.1 | 2.9 | 1.4 |
| . 6344 | 5 1n' 81 | 3.1 | 1.5 | 3.8 | 1.6 | 1.2 | 2.5 | - | - | - | - | 2.3 | 1.0 |
| G299 | 14 in' 81 | 8.9 | 6.7 | 7.1 | 3.5 | 3.5 | 0.5 | 11.3 | 2.7 | 6.2 | 6.1 | 5.7 | 3.2 |
| G281 | 4 in' 81 | 1.9 | 1.7 | 1.7 | 0.2 | 0.2 | 0.1 | 2.7 | 1.5 | 1.6 | 1.5 | 1.3 | 0.9 |
| G335 | 4 in' 82 | - | - | - | 2.4 | 2.0 | 0.9 | - | 1.4 | 1.5 | 1.9 | 1.7 | 0.5 |
| G340 | 4 in' 82 | - | - | - | 0.3 | 17.7 | 17.6 | - | 18.1 | 18.0 | 0.6 | 12.0 | 9.0 |
| G312 | 11 in' 81 | 2.1 | 0.6 | - | 1.6 | - | - | - | - | - | - | 1.4 | 0.8 |
| G379 | 6 in' 83 | - | - | - | - | - | 5.3 | - | - | 5.3 | 0.5 | 3.7 | 2.8 |
| G315 | 2 in' 80 |  |  |  |  |  |  | - | - | - | 0.8 | - | - |
| G381 | 3 in' 82 |  |  |  |  |  |  | - | - | 2.8 | 2.5 | 2.7 | - |
| G388 | 14 in' 83 |  |  |  |  |  |  | - | - | - | 0.8 | - | - |
| G396 | 9 in'83 |  |  |  |  |  |  | - | - | - | 9.0 | - | - |
| 6403 | 4 in' 83 |  |  |  |  |  |  | - | - | - | 2.2 | - | - |
| G407 | 4 in' 83 |  |  |  |  |  |  | - | - | - | 5.1 | - | - |
| ( FEMA | LES) $\begin{aligned} & \overline{\mathrm{x}}= \\ & \mathrm{s}=\end{aligned}$ | $\begin{aligned} & 3.9 \\ & 2.3 \end{aligned}$ | $\begin{aligned} & 2.8 \\ & 2.1 \end{aligned}$ | $\begin{aligned} & 3.3 \\ & 2.1 \end{aligned}$ | $\begin{aligned} & 2.7 \\ & 2.3 \end{aligned}$ | 4.2 | 3.9 | 5.4 4.0 | 4.7 6.6 | 5.7 | $\begin{aligned} & 3.0 \\ & 2.6 \end{aligned}$ |  | $\begin{array}{r} 3.8 \\ 4.0 \\ 18.1 \end{array}$ |

Table 42. (cont'd)


Note: For G341, distance between dens, $81 / 82$ to $85 / 86$, is 2.1 miles (not included in above calculations).

T'able 43. Number of observations and percent (in parentheses) of radio-marked black bears within nested impoundment proximity zones of the Watana Impoundment (den-related activities are not included).

ZONE 1 ZONE 2 ZONE 3 ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) (1-5 miles) (over 5 miles) TOTAL


Value of Chi-Square test of the null hypothesis that the use of each zone is equivalent to expected values based on the area of each zone for:

|  | ZONE 1 | ZONE 2 | ZONE 3 | $\mathrm{X}^{2}$ | d.f. |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | obs. E(x) | obs. E(x) | obs. E (x) |  |  |
| All months, 3 zones | 551119.6 | 469245.6 | 271926.0 | 2,222** | 2 |
| All months, zones $1 \& 2$ only | $551334.1$ | 469685.9 | -- -- | 210** | 1 |

* Reject null hypothesis, $p$ less than 0.10 .
** Reject null hypothesis, $p$ less than 0.05 .

Table 44. Number of observations and percent (in parentheses) of radio-marked black bears within nested impoundment proximity zones of the Devil's Canyon Impoundment (den-related activities are not included).

ZONE 1 ZONE 2 ZONE 3 ZONE 4
TIME PERIOD (impoundment) (shore-1 mile) ( $1-5$ miles) (over 5 miles) TOTAL

| 1. | April 1-30 | 0 | 1 | 0 | 0 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | May 1-15 | 2 | 33 | 16 | 2 | 53 |
| 3. | May 16-31 | 2 | 43 | 43 | 0 | 88 |
| 4. | June 1-15 | 8 | 70 | 86 | 0 | 164 |
| 5. | June 16-30 | 3 | 45 | 75 | 2 | 125 |
| 6. | July 1-15 | 0 | 21 | 29 | 1 | 51 |
| 7. | July 16-31 | 0 | 13 | 33 | 1 | 47 |
|  | August 1-15 | 0 | 17 | 17 | 2 | 36 |
| 9. | August 16-31 | 2 | 18 | 26 | 2 | 48 |
| 10. | Sept. 1-15 | 1 | 13 | 13 | 3 | 30 |
| 11. | Sept. 16March 31 TOTALS | $\frac{\because_{0}}{18(3)}$ | $\frac{18}{292(43)}$ | $\frac{16}{354(52)}$ | $\frac{2}{15(2)}$ | 36 679 |
| Area | within zone $\left(\mathrm{km}^{2}\right)$ | 28.92 | 164.78 | 689.01 | -- | 882.71 |
|  | \% | 3.28 | 18.67 | 78.06 | -- | 100.0 |

Value of Chi-Square test of the null hypothesis that the use of each zone is equivalent to expected values based on the area of each zone for:

|  | ZONE 1 |  | ZONE 2 |  | ZONE 3 |  | $\mathrm{X}^{2}$ | d.f. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | obs | E(x) | obs, | E(x) | obs | E(x) |  |  |
| All months, 3 zones | 18 | 21.8 | 292 | 124.0 | 354 | 518.3 | 275** | 2 |
| $\begin{aligned} & \text { May l-June } 30 \\ & 3 \text { zones } \end{aligned}$ | 12 | 9.9 | 146 | 56.6 |  | 236.5 | 177** | 2 |
| $\begin{aligned} & \text { May l-June } 30 \\ & 2 \text { zones } \end{aligned}$ | 12 | 23.6 |  | 134.4 | -- | -- | 6.7** | 1 |

[^7]Table 45. Numbers of point locations in each of 4 impoundment proximity zones for individual black bears for each impoundment and for both impoundments lumped. Subadult dispersers and den site locations are not included.

| BLACK BEARS-WATANA IMPOUNDMENT ONLY |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bear ID | Sex | Zone 1 | Zone 2 | Zone 3 | Zone 4 | Totals |
| 287 | M |  | 2 | 11 |  | 13 |
| 302 | M | 46 | 27 | 19 |  | 92 |
| 303 | M | 1 | 4 | 13 |  | 18 |
| 304 | M |  |  | 18 | 1 | 19 |
| 305 | M |  |  | 6 | 3 | 9 |
| 322 | M | 7 | 3 | 1 |  | 11 |
| 323 | M | 4 | 10 | 14 |  | 28 |
| 324 | M | 3 | 9 | 5 |  | 17 |
| 346 | M | 6 | 33 | 9 |  | 48 |
| 348 | M |  | 2 | 2 |  | 4 |
| 357 | M | 12 | 5 |  | 17 |  |
| 359 | M | 36 | 35 | 14 |  | 85 |
| 360 | M | 11 | 22 | 3 |  | 36 |
| 387 | M | 17 | 19 | 30 |  | 66 |
| 401 | M | 19 | 4 | 2 |  | 25 |
| 416 | M | 3 | 4 | 6 | 4 | 17 |
| 342 B | M | 13 | 14 | 10 | 3 | 40 |
| All Males |  | 178 | 193 | 163 | 11 | 545 |
| \% |  | 32.7 | 35.4 | 29.9 | 2.0 |  |
|  | Sex | Zone 1 | Zone 2 | Zone 3 | Zone 4 | Totals |
| 289 | F | 36 | 9 | 16 |  | 61 |
| 301 | F | 16 | 39 |  | 55 |  |
| 317 | F | . | 6 | 4 |  | 10 |
| 318 | F | 3 | 1 | 3 | 3 | 10 |
| 321 | F |  | 3 | 2 |  | 5 |
| 327 | F | 39 | 11 | 2 |  | 52 |
| 328 | F |  | 5 | 21 |  | 26 |
| 329 | F | 90 | 15 | 1 |  | 106 |
| 349 | F | 22 | 35 | 28 |  | 85 |
| 354 | F | 12 | 35 | 3 |  | 50 |
| 358 | F | 34 | 33 | 6 |  | 73 |
| 361 | F | 65 | 21 | 2 |  | 88 |
| 363 | F | 16 | 35 | 3 |  | 54 |
| 364 | F | 37 | 22 | 11 |  | 70 |
| Watana |  |  |  |  |  |  |
| All Females |  | 370 | 270 | 102 | 3 | 745 |
| \% |  | 49.7 | 36.2 | 13.7 | 0.4 | 100 |
| Watana |  |  |  |  |  |  |
| ALL BEARS |  | 548 | 463 | 265 | 14 | 1290 |
| \% |  | 42.5 | 35.9 | 20.5 | 1.1 | 100 |

Table 45. (cont'd)

| BLACK BEARS-DEVILS CANYON IMPOUNDMENT ONLY |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bear ID | Sex | Zone 1 | Zone 2 | Zone 3 | Zone 4 | Totals |
| 287 | M | 1 | 16 | 14 | 1 | 32 |
| 303 | M | 11 | 29 |  |  | 40 |
| 304 | M | 4 | 12 |  |  | 16 |
| 319 | M | 8 | 6 |  |  | 14 |
| 324 | M | 23 | 19 | 7 |  | 49 |
| 348 | M | 4 | 5 |  |  | 9 |
| 359 | M | 2 | 4 |  |  | 6 |
| 401 | M | 4 | 31 | 11 |  | 46 |
| 416 | M | 2 | 11 | 22 | 3 | 38 |
| A11 Males$\%$ |  | 7 | 110 | 122 | 11 | 250 |
|  |  | 2.8 | 44.0 | 48.8 | 4.4 |  |


|  |  | Zone 1 | Zone 2 | Zone 3 | Zone 4 | Totals |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
| 288 | F |  | 12 | 4 |  | 16 |
| 289 | F |  | 27 | 35 | 62 |  |
| 290 | $\mathbf{F}$ | 2 | 14 | 13 |  | 29 |
| 317 | F | 2 | 42 | 51 |  | 95 |
| 318 | $\mathbf{F}$ |  | 16 | 19 | 35 |  |
| 321 | $\mathbf{F}$ | 3 | 29 | 29 | 61 |  |
| 325 | $\mathbf{F}$ | 1 | 2 | 6 | 9 |  |
| 327 | $\mathbf{F}$ |  | 6 | 5 | 11 |  |
| 328 | $\mathbf{F}$ | 2 | 10 | 38 | 50 |  |
| 329 | F |  | 1 | 1 | 2 |  |


| Devils Canyon |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| All Females | 10 | 159 | 201 | 0 | 370 |
| \% | 2.7 | 43.0 | 54.3 | 0.0 | 100 |
| Devils Canyon |  |  |  |  |  |
| ALL BEARS | 17 | 269 | 323 | 11 | 620 |
| \% | 2.7 | 43.4 | 52.1 | 1.8 | 100 |
| Both impoundments |  |  |  |  |  |
| All Males | 185 | 303 | 285 | 22 | 795 |
| \% | 23.3 | 38.1 | 35.8 | 2.8 | 100 |


| Both fmpoundments |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | ---: |
| All Females | 380 | 429 | 303 | 3 | 1115 |
| $\%$ | 34.1 | 38.5 | 27.2 | 0.3 | 100 |
|  |  |  |  |  |  |
| Both impoundments |  |  |  |  |  |
| ALL BEARS | 565 | 732 | 588 | 25 | 1910 |
| $\%$ | 29.6 | 38.3 | 30.8 | 1.3 | 100 |

Table 46. Number of Susitna River crossings by radio-marked black bears, 1980-1984.

(continued)

Table 46. (continued)


Table 46. (cont inued)

** possible unreported hunter kill, collar fallure, or emigration.
Reprod. status: * = cub of year $\quad y=y r l g$.

Table 47. Scat analyses of brown bear and black bear scats collected in the Su-Hydro study area, 1980-1982. (Analyses done by Paul Smith, ADF\&G, Soldotna). Values are \% volume ( $T=$ trace, $2=6-25 \%, 3=26-50 \%$, $4=51-75 \%$, $5=76-100 \%$ ).

 Values are $\%$ volume ( $\mathrm{T}=$ trace, $2=5-25 \%$, $3=26-50 \%$, $4=51-75 \%$, $5=75-100 \%$ ).

|  | Date Collected | Species of bear | Place | $\begin{gathered} \text { Sample } \\ \text { No. } \end{gathered}$ | Comments | 1 | 2 | 3 | 4 | 5 | 7 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer - F | 11-Sloughs |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/18/82 | ? | upstm | 25 | Steigers-84 |  |  |  |  |  |  |  |  |  | 2 |  |  |  |  |  |  | 5 |
|  | 8/20/83 | ? | upstm | 27 | Steigers-84 |  | 2 |  |  |  | 3 | 3 |  |  |  |  |  |  |  |  |  |  |
|  | 8/25/83 | ? | dstm | 5 | Slough 8A |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  | 2 |  |  |
|  | 8/25/83 | ? | dstm | 7 | Slough 8A |  |  | T |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/25/83 | ? | dstm | 8 | Slough 8A |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/25/83 | ? | dstm | 28 | Slough 8A | T |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  | 2 |  |
|  | 8/25/83 | ? | dstm | 31 | Slough 8A |  |  |  |  | 4 |  | 2 |  |  |  |  |  |  |  | T | T |  |
|  | 8/24/83 | ? | dstm | 13 | Slough 8B |  |  | T |  | 5 |  |  |  |  |  | T |  |  |  | T |  |  |
|  | 8/24/83 | ? | dstm | 4 | Slough 8B |  |  |  |  | 5 |  |  |  |  |  |  |  |  | T | T | T |  |
|  | 8/24/83 | ? | dstm | 21 | Slough 8B |  |  | T |  | 5 |  |  |  |  |  |  |  |  |  | T |  |  |
|  | 8/24/83 | ? | dstm | 17 | Slough 8B |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  | T | T |  |
|  | 8/24/83 | ? | dstm | 30 | Slough 8B | T |  | T |  | 4 |  | T |  |  |  |  |  |  |  | T | T |  |
|  | 8/24/83 | ? | dstm | 6 | Slough 8B | T |  |  |  | 4 |  | T |  |  |  |  |  |  |  |  | 2 |  |
|  | 8/24/83 | ? | dstm | 18 | Slough 8B |  |  |  |  | 3 |  |  |  |  |  |  |  |  |  | T | 2 |  |
|  | 8/24/83 | ? | dstm | 9 | Slough 8B |  |  |  |  | 3 |  |  |  |  |  |  |  |  |  | 3 | T |  |
|  | 8/24/83 | ? | dstm | 15 | 8B + nematode | 3 |  | 3 |  | 3 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/25/83 | ? | dstm | 14 | Slough 8A |  |  |  |  | 4 |  |  |  |  |  |  |  |  | T | T | T |  |
|  | 8/25/83 | ? | dstm | 22 | Slough 8A | T |  | 2 |  | 2 |  | 5 |  |  |  |  |  | T |  |  | 2 |  |
|  | 8/25/83 | ? | dstm | 3 | Slough 11 |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/26/83 | BRB? | dstm | 43 | Slough 20 |  |  |  |  | 3 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/26/83 | BRB? | dstm | 33 | Slough 21 |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  | T |  |
|  | 8/26/83 | BRB? | dstm | 29 | Slough 21. |  |  | T |  | 5 |  | T |  |  |  |  |  |  |  | T |  |  |
|  | 8/26/83 | BRB? | dstm | 26 | Slough 21 |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
| $\stackrel{\square}{\square}$ | 8/26/83 | ? | dstm | 24 | Slough 21 | 3 |  |  |  | 3 |  |  |  |  |  |  |  |  | 2 |  | T |  |
| $\stackrel{\infty}{\infty}$ | 8/26/83 | ? | dstm | 16 | McKenzie Ck. |  |  |  |  | 5 |  |  |  |  | T |  |  | T |  | T | T |  |
|  | 8/25/83 | ? | dstm | 19 | Moose Ck. | 2 |  |  |  | 5 |  |  |  |  |  |  |  |  | T | T | T |  |
|  | 8/25/83 | ? | dstm | 27 | Moose Ck. |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  | T |  |
|  | 8/25/83 | ? | dstm | 11 | Moose Ck. |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/24/83 | ? | dstm | 12 | Slough 8 | T |  | T |  | 5 |  |  |  |  |  |  |  |  |  |  | T |  |
|  | 8/25/83 | ? | dstm | 23 | Slough 8A |  |  | T |  | 5 |  |  |  |  |  |  |  | T |  |  |  |  |
|  | 8/25/83 | ? | dstm | 20 | Slough 8A |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/25/83 | ? | dstm | 25 | Slough A' | T |  | 3 |  | 3 |  | T |  |  |  |  |  |  | T |  | T |  |
|  | 8/18/83 | ? | upstm | 42 | Berry Plot \#1 |  | 3 |  |  |  |  |  |  | T |  | T |  |  |  | 2 |  |  |
|  | 8/18/83 | ? | upstm | 44 | Berry Plot \#2 |  | 3 |  |  |  | 3 |  |  | T |  |  | T | T |  | T |  |  |
|  | 8/18/83 | ? | upstm | 45 | Berry Plot \#1 |  | T |  |  |  | 3 | $T$ | T |  |  |  |  |  | 3 |  |  |  |
|  | 8/18/83 | ? | upstm | 46 | Berry Plot \#1 | 2 | 3 |  |  |  |  | 2 |  |  |  |  |  |  |  |  |  |  |
|  | 9/16/83 | ? | upstm | 22 | Steigers-84 |  | 2 |  | 2 |  | 4 |  |  |  |  |  |  |  |  |  |  |  |
|  | Spring Sam |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 5/19/83 |  | upstm |  | Steigers-84 |  |  |  | 5 |  |  |  |  |  | T |  |  |  |  |  |  |  |
|  | 5/19/83 | BKB | upstim | 36 | B404 | 2 |  |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |
|  | 5/31/83 | ? | upstm | 24 | Steigers-84 | 3 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 5/19/83 | ? | upstm | 26 | Steigers-84 |  |  |  |  |  |  |  |  | 5 |  |  |  |  |  |  |  |  |
|  | 6/7/83 | ? | upstm | 32 | Forest area | 5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 6/7/83 | BKB | upstm | 34 | B361 den | 5 | T |  |  |  |  | 2 |  |  |  |  |  |  |  |  |  |  |
|  | 6/8/83 | $?$ | upstm | 35 | + nematodes |  | 3 |  | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 6/8/83 | BKB | upstm | 40 | B372 den |  |  |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |
|  | 6/9/83 | BKB | upstm | 10 | B3 74 |  |  |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |
|  | 6/10/83 | BKB | upstm | 37 | B358 den |  | 2 |  | 2 |  | 2 |  | T | T |  |  |  |  | T |  |  |  |
|  | 6/9/83 | ? | dstm | 38 | Deadhorse Ck. | 5 |  |  |  |  |  |  |  |  |  |  |  | T |  |  |  |  |

Table 48. (continued)

1. Equisetum spp. (horsetail)
2. Lichens
3. Grasses or sedges 19. Clover (Trifollum spp.)

## Berries

2. Vaccinium vitis-iadea (lowbush cranberry)
3. Fmpetrum nigrum (crowberry)
4. Oplopanax horridus (devil's Club)
5. Arctosptaphylos alpina (bearberry)
6. Vaccintum uliginosumat (blueberry)
7. Streptopus amplexifolius (watermelon berry)
8. Other berries

Sambucus racemosa (red el derberry)
Oxycoccus microcarpus (bog cranberry)
Sorbus scopulina (Greene Mt. ashberry) Shepherdia canadensis (soapberry) - \#42 Cornus canadensis (Cornus berry)
Vaccintum ovalifolium (early blueberry) Viburnum edule (htghbush cranberry)

## Animal Matter

11. Moose
12. Hare or ground squirrel, misc.
13. Feathers
14. Fish
15. Insects
 Values are of volume (T=trace, $2=6+25 \%, 3=26-50 \%, 4=51-75 \%, 5=76-100 \%$ ).

|  | $\begin{aligned} & \text { Date } \\ & \text { Collected } \end{aligned}$ | $\begin{gathered} \text { Species of } \\ \text { bear } \end{gathered}$ | Place | $\begin{gathered} \text { Sample } \\ \text { No. } \end{gathered}$ | Comments | 1 | 2 | 3 | 4 | 5 | 6 | 9 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer - | - Sloughs |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/3/84 | ? | upstm | 6 | 1700' elev. |  | 2 |  | 2 |  |  |  |  |  |  |  |  | T |  |  |  | 4 |
|  | 8/5/84 | ? | upstm | 19 | Watana Camp | 2 | 2 |  |  |  |  | 3 |  |  |  |  | T |  |  |  | 3 |  |
|  | 8/5/84 | ? | upstm | 4 | Watana Camp |  | T |  |  |  |  |  |  |  |  |  | 2 | T |  |  | 5 |  |
|  | 8/15/84 | ? | dstm | 55 | Lane Ck. |  |  |  |  | 4 |  |  |  |  |  |  |  |  | 2 | 2 |  |  |
|  | 8/15/84 | $?$ | dstm | 60 | Slough 8B |  |  |  |  | 3 |  | 3 |  |  |  |  |  |  |  | 2 |  |  |
|  | 8/15/84 | ? | dstm | 64 | Portage Ck. S. |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  | T |  |  |
|  | 8/15/84 | ? | dstm | 65 | McKensie Ck. |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 5/15/84 | ? | dstm | 66 | Lane Ck. |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  | T |  |  |
|  | 8/16/84 | ? | dstm | 28 | Slough 28 |  |  |  |  | 5 |  |  |  |  |  |  |  | T | T |  |  |  |
|  | 8/16/84 | $?$ | dstm | 29 | Slough BA |  |  |  |  | 4 |  |  |  |  | T |  |  |  | 2 |  |  |  |
|  | 8/16/84 | $?$ | dstm | 30 | Slough A |  |  |  |  | 4 |  |  |  |  |  |  |  |  | 2 | 2 |  |  |
|  | 8/16/84 | BKB | dstm | 31 | Slough 9 |  |  |  |  | 3 |  |  |  |  |  |  | T |  | 3 | 2 |  |  |
|  | 8/16/84 | $?$ | dstm | 32 | Slough A |  |  |  |  | 3 |  |  |  |  | T |  |  |  | 3 | T |  |  |
|  | 8/16/84 | ? | dstm | 33 | Slough A |  |  |  |  | 3 |  |  |  |  |  |  |  |  | 3 | 2 |  |  |
|  | 8/16/84 | ? | dstm | 34 | Slough 11 |  |  |  |  | 3 |  | T |  |  |  |  | T | T | 3 | T |  |  |
|  | 8/16/84 | ? | dstm | 35 | Slough 8A |  |  |  |  | 3 |  |  |  |  |  |  |  |  | 3 |  |  |  |
|  | 8/16/84 | ? | dstm | 36 | Slough 9A |  |  |  |  | 5 |  |  |  | T |  |  |  |  |  | T |  |  |
| $\stackrel{\square}{\square}$ | 8/16/84 | ? | dstm | 37 | Slough 11 |  |  |  |  | 4 |  | T |  |  |  |  |  |  | 2 | 2 |  |  |
| $\checkmark$ | 8/16/84 | ? | dstm | 38 | Slough 11 |  |  |  |  | 4 |  |  |  |  |  |  |  |  | 2 | 2 |  |  |
| $\bigcirc$ | 8/16/84 | ? | dstm | 39 | Slough 9A | $T$ |  |  |  | 5 |  |  |  |  |  |  |  |  | T |  |  |  |
|  | 8/16/84 | ? | dstm | 40 | Slough 21 | 2 |  |  |  | 2 |  | 2 |  |  |  |  |  | T | 2 | 2 |  |  |
|  | 8/16/84 | ? | dstm | 41 | Slough 21 |  |  |  |  | 2 |  | 2 |  | T |  |  |  |  | 2 | 2 |  |  |
|  | 8/16/84 | $?$ | dstm | 42 | Slough 21 |  |  |  |  | 3 |  |  |  |  |  |  |  |  |  | 2 |  |  |
|  | 8/16/84 | ? | dstm | 43 | Slough 21 | 2 |  |  |  | 3 |  | 2 |  |  |  |  |  |  |  | T |  |  |
|  | 8/16/84 | ? | dstm | 44 | Slough 21 |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  | T |  |  |
|  | 8/16/84 | ? | dstm | 45 | 4th July Ck. |  | : |  |  | 4 |  |  |  |  |  |  |  |  | 3 | T |  |  |
|  | 8/16/84 | ? | dstm | 46 | Slough 8A |  |  |  |  | 4 |  | T |  |  |  |  |  |  | 2 |  |  |  |
|  | 8/16/84 | ? | dstm | 47 | Slough 11 |  |  |  |  | 2 |  |  |  |  |  |  |  |  |  | 5 |  |  |
|  | 8/16/84 | $?$ | dstm | 48 | Slough 8A |  |  |  |  | T |  |  |  |  |  |  |  | T | 3 | T |  |  |
|  | 8/16/84 | ? | dstm | 49 | Slough 9A |  |  |  |  | 3 |  |  |  |  |  |  |  |  |  | 3 |  |  |
|  | 8/16/84 | $?$ | dstm | 50 | Riverbank |  |  |  |  | 3 |  |  |  |  |  |  |  |  | 3 |  |  |  |
|  | 8/16/84 | $?$ | dstm | 51 | Slough 8A |  |  |  |  | T |  |  |  |  |  |  |  |  | 3 |  |  |  |
|  | 8/16/84 | ? | dstm | 52 | Slough 8A |  |  |  |  | 5 |  | T |  |  |  |  |  |  | 2 |  |  |  |
|  | 8/16/84 | $?$ | dstm | 53 | Slough 8A | $T$ |  |  |  | 4 |  |  |  |  |  |  |  |  | T | 2 |  |  |
|  | 8/16/84 | ? | dstm | 54 | 5th July Ck. |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/16/84 | $?$ | dstm | 56 | 5 th July Ck. |  | T |  |  | 2 | 3 |  |  |  |  |  |  |  |  | 3 |  |  |
|  | 8/16/84 | ? | dstm | 57 | 5th July Ck. |  |  |  |  | 3 |  |  |  |  |  |  |  |  | 2 | 2 |  |  |
|  | 8/16/84 | ? | dstm | 58 | 5th July Ck. |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 4 |  |  |
|  | 8/16/84 | $\stackrel{3}{8}$ | dstm | 62 | Slough 9 |  |  |  |  | 2 |  |  |  |  |  |  |  |  | 3 | 2 |  |  |
|  | 8/16/84 | BKB | dstm | 61 | Slough 8A |  |  |  |  | 2 |  |  |  |  | 2 |  |  |  | 3 | T |  |  |
|  | 8/16/84 | ? | dstm | 59 | Slough A |  |  |  |  | 5 |  |  |  |  |  |  |  |  | $T$ | T |  |  |
|  | 8/16/84 | $?$ | dstm | 63 | Slough 9 |  |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8/23/84 | ? | upstm | 15 | E. Fk. Watana | 2 | T |  |  |  |  | 3 |  |  |  |  |  |  |  |  | 3 |  |
|  | 8/25/84 | ? | upstm | 16 | E. Fk. Watana | 3 | T |  |  |  |  | 3 |  |  |  |  |  | T |  |  | 3 |  |

Table 49. (cont'd)

| Date Collected | Spectes of bear | Place | $\begin{gathered} \text { Sample } \\ \text { No. } \end{gathered}$ | Comments | 1 | 2 | 3 | 4 | 5 | 6 | 9 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SPRING SAMPLES |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5/15/84 | BRB 299 | upstm | 7 | Susitna |  | 2 |  | 4 |  |  |  | T |  |  |  |  |  |  |  |  |
| 5/15/84 | BRB 418 | upstm | 5 | Ylg w/299 |  | 5 |  |  |  |  |  | T |  |  |  |  |  |  |  |  |
| 5/15/84 | BRB 417 | upstm | 11 | Ylg w/299 |  | T |  | 3 |  |  |  | 3 |  |  |  |  | $T$ |  |  |  |
| 5/15/84 | BRB 419 | upstm | 12 | ylg w/299 |  |  |  | 5 |  |  |  | T |  |  |  |  | T |  |  |  |
| 5/15/84 | BRB 399 | upstm | 14 | Susitna |  | T |  |  |  |  | 3 | 4 |  |  |  |  |  |  |  |  |
| 5/16/84 | BRB 312 | upstm | 8 | Stomach |  | $T$ |  |  |  |  | T |  | 5 |  |  |  |  |  |  |  |
| 5/16/84 | BKB 349 | upstm | 1 | Anal plug |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5/18/84 | BRB 422 | upstm | 9 | On old moose kill |  | 2 |  | 2 |  |  |  | 4 |  |  |  |  | T |  |  |  |
| 5/27/84 | BRB | upstm | 10 | On calf kill |  | T |  | 2 |  |  |  | 5 |  |  |  |  | T |  |  |  |
| 5/27/84 | BRB | upstm | 21 | On calf kill |  | 2 |  | 2 |  |  |  | 3 |  |  |  |  | T |  |  |  |
| 5/29/84 | BRB cub | upstm | 3 | Abandoned cub |  |  |  | 3 |  | 2 |  | T |  |  |  |  | T | 2 |  |  |
| 5/30/84 | BRB | upstm | 17 | On calf kill | 2 |  |  |  |  |  |  | 5 |  |  |  |  | $T$ |  |  |  |
| 5/31/84 | BRB | upstm | 2 | On calf kill |  |  |  | 4 |  |  |  | T |  | 2 |  | T |  |  |  |  |
| 5/31/84 | BRB | upstm | 13 | On calf kill | 5 | 2 |  |  |  |  |  | T |  |  |  |  | T |  |  |  |
| 5/31/84 | BRB | upstm | 18 | On calf kill | 2 | 2 |  | 2 |  |  | 3 | 3 |  |  |  |  | T |  |  |  |
| 6/20/84 | BKB | upstm | 20 | den of B401 | 3 | 3 |  | 2 |  |  |  | T |  |  |  |  | T |  |  |  |

1. Equisetum spp. (horsetail)
2. Lichens
3. Grasses or sedges
4. Clover (Trifolium spp.)

## Berries

2. Vaccinium vitis-idaea (lowbush cranberry)
3. Empetrum nigrum (crowberry)
4. Oplopanax horridus (devil's Club)
5. Arctosptaphyios alpina (bearberry)
6. Vaccinium uliginosum (blueberry)
7. Streptopus amplexifolius (watermelon berry)
8. Other berries

Sambucus racemosa (red el derberry)
Oyycoccus microcarpus (bog cranberry) Sorbus scopulina (Greene Mt. ashberry) Shepherdla canadens is (soapberry) - \#42 Corrius canadensis (Cornus berry)
Vaccintum ovalifolium (early blueberry) Viburnum edule (highbush cranberry) R1bes triste (red currant)

AnImal Matter
16. Other Misc.
11. Moose
12. Hare or ground squirrel, misc.
13. Feathers
14. Fish
15. Insects

Table 50. Salmon abundance in downstream sloughs and streams, 1981-1984.

| AREA | No. Adult Salmon Enumerated* |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | RIVER MILE | 1981( ${ }^{* * * \text { ) }}$ | 1982 ( ${ }^{* *}$ ) | 1983(N**) | 1984 ( ${ }^{* *}$ ) |
| Slough 21 | 141.0 | 747 (5) | 2424 (9) | 1904 (13) | 7197 (9) |
| Slough 11 | 135.3 | 5483 (9) | 4806 (11) | 5067 (23) | 9749 (8) |
| Slough 8A | 125.1 | 1283 (5) | 1804 (10) | 843 (20) | 3054 (8) |
| Slough 20 | 140.0 | 27 (2) | 220 (7) | 201 (20) | 695 (4) |
| Slough 9A | 133.3 | 484 (6) | 146 (3) | 217 (3) | 574 (5) |
| Moose Slough | 123.5 | 555 (5) | 115 (7) | 392 (15) | 405 (5) |
| Slough 8B | 122.2 | 1 (1) | 190 (6) | 240 (6) | 1749 (8) |
| Slough 8C | 121.9 | (0) | 105 (3) | (0) | 416 (5) |
| Slough 17 | 138.9 | 169 (7) | 29 (4) | 182 (8) | 240 (4) |
| Slough 15 | 137.2 | 1 (1) | 178 (3) | 20 (5) | 611 (1) |
| Slough B | 126.3 | NA | 225 (6) | 9 (1) | 196 (5) |
| Slough 9 | 128.3 | 380 (5) | 911 (6) | 1081 (9) | 499 (3) |
| Slough 6A | 112.3 | 27 (3) | 101 (4) | 2 (1) | 3 (1) |
| Sloughs A \& A | 124.7 | 437 (10) | (0) | 528 (16) | 338 (5) |
| Slough 8 | 113.7 | 858 (5) | (0) | (0) | 193 (6) |
| Slough 9B | 129.2 | 678 (7) | (0) | (0) | 181 (3) |
| Slough 19 | 139.7 | 84 (6) | (0) | 18 (6) | 147 (7) |
| Slough 22 | 144.5 | NA | NA | 274 (4) | 199 (3) |
| Mainstream Zone 3 | 135.2 | NA | NA | 252 (2) | No data |
| Slough 2 | 100.2 | 44 (5) | 0 | 103 (4) | 287 (9) |
| Indian River*** | 138.6 | 232 (7) | 6703 (12) | 7958 (16) | 14898 (9) |
| Lane Ck | 113.6 | 569 (7) | 2508 (11) | 118 (9) | 2837 (9) |
| 4th of July Ck. | 131.0 | 247 (6) | 2832 (11) | 636 (9) | 6160 (7) |
| Little Portage Ck . | 117.7 | NA | 407 (9) | 10 (2) | 384 (7) |
| Lower McKenzie Ck. | 116.2 | 97 (6) | 492 (6) | 46 (6) | 1067 (7) |
| 5th of July Ck. | 123.7 | 2 (1) | 224 (4) | 24 (4) | 834 (5) |
| Skull Ck. | 124.7 | 24 (3) | 36 (4) | 1 (1) | 216 (3) |
| Portage Ck. | 148.9 | 22 (1) | 2238 (7) | 4651 (13) | 15319 (19) |

Table 50. (cont'd)

| AREA | RIVER MILE | No. Adult Salmon Enumerated* |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1981(N**) | 1982 ( ${ }^{* * *}{ }^{\text {a }}$ ) | $1983\left(N^{* *}\right)$ | 1984 (N**) |
| Gash Ck. | 111.6 | 258 (2) | 163 (3) | 35 (2) | 711 (7) |
| Slash Ck. | 111.2 | NA | 6 (1) | 2 (1) | 8 (2) |
| Whiskers Ck. | 101.4 | 212 (7) | 626 (5) | 273 (9) | 899 (11) |
| Jack Long Ck. | 144.5 | 1 (1) | 54 (7) | 19 (5) | 27 (3) |
| Deadhorse Ck | 120.9 | 0 | NA | NA | 378 (2) |
| Upper McKenzie Ck. | 116.7 | 0 | 24 (2) | (0) | 23 (3) |
| Chase Ck. | 106.9 | 328 (8) | 332 (8) | 26 (5) | 1523 (9) |
| Gold Ck. | 136.7 | 0 | 37 (3) | 51 (3) | 83 (1) |
| Sherman Ck. | 130.8 | 32 (4) | 40 (4) | (0) | 126 (3) |

* These data sum all live and dead fish (Chinook, Sockeye, Pink, Chum, and Coho Salmon) recorded by Su-Hydro AA personnel (ADF\&G) during stream surveys. Different areas were surveyed from 1 to 11 times during the year which contributes to variation observed between areas and between years in this data, survey conditions also varied. Note that the same fish would likely be recorded numerous times in replicate surveys.
** N is the number of surveys conducted where salmon were enumerated, surveys where no salmon were seen are not counted.
*** The portion of the Indian River evaluated by Fisheries personnel varied in 1981 and 1982. Most fish were found in 1982 in a tributary about $\frac{1}{2}$ mile up from the mouth (Crowe, per. commun.) during our investigation of the Indian River we did not observe this location.

Table 51. Ranking of bear and salmon use of downstream sloughs and creeks on 24-25 August, 1983. ( $0=1$ owest on scale of $0-10$ ).
Index of Index of Apparent use by radio-

Slough No. salmon presence bear use Comments collared individuals

| 7 | 0 | 1 | entrance into slough |
| :--- | :--- | :--- | :--- | :--- |
| blocked |  |  |  |

* Had been lots of rain and sloughs were very high and muddy, salmon were difficult to spot in the sloughs.

Table 52. Ranking of bear and salmon use of downstream sloughs and creeks on 15-16 August, 1984. ( $0=1$ lowest on scale of $0-10$ ).

| Slough No.Index of <br> salmon presence | Index of <br> bear use | apparent use by radio- <br> collared individuals |
| :---: | :---: | :---: |


| 8 | 1 | 4 |  |  |
| :---: | :---: | :---: | :---: | :---: |
| 8A | 8 | 6 | some salmon eaten | B404, G379 |
| 8B | 3 | 6 |  |  |
| 8 C | 1 | 2 | . |  |
| 8D | 0 | 1 |  |  |
| A | 0 | 1 |  | B343, |
| $A^{\prime}$ | 4 | 1 | . | $\cdots$ |
| 9 | 3 | 2 |  |  |
| 9 B | 3 | 2 |  | G379 |
| 9A | 2 | 2 |  | B409 |
| 10 | ND | ND |  | B411 |
| 11 | 9 | 2 |  |  |
| 17 | 3 | 1 |  | . |
| 20 | 4 | 3 | . |  |
| 21 | 5 | 6 | salmon eaten |  |
| Lane Ck | 7 | 5 | lots of Pinks, some eaten |  |
| Lower McKensie Ck | 3 | 2 |  |  |
| McKensie Ck | 2 | 1 |  |  |
| Portage Ck | 3 | 2 | some salmon eaten |  |
| Deadhorse Ck | 2 | 2 | entrance perched |  |
| Moose and Clear Creeks | 1 | 3 |  |  |
| 5 th of July | 8 | 7 |  | B376 |
| 4 th of July | 7 | 8 | many salmon eaten | B405 |

Table 53. Summary of black bear litter size data based on observations of bears with litters of newborn cubs.

| MOTHER'S ID (age-year) | LITTER SIZE | COMMENTS |
| :---: | :---: | :---: |
| B289 (10 in spring '81) | 3 | lost 1 in August, 2 survived |
| B289 (12 in spring '83) | 2 | lost 1 cub in September, other survived to den exit |
| B289 (14 in spring '85) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | both survived to yearling age |
| B301 (8 in spring '81) | 2 | both survived to yearling age |
| B301 (10 in spring '83) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit }]} \end{aligned}$ | survivorship undetermined, female shed collar |
| B317 (7 in summer '80) | 2 (summer) | initial capture in summer, both survived to fall, cubs not seen with bear at initial capture |
| B317 (10 in '83) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | lost 1 in June, other survived to den exit |
| B317 (12 in spring '85) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | ```l survived to den entrance, l lost in July``` |
| B318 (5 in summer '80) | 1 (summer) | survived |
| Bi318 (8 in '83) | $\begin{aligned} & 2 \text { (den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | both lost by 6/6/83 apparently, shed collar |
| B328 (7 in summer '81) | 2 (summer) | ```bred in 1980. Lost 1 by 7/29/81, shed collar in den (not sure if survived until exit)``` |
| B328 (11 in spring '85) | $\begin{aligned} & 3 \text { (in den) } \\ & {[3 \text { at exit] }} \end{aligned}$ | lost 6/6-7/24 |
| E326 (5 in summer '80) | 2 (summer) | bear shot in 1980, cubs may have been adopted by B317 |
| E321 (11 in spring '81) | 2 | no cubs in summer 1980, both cubs lost by $8 / 24 / 81$, no litter in ' 82 , no litter verified in 1983 but may have lost a litter early in 1983, bred in 1983 |
| E321 (14 in '84) | 2 | lost 1 of 2 by 6/29, other survived to den entrance |
| E327 (5 in summer '80) | 2 (summer) | both survived to yearling age |
| B327 (8 in '83) | $\begin{aligned} & 2 \text { (den) } \\ & {[2 \text { at exit }]} \end{aligned}$ | cubs survived into June, female died in July |
|  | (continued | next page) |

Table 53. (cont'd)

| MOTHER'S ID (age-year) | LITTER SIZE | COMMENTS |
| :---: | :---: | :---: |
| B349 (6 in spring '83) | $\begin{aligned} & 2 \text { (den) } \\ & {[0 \text { at exit?] }} \end{aligned}$ | first litter, no cubs in summer ' 81 or spring ' 82 , cubs apparently lost in May '83, collar shed in July - no ylgs on 5/84 |
| B349 (8 in spring '85) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit }]} \end{aligned}$ | one survived to den entrance, 1 lost in August |
| B354 (5 in '82) | 2 | both survived to den entrance, at least 1 ylg at exit in ' 83 |
| B354 (7 in '84) | 2 | may have lost 1 by den entrance date |
| B354 (9 in '86) | 2 | both survived to den entrance |
| B361 (8 in '83) | $\begin{aligned} & 4 \text { (in den) } \\ & {[3 \text { at exit] }} \end{aligned}$ | lost 1 in den prior to exit, others survived to den exit in ' 84 |
| B370 (8 in '83) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | bear missing after 5/23/83, cubs alive at that time |
| B363 (6 in '84) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit }]} \end{aligned}$ | None lost to den entrance |
| B364 (10 in ${ }^{186}$ ) | 2 | both survived to den entrance |
| B369* (6 in '84) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit }]} \end{aligned}$ | none lost to den entrance |
| B372* (10 in '83) | $\begin{aligned} & 3 \text { (in den) } \\ & \text { [3 at exit] } \end{aligned}$ | lost 1 in early July, others survived to $7 / 20$, female lost in. September ' 83 |
| B374* (7 in '83) | 3 | think lost 2 in July, bear shot in September '83 |
| B375* (6 in '83) | 2 | both survived to exit in ${ }^{\text { }} 84$ |
| B376* (5 in ${ }^{\text { }} 83$ ) | $\begin{aligned} & 3 \text { (in den) } \\ & {[3 \text { at exit }]} \end{aligned}$ | all survived to exit in ' 84 |
| B377* (5 in '83) | $[1-2 ? ?]$ <br> NOT COUNTED | cubs may have been lost prior to or during capture, cubs not seen during capture but saw at least 1 cub 9 days earlier on 5/10/83 |
| B377 (6 in '84) | some (in den) [0 at exit] | heard at least 1 cub in den, none seen at exit |
| B377 (7 in '85) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit }]} \end{aligned}$ | lost 1 in June, other in AugustSeptember |


| MOTHER'S ID (age-year) | LITTER SIZE | COMMENTS |
| :---: | :---: | :---: |
| B378* (7 in ${ }^{\text {' }} 83$ ) | $\begin{aligned} & \text { 2(den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | both survived to '84 den exit |
| B378* (9 in '85) | 1 | survived to den entrance |
| B379 (9 in '83) | $\begin{aligned} & 3 \text { (den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | lost all cubs by 5/23/83, bred again, died in July |
| B402* (12 in '85) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit] }} \end{aligned}$ | both survived to den entrance |
| B404* (11 in '83) | 1 | survived thru 7/20/83 at least, not seen in ' 84 |
| B405* (17 in' 83 ) | 2 | both survived to den exit in ' 84 |
| B406* (11 in '83) | 2 | both survived to den exit in ' 84 |
| B409* (?)(6 in '84) | ? | not observed in '84 |
| B409* (8 in ' 86 ) | [2(?)] | data not conclusive, not included in means |
| B410* (7 in '83) | 2 | both survived thru June, bear shot in July |
| B411* (9 in '84) | 2 | status at entrance into '84 den unknown |
| B438 (9 in '86) | 3 | B438 probably shot by 9/5/86, cub status unknown |

Table 53. (cont'd)

| Total number <br> of cubs | Number of <br> litters | Mean litter <br> size (range) |
| :---: | :---: | :---: | | Comments (includes) |
| :--- |
| 90 |

Downstream study area

| MOTHER'S ID (age-year) | LITTER SIZE | COMMENTS |
| :---: | :---: | :---: |
| B288 (10 in 1980) | 3 | bred in 1980, ylgs with female into August, shed collar in 1980 |
| B290 (8 in 1980) | 2 | weaned by $6 / 23 / 80$, bred in 1981 , collar removed on $8 / 5 / 81$ (neck scarred) |
| B289 (9 in 1980) | 2 | weaned by $5 / 22 / 80$, bred, 3 cubs in ' 81 |
| B289 (13 in 1984) | 1 | with mom to September bred in June |
| B289 (11 in 1982) | 2 (in den) | weaned by $6 / 9 / 82$, bred, had 2 cubs in 1983 |
| B289 (15 in 1986) | 2 | weaned by 7/9/86 |
| B301 (7 in 1980) | 1 | weaned by $6 / 12 / 80$, bred, had 2 cubs in 1981 |
| B301 (9 in 1982) | 2 | weaned-by $6 / 17 / 82$, bred, had 3 cubs in 1983 |
| B317 (8 in 1981) | 2 | weaned by $6 / 18 / 81$, bred, 1 y 1 g returned and was with female until 9/9/81, no cubs in 1982 |
| B317 (11 in 1984) | 1 | weaned in June, bred |
| B318 (6 in 1981) | 1 (den) | ylg (B330) weaned by $5 / 29 / 81$, bred, ylg died by $8 / 24 / 81$, no (reason?) cubs in 1982, bred again, 2 cubs in 1983 |
| B318 (10 in 1985) | 2 | B318 not located after 6/11/85 |
| B327 (5 in 1981) | 2 (den) | y1g B329 and sibling, sibling weaned by $6 / 5 / 81$, B329 by $6 / 21$, bred, no cubs in 1982, bred again, cubs in 1983 |
| B349 (9 in 1986) | 1 |  |
| B354 (6 in 1983) | 1 (?) | at least 1 ylg exited den (perhaps both?), weaned by 6/2/83 |

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 updated 9/86Table 54. (cont'd)
MOTHER'S ID (age-year) LITTER SIZE COMMENTS

| B363 (8 in 1985) | 2 |
| :---: | :---: |
| B364 (8 in 1984) | 3 |
| B369* (7 in 1985) | $\begin{aligned} & 2 \text { (in den) } \\ & {[2 \text { at exit }]} \end{aligned}$ |
| B402* (10 in 1983) | 3 |
| B402* (13 in 1986) | 2 |
| B411* (8 in 1983) | 2 |
| B321 (15 in 1986) | 1 |
| B361 (9 in 1984) | 3 |
| B375* (11 in 1984) | 2 |
| B376* (8 in 1984) | 3 |
| B378* (8 in 1984) | 2 |
| B404* (12 in 1984) | [?] |
| B405* (18 in 1984) | 2 |
| B406* (12 in 1984) | 2 |
| B432 (6 in 1985) | 1 |

weaned by $9 / 4 / 85$
2 weaned early, bred, still with one in September
weaned in early July
weaned by September
weaned after 6/13
weaned by $6 / 27 / 85$
entered den w/mon, weaned at age 2
weaned in June
weaned 2 in June, 1 with mon in October

Not seen after June
'84 status not verified
with mom into August
weaned by September
weaned by $6 / 3 / 85$

| Total number . <br> of ylgs. observed | number of <br> litters | mean litter size (range) |
| :--- | :---: | :--- | comments | all litters with |
| :--- |
| 54 |

[^8]Table 55. Sex ratio and morphometrics of black bear cubs-of-year handled in the Susitna Hydro Project.

| $\begin{aligned} & \text { CUB } \\ & \text { ID } \\ & \hline \end{aligned}$ | $\begin{gathered} \text { MOTHER 'S } \\ \text { ID } \\ \hline \end{gathered}$ |  | $\begin{gathered} \text { DATE } \\ \text { HANDLED } \end{gathered}$ |  | SEX | WT (lbs) | COMMENTS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 355 | B354 | 26 | May 1982 |  | F | -- | ear tags |  |
| 356 | B354 | 26 | May 1982 |  | M | -- | ear tags |  |
| -- | B301 | 20 | March 1983 | (den) | F | 2.6 |  |  |
| -- | B301 | 20 | March 1983 | (den) | F | 2.5 |  |  |
| -- | B361 | 21 | March 1983 | (den) | M | 3.5 |  |  |
| -- | B361 | 21 | March 1983 | (den) | F | 3.8 |  |  |
| -- | B361 | 21 | March 1983 | (den) | F | 3.5 |  |  |
| -- | B361 | 21 | March 1983 | (den) | F | 2.8 |  |  |
| -- | B349 | 22 | March 1983 | (den) | F | 3.5 |  |  |
| -- | B349 | 22 | March 1983 | (den) | F | 3.4 |  |  |
| -- | B317 | 23 | March 1983 | (den) | M | 4.3 | neck $=175 \mathrm{~mm}$ |  |
| -- | B317 | 23 | March 1983 | (den) | M | 4.3 | neck $=180 \mathrm{~mm}$ |  |
| -- | B318 | 23 | March 1983 | (den) | M | 2.8 |  |  |
| -- | B318 | 23 | March 1983 | (den) | F | 2.7 | - |  |
| -- | B327 | 23 | March 1983 | (den) | M | 5.3 | neck $=190 \mathrm{~mm}$ |  |
| -- | B327 | 23 | March 1983 | (den) | F | 4.5 | neck $=180 \mathrm{~mm}$ |  |
| -- | B379 | 24 | March 1983 | (den) | M | 2.8 |  |  |
| -- | B379 | 24 | March 1983 | (den) | M | 3.3 |  |  |
| -- | B379 | 24 | March 1983 | (den) | M | 3.3 |  |  |
| -- | B372 | 15 | April 1983 | (den) | F | 3.7 |  |  |
| -- | B372 | 15 | April 1983 | (den) | F | 4.1 |  |  |
| --- | B372 | 15 | April 1983 | (den) | M | 4.5 |  |  |
| -- | B376 | 16 | April 1983 | (den) | M | 6.0 | neck $=190 \mathrm{~mm}$ |  |
| -- | B376 | 16 | April 1983 | (den) | F | 5.5 | neck=190mm |  |
| -- | B376 | 16 | April 1983 | (den) | $F$ | 5.8 | neck $=190 \mathrm{~mm}$ |  |
| -- | B370 | 16 | April 1983 | (den) | F | 7.5 | neck $=200 \mathrm{~mm}$ |  |
| -- | B370 | 16 | April 1983 | (den) | F | 7.0 | neck $=190 \mathrm{~mm}$ |  |
| 010 | B374 | 19 | May 1983 |  | F | -- | neck $=175 \mathrm{~mm}$, ear tags |  |
| 011 | - B374 | 19 | May 1983 |  | F | -- | neck $=200 \mathrm{~mm}$, ear tags |  |
| 012 | B374 | 19 | May 1983 |  | F | -- | neck $=195 \mathrm{~mm}$, ear tags |  |

(continued on next page)

Table 55. (cont'd)

| $\begin{aligned} & \text { CUB } \\ & \text { ID } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { MOTHER 'S } \\ & \text { ID } \\ & \hline \end{aligned}$ |  | $\begin{gathered} \text { DATE } \\ \text { HANDLED } \\ \hline \end{gathered}$ |  | SEX | WT (1bs) | COMMENTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 013 | B404 | 19 | May 1983 |  | F | 10.0 | neck $=215 \mathrm{~mm}$, ear tags |
| 014 | B405 | 19 | May 1983 |  | F | 6.5 | neck $=180 \mathrm{~mm}$, ear tags |
| 015 | B405 | 19 | May 1983 |  | F | 6.0 | neck $=175 \mathrm{~mm}$, ear tags |
| -- | B363 | 6 | April 1984 | (den) | M | 6.0 | neck $=190 \mathrm{~mm}$ |
| -- | B363 | 6 | April 1984 | (den) | M | 6.8 | neck $=192 \mathrm{~mm}$ |
| -- | B369 | 8 | April 1984 | (den) | M | 4.0 |  |
| -- | B369 | 8 | April 1984 | (den) | F | 3.8 |  |
| -- | B349 | 28 | Feb. 1985 | (den) | M | 1.8 | very small, eyes closed, sibling not handled |
| -- | B328 | 29 | March 1985 | (den) | M | 5.0 |  |
| -- | B328 | 29 | March 1985 | (den) | M | 4.1 |  |
| -- | B328 | 29 | March 1985 | (den) | F | 4.1 |  |
| -- | B404 | 30 | March 1985 | (den) | M | 4.1* |  |
| -- | B404 | 30 | March 1985 | (den) | M | 4.1* | - |
| -- | B404 | 30 | March 1985 | (den) | F | 3.5* |  |

Totals: 19 males and 25 females, In dens=18 males and 18 females.

* Estimated

Table 56. Morphometrics of black bear yearlings handled in the Susitna Hydro project, 1980-1985.


Totals: 5 males and 5 females.

Table 57. Summary of known losses of black bear cubs-of-the-year. Losses calculated during first season out of den (in dens or at emergence from dens as cubs to entrance into dens as cubs)

| Year | Upstream study area | Downistream study area | Both areas |
| :---: | :---: | :---: | :---: |
| 1980 | no data | no data | -- |
| 1981 | $\begin{aligned} & 4 \text { of } 9 \text { lost }(289,301, \\ & 321,328) \end{aligned}$ | no data | 4 of 9 lost |
| 1982 | 0 of 2 lost (354) | no data | 0 of 2 lost |
| 1983 complete data | $\begin{aligned} & 8 \text { of } 13 \text { lost }(289,317, \\ & 361,349) \end{aligned}$ | $\begin{aligned} & 1 \text { of } 12 \text { lost }(375,376 \text {, } \\ & 377 * *, 378,405,406) \end{aligned}$ | 9 of 25 lost |
| 1983 incomplete data* | [2 of 2 lost (318] | [3 of 6 lost ( 372,374 )] | [5 of 8 lost] |
| 1984 complete data | 1 of 4 lost ( 321,363 ) | 0 of 2 lost (369) | 1 of 6 lost |
| 1984 incomplete data* | [1 of 2 lost (354)] | [1 of ? lost (377)] | [1 of 2 lost] |
| 1985 complete data | $\begin{aligned} & 7 \text { of } 11 \text { lost }(289,317, \\ & 328,349,377) \end{aligned}$ | 0 of 3 lost (378, 402) | 7 of 14 lost |
| 1986 complete data*** | 0 of 4 lost (354, 364) | 0 of 0 lost | 0 of 4 10st |
| TOTALS (all years) | 20 of $43=47 \%$ lost | 1 of $17=6 \%$ lost | 21 of $60=35 \%$ lost |

* incomplete data resulted from not observing the family status of the bear before it entered its winter den, shed collars, collar failures, or early hunter kills. Tabulated losses occurred prior to loss of the female to these causes. These are not included in totals.
** 3377 may have lost 2 of 2 rather than the 1 of 1 tabulated in 1983, the initial litter size was not known with certainty.
*** B438 and B409 had inadequate data.

Table 58．Reproductive histories of radio－marked female black bears．（＂Shed＂refers to removal by bear of radio collar．）Bears were in upstream study area unless otherwise indicated．

| Year | 289 （9 in ． 80 ） | 290 （8 in＇80） | $301\left(7 \mathrm{in}^{\prime} 80\right)$ | 317 （7 in $\left.{ }^{\prime} 80\right)$ |
| :---: | :---: | :---: | :---: | :---: |
| 1980 | w／2＠1 weaned in May－bred | w／2＠l weaned in June | w／1＠1 weaned in June | w／200 in Aug． |
| 1981 | w／3＠0， 1 lost In Aug． | alone，bred collar removed | W／2＠0， | ```w/2@⿴囗⿱一一⿻儿口一寸, weaned In June, bred, reunitd w/1@1 thru Sept.``` |
| $\overline{1982}$ | weaned 2＠1，May－June，bred | －－ | w／2＠1，weaned in June， bred | no newborns，possibly w／1＠2 into June， |
| 1983 | w／2＠0， 1 lost in Sept．， |  | w／2＠0，shot in Sept． | w／2＠0， 1 lost in June |
| 1984 | ```weaned 1@l in May, bred, reunited June-Sept. weaned in Sept.``` |  | －－ | ```w/l@l, weaned, June, bred, reunited predenning``` |
| 1985 | w／2＠0，survived |  | － | ```w/2@0, l lost in July, other okay thru Sept. at least``` |
| 1986 | w／2＠1，weaned（date？） |  | －－ | alone in June |

Table 58. (cont'd)

| Year | $\begin{array}{rl} 318 \\ 5 & \mathrm{in} \quad 80 \\ \hline \end{array}$ | $\begin{gathered} 321 \\ 10^{\mathrm{in} \cdot} 80 \\ \hline \end{gathered}$ | $\begin{aligned} & 325 \\ & 11 \text { in } 180 \\ & \hline \end{aligned}$ | $\begin{aligned} & 327 \\ & 5 \text { in ' } 80 \\ & \hline \end{aligned}$ | $\begin{array}{r} 328 \\ 6 \text { in } 80 \\ \hline \end{array}$ | $\begin{array}{r} 329 \\ 1 \quad \mathrm{in} \text { ' } 81 \\ \hline \end{array}$ | $\begin{array}{r} 349 \\ 4 \text { in }^{\prime} 81 \\ \hline \end{array}$ | $\begin{aligned} & 354 \\ & 5 \text { in } 182 \\ & \hline \end{aligned}$ | $7^{361} \mathrm{in}^{\prime} 82$ | $\begin{aligned} & 363 \\ & 4 \\ & 4 \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | w/1@0 <br> in Aug. | alone in Aug. | alone in Aug. | w/2@o in in Aug. | alone in Aug. | with <br> mother 327 | -- | -- | -- | -- |
| 1981 | w/l@1, weaned in May, bred | w/2@o, lost both in Aug. | alone, shed in next den | w/2@1 in den, 1 weaned in May, other in June, bred | W/2 20 , 1 lost in July, other okay thru Sept., collar shed | weaned from 327 in June | alone | -- | -- | -- |
| 1982 | alone | alone | -- | alone, bred | ? | alone | alone | w/2œ0 to den entrance | alone | alone, bred? |
| 1983 | W/2@o, suspect lost both June, shed | think lost litter very early, bred | -- | $\begin{aligned} & \text { w/2@o, } \\ & \text { mother } \\ & \text { died in } \\ & \text { July } \end{aligned}$ | ? | alone, bred? | w/2@0, both lost in den | w/1@1 weaned in May, bred | ```W/4@0 in den, l lost in den``` | alone, bred |
|  | ```must have had at least 2@0 based on 1985]``` | $\begin{aligned} & \text { w/I@o } \\ & \text { (in July) } \end{aligned}$ | -- | -- | alone, bred | alone, bred? | alone | $\begin{aligned} & \text { w/2@0, } \\ & \text { l lost in } \\ & \text { Sept. } \end{aligned}$ | $\begin{aligned} & \text { W/3ब1 } \\ & \text { not } \\ & \text { weaned-- } \\ & \text { seen in den } \end{aligned}$ | $\begin{aligned} & \text { w/2@0 } \\ & \text { survived } \end{aligned}$ |
| 1985 | w/2@1 in June when reported | w/191 weaned in June | -- | -- | $\begin{aligned} & \text { w/3@0, all } \\ & \text { lost in } \\ & \text { June-July } \end{aligned}$ | alone, bred? | w/2@o <br> in den, 1 <br> lost in <br> Aug. | $\begin{aligned} & \text { alone } \\ & \text { (June) } \end{aligned}$ | W/3@2, weaned in June | W/2@I weaned, date? |
| 1986 | ? | alone | -- | -- | alone | alone | w/1@1, weaned (date?) | ```w/2@0 (Sept.), l lost in Sept.?``` | $\begin{aligned} & \text { alone in } \\ & \text { June } \end{aligned}$ | alone, bred |

Table 58. (cont'd)

| Year | $\begin{aligned} & 364 \\ & 6 \text { in }^{\prime} 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 367 \\ & 4 \text { in ' } 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 369 \\ & 4 \text { in } 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 370 \\ & .7 \text { in '82 } \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 372 \\ & 9 \text { in } 182 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 374 \\ & 7 \text { in } 182 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 375 \\ & 9 \text { in } 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 376 \\ & 6 \text { in ' } 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 377 \\ & 4 \text { in ' } 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 378 \\ & 6 \text { in ' }^{\prime} 82 \end{aligned}$ | $\begin{aligned} & \text { Downstream } \\ & 402 \\ & 10 \text { in ' } 83 \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | alone, bred, collar failed | alone | alone | alone | alone, bred | alone? | w/3@1? | alone? | alone | alone | - |
| 1983 | [must have had cubs based on 1984] | alone- shot | alone | $\begin{aligned} & \text { w/2@o, } \\ & \text { failed } \\ & \text { collar } \end{aligned}$ | $\begin{aligned} & \text { w/2@o, } \\ & \text { failed } \\ & \text { collar } \end{aligned}$ | w/3@0, 2 died in July, shot in fall | $\begin{aligned} & \text { w/2@o, } \\ & \text { survived } \end{aligned}$ | W/3@0 | alone? | W/2@0, survived | w/3@1, weaned in June |
| 1984 | w/3@1, weaned in June-July, bred, reunited w/1 in Sept. | - | $\begin{aligned} & \text { 2@o } \\ & \text { in den } \\ & \text { lost } 1 \\ & \text { in Sept. } \end{aligned}$ | -- | -- | -- | w/2@1 weaned in July | w/3@1, weaned in May, reunited in July and Sept. | alone | w/2@1, weaned | alone |
| $\infty_{\infty}^{\infty} \overline{1985}$ | w/1@2 in June | -- | W/1@1 weaned in JuneJuly | -- | -- | -- | shot in spring | alone? | w/2@0, 1 lost in June, other in July Aug. | W/1@0, survived | w/2@0 |
| 1986 | $\begin{aligned} & \text { w/2@o, } \\ & \text { survived } \\ & \text { thru Sept. } \end{aligned}$ | -- | alone? | -- | -- | - | -- | alone | alone | alone | $\begin{aligned} & \text { w/2@1, } \\ & \text { survived } \end{aligned}$ |

Table 58. (cont'd)


Table 59. Summary of reproductive intervals for black bears by bear ID. (* indicates bear from downstream study area. Year of litter and reason for intervals >2 years are indicated in parentheses - "lost" means lost complete litter).

COMPLETE INTERVALS OF:
2 YEARS
3 YEARS
4 YEARS
5 YEARS

| 289 | (81) | 363 (84) | 317 (85, skipped 1) | 318 (83, lost 2) | 321 (84, lost 1-2) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 289 | (83) | 364 (83) | 361 (83, weaned @2) | 349 (85, 1 lost, 1 skip) |  |
| 289 | (85) | 369* (84) | 402* (85, skipped 1) |  |  |
| 301 | (81) | 375* (83) | 405* (83, weaned @2) |  |  |
| 317 | (80) | 376* (83) |  |  |  |
| 318 | (80) | 378* (83) |  |  |  |
| 327 | (80) | 378* (85) |  |  |  |
| 354 | (82) | 406* (83) |  |  |  |
| 354 | (84) | 410* (84) |  |  |  |

INCOMPLETE INTERVALS THAT WILL BE AT LEAST INDICATED LENGTH:


Table 60. Summary of age at first reproduction for Su-hydro area black bears by bear ID. Based on first observed litter, status in previous year(s) is given in parentheses.

FIRST REPRODUCTION AT AGE:

| 5 YEARS | 6 YEARS | 7 YEARS | 8 YEARS |
| :---: | :---: | :---: | :---: |
| 327 (?) | 349 (alone prev. 2) | 377 (alone prev. 3) | 448 (alone prev. 2 expected '87) |
| 354 (?) | 363 (alone prev. 2) | 409 (alone prev. 2) | *361 (alone prev. 1) |
| 432 (?) | 369 (alone prev. 2) | 329 (expected '87) | *370 (alone prev. 1) |
|  | 328 (alone prev. 1) | *374 (alone prev. 1) |  |
|  | 364 (alone prev. 1) |  |  |
|  | 376(alone prev. 1) |  |  |
|  | $\begin{aligned} & 378 \text { (alone prev. 1) } \\ & * 410(?) \end{aligned}$ |  |  |
|  | *411(?) |  |  |

* Not included in calculations of mean age at first reproduction as possible earlier litter could easily have been missed.

Table 61. Black bear hunter kills in the Su-hydro study area.

| Year | Males | Females | Sex Unk. | Total | \% in Spring |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1973 | 14 | 6 | 2 | 22 | 0 |
| 1974 | 2 |  |  | 2 | 0 |
| 1975 | 6 | 2 | 2 | 10 | 0 |
| 1976 | 4 | 4 | 1 | 9 | 11 |
| 1977 | 1 |  | 1 | 2 | 50 |
| 1978 | 10 |  | . | 10 | 0 |
| 1979 | 8 | 4 |  | 12 | 17 |
| 1980 | . 14 | 9 | 1 | 24 | 13 |
| 1981 | 10 | 4 | 2 | 16 | 31 |
| 1982 | 9 | 5 |  | 14 | 29 |
| 1983 | 5 | 5 |  | 10 | 20 |
| 1984 | 11 | 5 |  | 16 | 38 |
| 1985 | 11 | 5 | 1 | 17 | 29 |

Table 62. Status of black bears first marked during Su-Hydro studies, 1980-1985. (A=alive, ND=no data available, F=shot in fall season, SP=shot in spring season). ND in year of capture indicates bear was not collared or soon shed its collar and no subsequent data were collected

|  | Bear ID | Sex/age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1980 Captures |  |  |  |  |  |  |  |  |
|  | 287 | M/10 in '80 | A | A | Shot-F | - | - | - | - |
|  | 288 | F/10 in ' 80 | Shed/dead? | ND | ND | ND | ND | ND | ND |
|  | 289 | F/9 in 180 | A | A | A | A | A | A | A |
|  | 290 | F/8 in ' 80 | Removed-F | ND | ND | ND | ND | ND | ND |
|  | 301 | F/7 $/ 7{ }^{\text {in }} 80$ | A | A | A | A | Shot-F | - | - |
|  | 302 | M/8 in ${ }^{1} 80$ | A | A | A | A | A | A | ND |
|  | 303 | M/8 in ' 80 | A | A | A | Shot-F | - | - | - |
|  | 304 | M/10 in '80 | A | A | Shed | ND | ND | ND | ND |
|  | 305 | M/9 in ${ }^{1} 80$ | Shot-F | - | - | - | , | - | ND |
|  | 307 | M/2 in 180 | A | Shot-Sp | - | - | - | - | - |
|  | 310 | M/2 in ${ }^{1} 80$ | A | A | A | A | A | A* | A |
|  | 316 | M/2 in ${ }^{1} 80$ | Shot-Sp | - | - | - | - | - | - |
|  | 317 | F/7 in ${ }^{1} 80$ | A | A | A | A | A | A | A |
|  | 318 | F/5 in ${ }^{1} 80$ | A | A | A | A | $A^{*}$ | A* | ND |
|  | 319 | M/3 $\mathbf{1 2} 180$ | A | Died-F | - | , | - | - | - |
|  | 320 |  | Shot-F | - | - | - | - | - | - |
|  | 321 | F/10 in ${ }^{1} 80$ | A | A | A | A | A | A | A |
|  | 322 | M/4 in ' 80 | A | A | Died-Sum | - | - | - | - |
|  | 323 | M/2 in ${ }^{\prime} 80$ | A | A | A | Shot-F | - | - | - |
|  | 324 | M/5 in '80 | A | A | $\mathrm{A}^{5}$ | A | Shot-F | - | - |
| $\mapsto$ | 325 | F/11 in ' 80 | A | Shed in den | ND | ND | ND | ND | ND |
| 6 | 326 | F/5 in ' 80 | Shot-F | - |  | - | , | , | - |
| $\omega$ | $327$ | $\text { F/5 in } 180$ |  |  | A | Died-Sum |  |  | - |
|  | 328 | F/6 in ' 80 | A | A | A | A | A | A | A |
|  | 1981 Captures |  |  |  |  |  |  |  |  |
|  | 329 | F/1 Ln '81 | - | Y1g | A | A | A | A | A |
|  | 330 | M/1 in ' 81 | - | Ylg, died-Sum | - | - | - | - | - |
|  | $342$ | M/5 in 181 | - | Shot-F | $\bar{\square}$ | $\bar{\square}$ | - | - | - |
|  | 343** | M/5 in '81 | - | A | A | A | Died-F | - | - |
|  | $346$ | M/9 in '81 | - | A | $\mathrm{A}$ | A | D1ed-Sp | - | - |
|  | $348$ | M/9 in '81 | - | A | Shot-F | - | - | $\overline{-}$ | - |
|  | 349 | F/4 in ' 81 | - | A | A | A | A | A | A |
|  | 1982 Captures |  |  |  |  |  |  |  |  |
|  | 350 | M/1 in ${ }^{\prime} 82$ | - | - | Ylg | - | - | - | - |
|  | 351 | M/1 in 182 | - | - | Y1g | A | A | $A^{*}-S p$ | ND |
|  | 354 | F/5 in ${ }^{\prime} 82$ | - | - | A | A | A | $\AA$ | A |
|  | 357 | M/4 in ${ }^{\prime} 82$ | - | - | Died winter | - | - | - | - |
|  | 358 | F/2 in 182 | - | - | A | A | Died-F | - | - |
|  | 359 | M/4 in ${ }^{\prime} 82$ | - | - | A | A | A | A | A |
|  | 360 | M/7 in 182 | - | - | A | A | Shed-Sp | ND | ND |
|  | 361 | F/7 in '82 | - | - | A | A | A | A | A |
|  | 362 | F/2 in '82 | - | - | $A-S p$ | ND | ND | ND | ND |
|  | $363$ | F/4 in '82 | - | - | A | A | A | A | A |
|  | $364$ | F/9 in 182 | - | - | A | A | A | A | A |
|  | 365** | M/S in ' 82 | - | - | A | Died-F | - | - | - |

continued on next page)

Table 62. (cont'd)

|  | Bear ID | Sex/age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1982 Captures | (cont 'd) |  |  |  |  |  |  |
|  | 366** | M/6 in ' 82 | - | Shot-F | - | - | - | - |
|  | 367** | F/4 in ' 82 | - | A | Shot-Sum | - | - | - |
|  | 369** | F/4 in '82 | - | A | A | A | A | A |
|  | 370** | F/7 in ' 82 | - | A | 3Shot?-Sp | - | - | - |
|  | 372** | F/9 in '82 | - | A | ?Shot?-F | - | - | - |
|  | 374** | F/7 in '82 | - | A | Shot-F | - | - | - |
|  | 375** | F/9 in '82 | - | A | A | A | Shot-F | - |
|  | 376** | F/6 in '82 | - | A | A | A | A | A |
|  | 377** | F/4 in '82 | - | A | A | A | A | A |
|  | 378* | F/6 in '82 | - | A | A | A | A | A |
|  | 1983 Captures |  |  |  |  |  |  |  |
|  | 379 | F/9 in '83 | - | - | Died-Sum | - | - | - |
|  | 387 | M/4 in '83 | - | - | A | A | Shot-F | - |
|  | 401 | M/3 in ' 83 | - | - | A | A | A | Shot?-Sp |
|  | 402** | F/10 in '83 | - | - | A | A | A | A |
|  | 404** | F/11 in '83 | - | - | A | ${ }^{\text {A }}$ | Shot?-Sp | - |
|  | 405** | F/17 in '83 | - | - | A | A | Shot-F | - |
|  | 406** | F/11 in '83 | - | - | A | A | ND | ND |
|  | 408** | M/3 in ${ }^{\text {c }} 83$ | - | - | A | A | A | ND |
|  | 409** | F/5 in '83 | - | - | A | A | A | A |
| $\stackrel{\square}{\circ}$ | 410** | F/7 in '83 | - | - | Shot-Sum | - | - | - |
| $\stackrel{\square}{\bullet}$ | 411** | F/8 in '83 | - | - | A | A | A | A |
|  | 1984 Captures |  |  |  |  |  |  |  |
|  | 412 | M/1 in '84 | - | - | - | Y1g w/361 | ND-Weaned | ND |
|  | 413 | F/l in ' 84 | - | - | - | Y1g w/361 | ND-Weaned | ND |
|  | 414 | F/1 in '84 | - | - - | - | Y1g w/361 | ND-Weaned | ND |
|  | 415 | F/1 in '84 | - | - | - | Ylg w/289-ND | - | - |
|  | 416 | M/9 in '84 | - | - | - | A | A | ND |
|  | 1985 Captures |  |  |  |  |  |  |  |
|  | 428 | M/5 in ' 85 | - | - | - | - | A | A |
|  | 430 | M/9 in ' 85 | - | - | - | - | A | ND |
|  | 431 | F/11 in ' 85 | - | - | - | - | A | A |
|  | 432 | F/6 in '85 | - | - | - | - | A | A |
|  | 434 | F/1 in '85 | - | - | - | - | Ylg w/432-W | ND |
|  | 433 | M/3 in '85 | - | - | - | - | A | ND |
|  | 435 | M/7 in '85 | - | - | - | - | Shot-F | - |
|  | 436 | M/2 in '85 | - | - | - | - | ND w/436-W | ND |
|  | 438 | F/8 in '85 | - | - | - | - | A | Shot-F |
|  | 441 | F/9 in '85 | - | - | - | - | A | A |
|  | 444 | M/3 in '85 | - | - | - | - | A | ND |
|  | 445 | M/8 in '85 | - | - | - | - | A | ND |
|  | 448 | F/6 in '85 | - | - | - | - | A | A |
|  | 449 | M/6 in ' 85 | - | - | - | - | A | ND |
|  | 451 | F/2 in ' 85 | - | - | - | - | A | ND |

Table 62. (continued)

|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 prelim. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| A. Max. no. marked bears potentially alive in year, includes ND. Excludes tagging and natural spring mortalities and coy and ylgs ( $\mathrm{M}: \mathrm{F}$ ) | 24(12:12) | 25(14:11) | $43(16: 27)$ | 50(15:35) | 41(13:28) | $53(20: 33)$ | 48(17:31) |
| B. No. KNOWN shot <br> Ln year ( $M$ : $F$ ) | 4(2:2) | 2(2:0) | 3(3:0) | $5(2 ; 3)$ | 2(1:1) | $5(2: 3)$ | $1(0: 1)$ |
| Min: \% known shot (Row B/Row A) | 17\% | 8.0\% | 7.0\% | 10.0\% | 4.9\% | 9.4\% | 2.1\% |
| C. No. known shot plus suspected (unreported) shot in year (M:F) | 4(2:2) | $2(2: 0)$ | $3(3: 0)$ | $7(2: 5)$ | 2 (1:1) | 5(2:3) | 2(1:1) |
| Probable min. \% shot (Row C/Row A) | 17\% | 8.0\% | 7.0\% | 14.0\% | 4.9\% | 9.4\% | 4.2\% |
| D. No, bears known alive (excludes ND, died, lost, cubs or ylgs) | 24 (12:12) | $24(14: 10)$ | 40(16:24) | 45(14:31) | $35(11: 24)$ | 45(16:29) | 26(4:22) |
| ${ }^{\circ}$ Probable \% shot (Row C/Row D) | 17\% | 8\% | 7.5\% | 15.6\% | 5.7\% | 11.1\% | 7.7\% |
| Cumulative \% shot (based on bear-years available, <br> from Row $A$ and Row C). | 17\% | 12,5\% | 9.8\% | 11.3\% | 9.8\% | 9.7\% | 8.8\% |

Not included: in 1980:
291 (M@3), 296 (M@10), 300 (M@7)
in 1982:
352 (M@2), 353 (M@1), $368^{* *}(F @ 3), 371$ (M@2), 2 coy $W / B 354$
in 1983:
3 coy w/B374, I coy w/B404, 2 coy w/B405
in 1984:
2 Coy w/B369
in 1985:
426(M@2), 439 (M@2 w/B438-hurt leg), B446(F@5), 2 coy w/B349,
3 coy w/B328, 3 coy w/B404

* Previous alive status based in part at least, on knowledge from this year.
** Bear in downstream study area.

Table 63. Status of black bears marked during Su-Hydro studies, 1980-1983. (A=alive, ND=no data, F=shot in fall season, Sp=shot in spring season, S=Summer capture or mortallty).

|  | Bear ID | Sex/Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Upstream Study Area |  |  |  |  |  |  |  |  |
|  | 287 | M/10 in ' 80 | A | A | Shot-F | - | - | - | - |
|  | 288 | F/10 in ' 80 | A (shed) | ND | ND | ND | ND | ND | ND |
|  | 289 | F/9 in ' 80 | A | A | A | A | A | A | A |
|  | 290 | F/8 in ' 80 | A | A (remvd) | ND | ND | ND | ND | ND |
|  | 301 | F/7 in '80 | A | A | A | A (shed) | Shot-F | - | - |
|  | 302 | M/8 in '80 | A | A | A | A | A | A | ND |
|  | 303 | M/8 in ' 80 | A | A | A | Shot-F | - | - | - |
|  | 304 | M/10 in '80 | A | A | A (shed) | ND | ND | ND | ND |
|  | 305 | M/9 in '80 | Shot-F | - | - | - | - | - | - |
|  | 307 | M/2 in '80 | A | Shot-Sp. | - | - | - | - | - |
|  | 310 | M/2 in '80 | A | A | A | A | A | A* | A |
|  | 316 | F/12 in ' 80 | Shot-F | - | - | - | - | - | - |
|  | 317 | F/7 in '80 | A-S | A | A | A | A | A | A |
| $\stackrel{\rightharpoonup}{\bullet}$ | 318 | F/5 in '80 | A-S | A | A | A* | A* | A | ND |
| $\bigcirc$ | 319 | M/3 in '80 | A-S | Died | - | - | - | - | - |
|  | 320 | M/4 in '80 | Shot-F | - | - | - | - | - | - |
|  | 321 | F/70 in '80 | A-S | A cubs | A | A | A | A | A |
|  | 322 | M/4 in '80 | A-S | A | Died | - | - | - | - |
|  | 323 | M/2 in '80 | A-S | A | A | Shot-F | - | - | - |
|  | 324 | M/5 in ' 80 | A-S | A | A | A | Shot-F | - | - |
|  | 325 | F/11 in '80 | A-S | A | Shed | ND | ND | ND | ND |
|  | 326 | F/5 in '80 | Shot-F | - | - | - | - | - | - |
|  | 327 | F/5 in ' 80 | A-S | A | A | Died-S | - | - | - |
|  | 328 | F/6 in '80 | A-S | A | A | A | A | A | A |
|  | 329 | F/1 in '81 | - | Ylg. | A | A | A | A | A |
|  | 330 | M/1 in '80 | - | Ylg. died-S | - | - | - | - | - |
|  | 342b | M/5 in '81 | - | Shot-F | - | - | - | - | - |
|  | 346 | M/9 in ' 81 | - | A | A. | A | Died | - | - |
|  | 348 | M/9 in '81 | - | A-S | Shot-F | - | - | - | - |
|  | 349 | F/4 in '81 | - | A-S | A | A | A | A | A |
|  | 350 | M/1 in '82 | - | - | Ylg. | - | - | - | - |
|  | 351 | M/1 in '82 | - | - | Ylg. | A | A | A* | ND |
|  | 354 | F/5 in '82 | - | - | A | A | A | A | A |
|  | 357 | $\mathrm{M} / 4$ in '82 | - | - | Died-W | - | - | - | - |
|  | 358 | M/2 in ' 82 | - | - | A | A | Died-F | - | - |
|  | 359 | M/4 in ' 82 | - | - | A | A | A | A | A |

(continued on next page)

Table 63. (cont'd)

| Bear ID | Sex/Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Upstream Study Area (cont'd) |  |  |  |  |  |  |  |
| 360 | M/7 in ' 82 | - | A | A | A | ND | ND |
| 361 | F/7 in ' 82 | - | A | A | A | A | A |
| 362 | F/2 in ' 82 | - | A-Sp. | ND | ND | ND | ND |
| 363 | F/4 in ' 82 | - | A | A | A | A | A |
| 364 | F/9 in '82 | - | A | A | A | A | A |
| 379 | F/9 in '83 | - | - | Died-S | - | - | - |
| 387 | F/4 in '83 | - | - | A | A | Shot-F | - |
| 401 | M/3 in '83 | - | - | A | A | A | Shot?-Sp. |
| 412 | M/1 in '84 | - | - | - | Y1g. | A | ND |
| 413 | F/l in '84 | - | - | - | Y1g. | A | ND |
| 414 | F/1 in '84 | - | - | - | Y1g. | A | ND |
| 416 | M/9 in ' 84 | - | - | - | A | A | A |
| 428 | M/5 in '85 | - | - | - | - | A | A |
| 430 | M/9 in ' 85 | - | - | - | - | A | ND |
| 431 | F/ll in ' 85 | - | - | - | - | A | A |
| 432 | F/6 in ' 85 | - | - | - | - | A | A |
| 433 | M/3 in ' 85 | - | - | - | - | A | ND |
| 434 | F/1 in ' 85 | - | - | - | - | Y1g. | - |
| 435 | M/7 in '85 | - | - | - | - | Shot-F | - |
| 436 | M/2 in ' 85 | - | - | - | - | ND | ND |
| 438 | F/8 in ' 85 | - | - | - | - | A | Shot-F |
| 441 | F/9 in ' 85 | - | - | - | - | A | A |
| 444 | M/3 in ' 85 | - | - | - | - | A | ND |
| 445 | M/8 in ' 85 | - | - | - | - | A | ND |
| 448 | F/6 in '85 | - | - | - | - | A | A |
| 449 | M/6 in '85 | - | - | - | - | A | ND |
| 451 | F/2 in ' 85 | - | - | - | - | A | ND |

(continued on next page)

|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Upstream subtotals |  |  |  |  |  |  |  |
| Maximum no. bears potentially alive (includes ND) in year (excludes natural mortalities) (M;F) | 24(12:12) | 24 (13:11) | $31(14: 17)$ | 31 (12:19) | $28(11: 17)$ | 41(17:24) | 39(16:21) |
| No. known shot (M:F) | 4(2:2) | 2(2:0) | 2(2:0) | 2(2:0) | 2(1:1) | 2(1:1) | 2(1:1) |
| No, additional bears suspected shot | 0 | . 0 | 0 | 0 | 0 | 0 | 0 |
| \% known or suspected shot | 178 | $8 \%$ | 7\% | 7\% | 7\% | 5\% | 5\% |

Table 63. (cont'd)

|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Downstream Study Area |  |  |  |  |  |  |
| 343 M/5 in ' 81 | A | A | A | Died-F | - | - |
| 365 M/5 in! 82 | - | A | Died-F | - | - | - |
| 366 M/6 in '82 | - | Shot-F | - | - | - | - |
| 367 F/4 in 'g2 | - | A | Shot-Sum. | - | - | - |
| $369 \quad F / 4$ in '82 | - - | A | A | A | A | A |
| $370 \quad F / 7$ in ' 82 | - | A | (Shot?) -S | - | - | - |
| 372 F/9 in '82 | - | A | (Shot?) -F | - | - | - |
| 374 F/7 in '82 | - | A | Shot-F | - | - | - |
| $375 \quad F / 5$ in ' 82 | - | A | A | A | Shot-F | - |
| 376 F/6 in '82 | - | A | A | A | A | A |
| 377 F/5 in ' 82 | - | A | A | A | A | A |
| 378 F/6 in '82 | - | A | A | A | A | A |
| $402 \mathrm{~F} / 10$ in '83 | - | - | A | A | A | A |
| $404 \quad \mathrm{~F} / 11$ in ' 83 | - | - | A | A | Shot?-Sp. | - |
| $405 \quad \mathrm{~F} / 17$ in '83 | - | - | A | A | Shot-F | - |
| 406 F/ll in '83 | - | - | A | A | ND | ND |
| 408 M/3 in ' 83 | - | - | A | A | A | ND |
| $409 \quad \mathrm{~F} / 5 \mathrm{in}$ ' 83 | - | - | A | A | A | A |
| $410 \quad \mathrm{~F} / 7$ in ' 83 | - | - | Shot-S | - | - | - |
| 411 F/8 in ' 83 | - | - | A | A | A | A |
| Downstream subtotals |  |  |  |  |  |  |
| Max. no. bears potentially <br> alive (includes ND) in year <br> (excludes natural mortalities) (M:F) | 1(1:0) | 12(3:9) | 19(3:16) | 13(2:11) | 12(1:11) | 9(1:10) |
| No. known shot (M:F) | 0 | 1(1:0) | $3(0: 3)$ | 0 | 2(0:2) | 0 |
| No. additional bears suspected shot (M:F) | 0 | 0 | 2(0:2) | 0 | $1(0: 1)$ | 0 |
| \% known or suspected shot (M;F) | - | $8 \%$ | 26\% | 0 | 25\% | 0 |

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 updated 10/86Table 63. (cont'd)

|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Upstream subtotals
Upstream and Downstream Areas Combined

Total bears potentially alive in year (excludes natural mortalities,

| includes ND) (M:F) | 24(12:12) | 25 (14:11) | 43(17:26) | 50(15:35) | 41(13:28) | 53(18:35) | 48(17:31) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. known shot (M:F) | 4(2:2) | 2(2:0) | $3(3: 0)$ | 5(2:3) | $2(1: 1)$ | 4(1:3) | 2(1:1) |
| No. additional bears suspected shot (M:F) | 0 | 0 | 0 | 2(0:2) | 0 | 1(0:1) | 0 |
| \% known or suspected shot | 17\% | 8\% | 7\% | 14\% | 5\% | 9\% | 4\% |

* Based on information obtained after this year.

Table 64. Black bear home range size. Code 99 in year or age column indicates lumping of all years. Area $1=$ upstream area, area $2=$ downstream study areas; sex $1=$ male, and $2=$ female; $0=$ w/o cubs-of-the-year and $1=$ with COY.

|  | ID |  |  |  | Age No. Points <br> (yrs.) Locations |  | Size |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | No. | Area | Sex | Year |  |  | Sq. Km. | Period | Comments | COY |
|  | 287 | 1 | 1 | 80 | 10 | 17 | 136.3 | May-0ct | w/o atypical den | 0 |
|  | 287 | 1 | 1 | 81 | 11 | 15 | 268.2 | Apr-Oct | w/o atypical den | 0 |
|  | 287 | 1 | 1 | 82 | 12 | 18 | 250.0 | Apr-Sept | shot 9/82 | 0 |
|  | 287 | 1 | 1 | 99 | 99 | 50 | 313.7 | 1980-82 |  | 0 |
|  | 302 | 1 | 1 | 81 | 9 | 36 | 325.7 | Apr-0ct | captured 5/80 | 0 |
|  | 302 | 1 | 1 | 82 | 10 | 11 | 51.1 | Apr-Jul | missing 7/82 | 0 |
|  | 302 | 1 | 1 | 84 | 11 | 42 | 351.6 | May-Aug | recaptured | 0 |
|  | 302 | 1 | 1 | 99 | 99 | 03 | 498.3 | 1980-85 |  | 0 |
|  | 303 | 1 | 1 | 80 | 8 | 15 | 94.9 | May-0ct |  | 0 |
|  | 303 | 1 | 1 | 81 | 9 | 18 | 92.5 | Apr-0ct |  | 0 |
|  | 303 | 1 | 1 | 82 | 10 | 20 | 73.6 | Apr-0ct |  | 0 |
| N | 303 | 1 | 1 | 83 | 11 | 11 | 43.2 | Apr-Sept | shot 9/83 | 0 |
| - | 303 | 1 | 1 | 99 | 99 | 64 | 167.0 | 1980-83 |  | 0 |
|  | 304 | 1 | 1 | 80 | 10 | 15 | 35.1 | May-Sept | w/o atypical den | 0 |
|  | 304 | 1 | 1 | 81 | 11 | 18 | 40.8 | Apr-0ct | shed 7/82 | 0 |
|  | 304 | 1 | 1 | 99 | 99 | 39 | 138.7 | 1980-82 | shed 7/82 | 0 |
|  | 305 | 1 | 1 | 80 | 9 | 9 | 47.9 | May-Aug | shot 8/80 | 0 |
|  | 305 | 1 | 1 | 99 | 9 | 9 | 47.9 | 1980 |  | 0 |
|  | 319 | 1 | 1 | 81 | 4 | 10 | 43.1 | Apr-July | captured $8 / 80$ | 0 |
|  | 319 | 1 | 1 | 99 | 99 | 16 | 455.8 | 1980-1981 | died 7/81 | 0 |
|  | 322 | 1 | 1 | 99 | 99 | 12 | 48.5 | 1980-82 | shed=2, died 7/82 | 0 |
|  | 323 | 1 | 1 | 81 | 3 | 19 | 382.9 | Apr-0ct | captured 8/80 | 0 |
|  | 323 | 1 | 1 | 82 | 4 | 20 | 1126.0 | Apr-0ct |  | 0 |
|  | 323 | 1 | 1 | 83 | 5 | 17 | 1089.3 | Apr-Sept | shot 9/83 | 0 |
|  | 323 | 1 | 1 | 99 | 99 | 62 | 1514.3 | 1980-83t |  | 0 |
|  | 324 | 1 | 1 | 81 | 6 | 20 | 247.8 | Apr-0ct | captured 8/80 | 0 |
|  | 324 | 1 | 1 | 82 | 7 | 21 | 139.9 | Apr-0ct |  | 0 |
|  | 324 | 1 | 1 | 83 | 8 | 17 | 170.2 | Apr-0ct |  | 0 |
|  | 324 | 1 | 1 | 84 | 9 | 11 | 236.8 | Apr-Sept | shot 9/84 | 0 |
|  | 324 | 1 | 1 | 99 | 99 | 75 | 776.5 | 1980-1984 |  | 0 |
|  | 330 | 1 | 1 | 81 | 1 | 14 | 10.0 | May-0ct | died 7/81 | 0 |
|  | 330 | 1 | 1 | 99 | 99 | 14 | 10.0 | 1981,82 |  | 0 |
|  | 346 | 1 | 1 | 81 | 9 | 16 | 61.5 | May-0ct |  | 0 |

[^9]Table 64. (continued)

| $\begin{aligned} & \overline{\mathrm{ID}} \\ & \text { No. } \end{aligned}$ | Area | Sex | Year | $\begin{gathered} \text { Age } \\ \text { (yrs.) } \end{gathered}$ | No. Points <br> Locations | $\begin{gathered} \text { Size } \\ \text { Sq. Km. } \end{gathered}$ | Period | Comments | COY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 346 | 1 | 1 | 82 | 10 | 22 | 90.5 | Apr-0ct |  | 0 |
| 346 | 1 | 1 | 83 | 11 | 16 | 119.0 | Apr-0ct |  | 0 |
| 346 | 1 | 1 | 99 | 99 | 56 | 175.0 | 1981-1983 | died 6/84 | 0 |
| 348 | 1 | 1 | 82 | 10 | 9 | 135.5 | 8/81-9/82 | shot 9/82 | 0 |
| 348 | 1 | 1 | 99 | 99 | 16 | 522.4 | 1981-82 | 8/81-shot 9/82 | 0 |
| 357 | 1 | 1 | 82 | 4 | 18 | 11.2 | May-Oct | died 10/82 | 0 |
| 357 | 1 | 1 | 99 | 4 | 18 | 11.2 | 1982 |  | 0 |
| 359 | 1 | 1 | 82 | 4 | 18 | 83.2 | May-Oct |  | 0 |
| 359 | 1 | 1 | 83 | 5 | 19 | 154.2 | Apr-Oct |  | 0 |
| 359 | 1 | 1 | 84 | 6 | 59 | 455.5 | Apr-0ct |  | 0 |
| 359 | 1 | 1 | 99 | 99 | 05 | 698.8 | 1982-85 ${ }^{\text {, }}$ |  | 0 |
| 387 | 1 | 1 | 83 | 4 | 16 | . 163.8 | May-0ct |  | 0 |
| 387 | 1 | 1 | 84 | 5 | 53 | 152.5 | Apr-Oct | shot 9/85 | 0 |
| 387 | 1 | 1 | 99 | 99 | 75 | 207.7 | 1983-85 |  | 0 |
| 401 | 1 | 1 | 83 | 3 | 18 | 91.4 | May-Oct |  | 0 |
| 401 | 1 | 1 | 84 | 4 | 56 | 240.5 | Apr-0ct |  | 0 |
| 401 | 1 | 1 | 99 | 99 | 86 | 241.0 | 1983-85 |  | 0 |
| 416 | 1 | 1 | 84 | 9 | 55 | 377.6 | May-Oct |  | 0 |
| 416 | 1 | 1 | 99 | 99 | 61 | 377.6 | 1984-85 |  | 0 |
| 342b | 1 | 1 | 81 | 5 | 40 | 611.3 | May-Sept | shot 9/81 | 0 |
| 342b | 1 | 1 | 99 | 5 | 40 | 611.3 | 1981 | shot 9/81 | 0 |
| 343 | 2 | 1 | 81 | 5 | 16 | 288.7 | May-Oct |  | 0 |
| 343 | 2 | 1 | 82 | 6. | 19 | 369.5 | Apr-Oct | . | 0 |
| 343 | 2 | 1 | 83 | 7 | 20 | 500.6 | Apr-Oct |  | 0 |
| 343 | 2 | 1 | 84 | 8 | 14 | 653.8 | Apr-Oct | died 12/84 | 0 |
| 343 | 2 | 1 | 99 | 99 | 69 | 1000.2 | 1981-85 |  | 0 |
| 365 | 2 | 1 | 82 | 5 | 11 | 656.4 | May-Oct |  | 0 |
| 365 | 2 | 1 | 83 | 6 | 15 | 251.6 | Apr-Sept | died 9/83 | 0 |
| 365 | 2 | 1 | 99 | 99 | 26 | 711.7 | 1982-83 |  | 0 |
| 366 | 2 | 1 | 82 | 6 | 10 | 136.1 | May-Aug | shot 9/82 | 0 |
| 366 | 2 | 1 | 99 | 6 | 10 | 136.1 | 1982 |  | 0 |
| 408 | 2 | 1 | 83 | 3 | 16 | 226.6 | May-Oct |  | 0 |
| 408 | 2 | 1 | 84 | 4 | 11 | 230.5 | Apr-Oct |  | 0 |
| 408 | 2 | 1 | 99 | 99 | 30 | 652.7 | 1983-85 |  | 0 |

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Table 64. (continued)

|  | ID |  |  |  | Age | No. Points | Size |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | No. | Area | Sex | Year | (yrs.) | Locations | Sq. Km. | Period | Comments | COY |
|  | 288 | 1 | 2 | 80 | 10 | 16 | 7.4 | May-Aug | shed 8/80 | 0 |
|  | 288 | 1 | 2 | 99 | 10 | 16 | 7.4 | 1980 |  | 0 |
|  | 289 | 1 | 2 | 80 | 9 | 14 | 43.4 | May-Oct |  | 0 |
|  | 289 | 1 | 2 | 81 | 10 | 20 | 26.1 | Apr-0ct | w/coy, survived | 1 |
|  | 289 | 1 | 2 | 82 | 11 | 20 | 29.0 | Apr-0ct |  | 0 |
|  | 289 | 1 | 2 | 83 | 12 | 17 | 18.6 | Apr-0ct | w/coy, survived | 1 |
|  | 289 | 1 | 2 | 84 | 13 | 62 | 52.7 | Apr-Oct ${ }^{\text { }}$ | w/@1 | 0 |
|  | 289 | 1 | 2 | 99 | 99 | 42 | 79.8 | 1980-85 | coy in $\pm 85$ | 0 |
|  | 290 | 1 | 2 | 80 | 8 | 18 | 44.7 | May-0ct |  | 0 |
|  | 290 | 1 | 2 | 81 | 9 | 15 | 116.3 | Apr-Aug | collar removed 8/81 | 0 |
|  | 290 | 1 | 2 | 99 | 99 | 33 | 163.4 | 1980-81 |  | 0 |
|  | 301 | 1 | 2 | 80 | 7 | 20 | 18.1 | May-0ct |  | 0 |
|  | 301 | 1 | 2 | 81 | 8 | 15 | 12.5 | Apr-Oct | w/coy, survived | 1 |
|  | 301 | 1 | 2 | 82 | 9 | 18 | 18.3 | Apr-Oct | shed 7/83 | 0 |
|  | 301 | 1 | 2 | 99 | 99 | 62 | 29.6 | 1980-83 |  | 0 |
| - | 317 | 1 | 2 | 81 | 8 | 19 | 13.9 | May-Oct | captured 8/80 | 0 |
| $\omega$ | 317 | 1 | 2 | 82 | 9 | 18 | 44.2 | Apr-Oct |  | 0 |
|  | 317 | 1 | 2 | 83 | 10 | 19 | 16.8 | Apr-Oct | w/coy, survived | 1 |
|  | 317 | 1 | 2 | 84 | 11 | 58 | 40.4 | Apr-0ct | w/@1 | 0 |
|  | 317 | 1 | 2 | 99 | 99 | 30 | 59.0 | 1980-85 |  | 0 |
|  | 318 | 1 | 2 | 81 | 6 | 20 | 1036.4 | Apr-Oct | captured 8/80 | 0 |
|  | 318 | 1 | 2 | 82 | 7 | 20 | 471.9 | Apr-Oct | shed 7/83 | 0 |
|  | 318 | 1 | 2 | 99 | 99 | 58 | 1095.4 | 1980-83, 85 | recaptured 6/85 | 0 |
|  | 321 | 1 | 2 | 81 | 11 | 14 | 771.0 | Apr-0ct | w/coy, lost coy 7/81 | 1 |
|  | 321 | 1 | 2 | 82 | 12 | 20 | 13.8 | Apr-Oct | prev. lost 8/81 | 0 |
|  | 321 | 1 | 2 | 83 | 13 | 18 | 28.5 | Apr-Oct |  | 0 |
|  | 321 | 1 | 2 | 84 | 14 | 17 | 14.5 | Apr-Oct | w/coy, survived | 1 |
|  | 321 | 1 | 2 | 99 | 99 | 81 | 836.0 | 1980-85 |  | 0 |
|  | 325 | 1 | 2 | 99 | 99 | 15 | 145.7 | 1980-81 | fall data only | 0 |
|  | 327 | 1 | 2 | 81 | 6 | 35 | 31.3 | Apr-0ct | captured 7/80 | 0 |
|  | 327 | 1 | 2 | 82 | 7 | 19 | 34.2 | Apr-Oct |  | 0 |
|  | 327 | 1 | 2 | 99 | 99 | 69 | 51.5 | 1980-83 | died 7/83 | 0 |
|  | 328 | 1 | 2 | 81 | 7 | 19 | 28.0 | Apr-Oct | w/coy, captured 8/80 | 1 |
|  | 328 | 1 | 2 | 84 | 10 | 56 | 31.7 | May-Oct | recaptured, shed 1981 | 0 |
|  | 328 | 1 | 2 | 99 | 99 | 89 | 64.4 | 1980-85 | lost coy in 1985 | 0 |

Table 64. (continued)

|  | $\overline{\bar{I}}$ |  |  |  | Age | No. Points | Size |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | No. | Area | Sex | Year | (yrs.) | Locations | Sq. Km. | Period | Comments | COY |
|  | 329 | 1 | 2 | 81 | 1 | 19 | 14.7 | May-Oct |  | 0 |
|  | 329 | 1 | 2 | 82 | 2 | 19 | 9.4 | Apr-Oct |  | 0 |
|  | 329 | 1 | 2 | 83 | 3 | 18 | 24.1 | Apr-Oct |  | 0 |
|  | 329 | 1 | 2 | 84 | 4 | 62 | 36.0 | Apr-Oct |  | 0 |
|  | 329 | 1 | 2 | 99 | 99 | 28 | 100.0 | 1981-85 | never had coy | 0 |
|  | 349 | 1 | 2 | 82 | 5 | 20 | 47.4 | Apr-Oct | captured 8/81 | 0 |
|  | 349 | 1 | 2 | 84 | 7 | 56 | 53.9 | May-Oct | recaptured, alone | 0 |
|  | 349 | 1 | 2 | 99 | 99 | 00 | 82.7 | 1981-85 | shed 7/83 | 0 |
|  | 354 | 1 | 2 | 82 | 5 | 19 | 64.8 | May-Oct | w/2@0 | 1 |
|  | 354 | 1 | 2 | 83 | 6 | 17 | 61.6 | Apr-Oct |  | 0 |
|  | 354 | 1 | 2 | 84 | 7 | 23 | 118.3 | Apr-Oct | w/coys, lost 6/84 | 0 |
|  | 354 | 1 | 2 | 99 | 99 | 63 | 140.9 | 1982-1985 |  | 0 |
|  | 358 | 1 | 2 | 82 | 2 | 17 | 10.7 | May-Oct |  | 0 |
|  | 358 | 1 | 2 | 83 | 3 | 17 | 53.2 | Apr-Oct |  | 0 |
|  | 358 | 1 | 2 | 84 | 4 | 43 | 57.5 | Apr-Aug | died 8/84 | 0 |
| N | 358 | 1 | 2 | 99 | 99 | 77 | 71.1 | 1982-84 |  | 0 |
| $\stackrel{\square}{*}$ | 360 | 1 | 2 | 82 | 7 | 20 | 144.5 | May-Oct |  | 0 |
|  | 360 | 1 | 2 | 83 | 8 | 19 | 299.2 | Apr-Oct |  | 0 |
|  | 360 | 1 | 2 | 99 | 99 | 42 | 429.1 | 1982-84 |  | 0 |
|  | 361 | 1 | 2 | 82 | 7 | 18 | 87.9 | May-Oct |  | 0 |
|  | 361 | 1 | 2 | 83 | 8 | 16 | 59.9 | Apr-Oct | w/coy, survived | 1 |
|  | 361 | 1 | 2 | 84 | 9 | 59 | 66.6 | Apr-Oct | w/@l all year | 0 |
|  | 361 | 1 | 2 | 99 | 99 | 07 | 111.3 | 1982-1985 |  | 0 |
|  | 363 | 1 | 2 | 82 | 3 | 18 | 19.9 | May-Oct |  | 0 |
|  | 363 | 1 | 2 | 83 | 4 | 18 | 20.6 | Apr-Oct |  | 0 |
|  | 363 | 1 | 2 | 84 | 5 | 23 | 19.6 | Apr-Oct | w/2@0, survived | 1 |
|  | 363 | 1 | 2 | 99 | 99 | 65 | 30.0 | 1982-85 | no coy in 85 or 86 | 0 |
|  | 364 | 1 | 2 | 82 | 9 | 16 | 121.5 | May-Sept | lost 9/82 | 0 |
|  | 364 | 1 | 2 | 99 | 9 | 16 | 121.5 | 1982 |  | 0 |
|  | 367 | 2 | 2 | 82 | 4 | 17 | 17.5 | May-Oct |  | 0 |
|  | 367 | 2 | 2 | 99 | 99 | 26 | 17.7 | 1982-83 | shot 7/83 | 0 |
|  | 369 | 2 | 2 | 82 | 4 | 19 | 10.2 | May-Oct |  | 0 |
|  | 369 | 2 | 2 | 83 | 5 | 20 | 26.0 | Apr-Oct |  | 0 |
|  | 369 | 2 | 2 | 84 | 6 | 12 | 20.0 | Apr-Oct | w/coy, survived | 1 |
|  | 369 | 2 | 2 | 99 | 99 | 59 | 30.9 | 1982-85 |  | 0 |

[^10]Table 64. (continued)

|  | $\begin{aligned} & \text { ID } \\ & \text { No. } \end{aligned}$ | Area | Sex | Year | $\begin{gathered} \text { Age } \\ \text { (yrs.) } \end{gathered}$ | No. Points Locations | $\begin{gathered} \text { Size } \\ \text { Sq. Km. } \end{gathered}$ | Period | Comments | COY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 370 | 2 | 2 | 82 | 7 | 18 | 16.0 | May-Oct | lost 5/83 | 0 |
|  | 370 | 2 | 2 | 99 | 99 | 22 | 16.0 | 1982-83 |  | 0 |
|  | 372 | 2 | 2 | 82 | 9 | 17 | 56:1 | May-Oct |  | 0 |
|  | 372 | 2 | 2 | 83 | 10 | 13 | 75.6 | Apr-Aug | w/coy, failed 9/83 | 1 |
|  | 372 | 2 | 2 | 99 | 99 | 30 | 101.6 | 1982-83 |  | 0 |
|  | 374 | 2 | 2 | 83 | 8 | 16 | 30.3 | Apr-Sept | shot 9/83 | 0 |
|  | 374 | 2 | 2 | 99 | 99 | 20 | 34.5 | 1982-9/83 |  | 0 |
|  | 375 | 2 | 2 | 82 | 9 | 16 | 16.8 | Jun-0ct |  | 0 |
|  | 375 | 2 | 2 | 83 | 10 | 19 | 19.3 | Apr-Oct | w/coy | 1 |
|  | 375 | 2 | 2 | 84 | 11 | 14 | 38.2 | Apr-0ct | shot 5/85 | 0 |
|  | 375 | 2 | 2 | 99 | 99 | 49 | 53.1 | 1982-85 |  | 0 |
|  | 376 | 2 | 2 | 82 | 6 | 13 | 21.1 | Jun-Oct |  | 0 |
|  | 376 | 2 | 2 | 83 | 7 | 21 | 34.0 | Apr-0.ct | w/coy | 1 |
|  | 376 | 2 | 2 | 84 | 8 | 14 | 36.3 | Apr-Oct | w/@1 | 0 |
|  | 376 | 2 | 2 | 99 | 99 | 56 | 108.5 | 1982-85 |  | 0 |
| O | 377 | 2 | 2 | 82 | 4 | 15 | 11.9 | Jun-0ct |  | 0 |
| or | 377 | 2 | 2 | 83 | 5 | 18 | 24.5 | Apr-0ct | w/coy, lost 5/83 | 0 |
|  | 377 | 2 | 2 | 84 | 6 | 12 | 13.2 | Apr-Oct |  | 0 |
|  | 377 | 2 | 2 | 99 | 99 | 52 | 81.5 | 1982-85 | coy in 85 | 0 |
|  | 378 | 2 | 2 | 82 | 6 | 14 | 8.0 | Jun-Oct |  | 0 |
|  | 378 | 2 | 2 | 83 | 7 | 20 | 9.8 | Apr-Oct | w/coy, survived | 1 |
|  | 378 | 2 | 2 | 84 | 8 | 12 | 7.3 | Apr-Oct |  | 0 |
|  | 378 | 2 | 2 | 99 | 99 | 52 | 17.5 | 1982-85 | coy in $\pm 85$ | 0 |
|  | 402 | 2 | 2 | 83 | 10 | 17 | 13.4 | May-Oct |  | 0 |
|  | 402 | 2 | 2 | 84 | 11 | 16 | 15.6 | Apr-Oct | alone | 0 |
|  | 402 | 2 | 2 | 99 | 99 | 40 | 26.8 | 1983-85 | coy in $\pm 85$ | 0 |
|  | 404 | 2 | 2 | 83 | 11 | 16 | 36.3 | May-Oct | coy | 1 |
|  | 404 | 2 | 2 | 84 | 12 | 13 | 89.5 | Apr-Oct | coy in ' 85 | 0 |
|  | 404 | 2 | 2 | 99 | 99 | 31 | 137.8 | .1983-85 | died 5/85 | 0 |
|  | 405 | 2 | 2 | 83 | 11 | 17 | 24.7 | May-Oct | w/coy | 1 |
|  | 405 | 2 | 2 | 84 | 12 | 11 | 53.6 | Apr-0ct | w/@1 | 0 |
|  | 405 | 2 | 2 | 99 | 99 | 32 | 70.7 | 1983-85 | coy in $\pm 85$ | 0 |
|  | 406 | 2 | 2 | 83 | 11 | 17 | 17.9 | May-Oct | w/coy, survived | 1 |
|  | 406 | 2 | 2 | 84 | 12 | 13 | 16.2 | Apr-Sept. | lost 9/84 | 0 |

Tabie 64. (continued)

| ID |  |  |  | Age | No. Points | Size |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. | Area | Sex | Year | (yrs.) | Locations | Sq. Km. | Period | Comments | COY |
| 406 | 2 | 2 | 99 | 99 | 30 | 20.7 | 1983-84 |  | 0 |
| 409 | 2 | 2 | 83 | 5 | 16 | 26.4 | May-0ct |  | 0 |
| 409 | 2 | 2 | 84 | 6 | 14 | 15.9 | Apr-0ct |  | 0 |
| 409 | 2 | 2 | 99 | 99 | 35 | 32.3 | 1983-85 |  | 0 |
| 411 | 2 | 2 | 83 | 8 | 17 | 31.3 | May-Oct |  | 0 |
| 411 | 2 | 2 | 84 | 9 | 12 | 45.7 | Apr-0ct | w/coy, survived | 1 |
| 411 | 2 | 2 | 99 | 99 | 36 | 105.5 | 1983-85 |  | 0 |

Table 65. Black bear home range slze by sex and age categories. (COY = cubs-of-year).

| No. $\begin{gathered}\text { Number of } \\ \text { radio-location points }\end{gathered}$ |  |  |  |  | Home Range Size ( $\mathrm{km}^{2}$ )* |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Category | Individuals | Mean | Max. | Min | Mean | S.D. | Max. | Min. |

TOTAL HOME RANGE (Summation all years)

| A11 bears | 55 | 52.7 | 142 | 9 | 250.7 | 324.8 | 1514.3 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| All males | 22 | 47.1 | 105 | 9 | 423.5 | 372.8 | 1514.3 |
| All females | 33 | 56.5 | 142 | 15 | 135.6 | 229.4 | 1095.7 |

ANNUAL HOME RANGES (all points in calendar year)

| All bears | 123 | 20.9 | 62 | 9 | 134.6 | 212.8 | 1126.0 | 7.3 |
| :--- | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| All males | 45 | 20.9 | 59 | 9 | 251.5 | 250.8 | 1126.0 | 10.0 |
| All females | 78 | 20.8 | 62 | 11 | 67.1 | 152.3 | 1036.4 | 7.3 |
| Females 5.0+, <br> without coy | 47 | 21.6 | 62 | 11 | 77.3 | 163.5 | 1036.4 | 7.3 |
| Females 5.0+, <br> with coy | 19 | 17.2 | 23 | 12 | 69.2 | 171.0 | 771.0 | 9.8 |

* Standard minimum grid method.
 based on status on 15 June. If another bear or wolves also on kill, each credited with 0.5 kills. Consecutive observation day sums all days, for periods of $>2$ consecutive days. Only spring data included, summer 1984 not included. Misc. kills include suspected and probable kills.


Table 67. Subjective characterization of berry abundance in the upstream study area since 1980.
Characterization of
Berry Abundance $\quad$ Comments
1980 normal

1981
very poor
slightly subaverage

No special effort was made to evaluate berry abundance, black bears were very common in the shrublands adjacent to forested habitats and in forested habitats.

Extensive unanticipated movements of radio-marked black bears in late summer provided first clue that something was amiss. On the ground inspection supported hypothesis that blueberries were very scarce. Bears were in very poor condition the following spring in both upstream and downstream area. Three marked black bears died (Table 34) in 1981 following the summer berry failure. Bears were common in semi-open shrublands.

Berry transects supported hypothesis that berries were more abundant in shrublands than in adjacent forests. Low reproductive success evident in spring 1982 and bears tended to be very skinny. In summer bears foraged in shrublands but there appeared to be many fewer bears in the study area than in 1980. Would have concluded a massive emigration in 1981 except that the marked bears that moved away had all returned. Possibly there was an increased mortality rate resulting from the 1981 berry failure. One marked bear died in 1982 compared to 3 in the previous and following years. Mortality could have been most marked on subadults, only 2 of these were radio-marked.

Table 67. (cont'd)
Characterization of
Year
Berry Abundance
Comments
1983 above average

1984
below average

Berry transects suggest more berries than in 1982, especially crowberries and lowbush cranberries. Although not evident in the transect data, it appeared that blueberries were locally very abundant in forested habitats and bears did not have to, and didn't, move into the shrubland habitat types to forage for berries in late summer. Some black bears expected to produce their first litters in 1983 failed to do so suggesting delayed age of first reproduction may have resulted from 1981 berry failure. Appeared to be many fewer bears present than in 1980. Craig Gardner noted that along the Denali highway "Berries were very abundant along the Denali Hwy from Paxton to the McClaren River."

Berry transects support substantially fewer blueberries and crowberries in upstream areas, about average in downstream areas. Berries appeared to be very abundant in highly localized pockets, more patchy than is typically the case. Black bear movements appeared normal but some brown bears made atypically large movements in fall 1984. Between Paxton and the McClaren River, Craig Gardner (pers. comm.) reported "Berries were less abundant than in 1983 but more abundant than in 1981."

In the, vicinity of Watana Camp berries appeared to be slightly below average in abundance. In more upstream habitat they appeared to be slightly above average. Saw nowhere where blueberries were really thick, pretty well dispersed. Along the Denali Hwy both Craig Gardner and Jack Whitman noted independently that berry crops "appeared to be a bust" -- very few were seen.

No data collected in study area. Along the Denali Highway on 8/10/86, Jack Whitman noted "I spent 3 days on west end of Denali Highway. Walked many miles in vicinity of 25 mile, 22 mile, and 15 mile. Excellent berry crop in all locations. Best I've noted in 4 years."

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Table 68. Den entrance and emergence dates of radio-collared black bears for the winter of 1980-81 ("S" is the standard deviation, but it includes variability from the fluctuating time between observations, as well as variability in denning times).

| Bear ID | Sex | Reproduct Ive status at exit | 1980 Entrance |  |  |  | Emergen |  | Days In Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Min. | Max. | MId. | Min. | Max. | Mid. | MIn. | Max. | MId. |
| 287 | M |  | 9 Sept. | 29 Sept. | 19 Sept. | 30 Apr. | 5 May | 2 May | 213 | 238 | 212 |
| 289 | F | $3 @ 0$ | 9 Sept. | 29 Sept. | 19 Sept. | 5 May | 15 May | 10 May | 221 | 248 | 235 |
| 290 | F | w/o | 1 Oct. | 9 Oct. | 5 Oct. | 5 May | 10 May | 8 May | 208 | 221 | 215 |
| 301 | F | $2 @ 0$ | 29 Sept. | 13 Oct. | 6 Oct. | 9 May | 29 May | 19 May | 208 | 242 | 225 |
| 303 | M |  |  |  |  | 30 Apr . | 5 May | 2 May |  |  |  |
| 304 | M |  |  |  |  | 5 May | 10 May | 8 May |  |  |  |
| 317 | F | 201 | 9 Sept. | 29 Sept. | 19 Sept. | 5 May | 15 May | 10 May | 218 | 248 | 233 |
| 318 | F | $1 @ 1$ | 29 Sept. | 13 Oct. | 6 Oct. | 30 Apr . | 5 May | 2 May | 199 | 218 | 209 |
| 319 | M |  | 29 Sept. | 13 Oct. | 6 Oct. | 30 Apr . | 5 May | 2 May | 199 | 218 | 209 |
| 321 | F | $2 @ 0$ | 9 Sept. | 29 Sept. | 19 Sept. | 10 May | 15 May | 12 May | 223 | 248 | 236 |
| 322 | M |  | 9 Sept. | 13 Oct. | 26 Sept. |  |  |  |  |  |  |
| 323 | M |  | 29 Sept. | 13 Oct. | 6 Oct. | 6 May | 8 May | 7 May | 205 | 228 | 217 |
| 324 | M |  | 29 Sept. | 13 Oct. | 6 Oct. | 30 Apr. | 5 May | 2 May | 199 | 218 | 209 |
| 325 | F | w/o | 29 Sept. | 9 Oct. | 4 Oct. |  |  |  |  |  |  |
| 327 | F | $1 @ 1$ | 9 Sept. | 29 Sept. | 19 Sept. | 8 May | 10 May | 9 May | 221 | 243 | 232 |
| 328 | F | 2@0 | 9 Sept. | 29 Sept. | 19 Sept. | 21 May | 29 May | 25 May | 234 | 262 | 248 |
|  | $\begin{aligned} & \text { MALES } \\ & \text { "S" } \end{aligned}$ |  | 19 Sept. | $\begin{aligned} & 70 \text { oct. } \\ & 7 \end{aligned}$ | $\underset{8}{28 \text { Sept. }}$ | ${ }_{6}^{5}$ May | ${ }_{8}^{\text {I2 May }}$ | ${ }_{7}^{8}$ May | 12 11 | 236 15 | 223 13 |
|  | n |  | 14 | 14 | 14 | 14 | 14 | 14 | 12 | 12 | 12 |

Table 69. Den entrance and emergence dates of radio-collared black bears for the winter of 1981-82 ("S" is the standard deviation, but it includes variability from the fluctuating time between observations, as well as variability in denning times).

| Bear ID | Sex | Reproductive status at exit | 1981 Entrance |  |  | 1982 Emergence |  |  | Days In Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Min. | Max. | Mid. | Min. | Max. | Mid. | Min. | Max. | MId. |
| 287 | M |  | 24 Aug. | 9 Sept. | 9 Sept. | 4 May | 6 May | 5 May | 237 | 255 | 246 |
| 289 | F | $2 @ 1$ | 23 Sept. | 1 Oct. | 28 Sept. | 12 May | 18 May | 15 May | 223 | 237 | 230 |
| 301 | F | $2 @ 1$ | 16 Sept. | 22 Sept. | 19 Sept. | 6 May | 18 May | 12 May | 226 | 244 | 235 |
| 302 | M |  | 16 Sept. | 22 Sept. | 19 Sept. | ? | 6 May | 6 May* |  | 232 | 229 |
| 303 | M |  | 16 Sept. | 22 Sept. | 19 Sept. | 12 May | 18 May | 15 May | 232 | 244 | 238 |
| 304 | M |  | 16 Sept. | 1 Oct. | 24 Sept. | 6 May | 12 May | 9 May | 217 | 238 | 228 |
| 317 | F | w/o | 9 Sept. | 16 Sept. | 12 Sept. | 12 May | 18 May | 15 May | 238 | 251 | 244 |
| 318 | F | w/o | 16 Sept. | 22 Sept. | 19 Sept. | 18 May | 26 May | 22 May | 238 | 252 | 245 |
| 321 | F | w/o | 16 Sept. | 22 Sept. | 19 Sept. | 6 May | 12 May | 9 May | 226 | 238 | 232 |
| 323 | M |  | 22 Sept. | 1 Oct. | 27 Sept. | 6 May | 12 May | 9 May | 217 | 232 | 224 |
| 324 | M |  | 1 Oct. | 7 Oct. | 4 Oct. | 4 May | 6 May | 5 May | 209 | 217 | 213 |
| 327 | F | w/o | 16 Sept. | 22 Sept. | 19 Sept. | 12 May | 18 May | 15 May | 232 | 244 | 238 |
| 329 | F | w/o | 22 Sept. | 1 Oct. | 27 Sept. | 12 May | 18 May | 15 May | 223 | 238 | 230 |
| 343 | M |  | 16 Sept. | 22 Sept. | 19 Sept. | 12 May | 18 May | 15 May | 232 | 244 | 238 |
| 346 | M |  | 9 Sept. | 16 Sept. | 12 Sept. | ? | 6 May | 6 May* |  | 239 | 236 |
| 348 | M |  | 16 Sept. | 22 Sept. | 19 Sept | 4 May | 6 May | 5 May | 224 | 232 | 228 |
| 349 | F | w/o | 9 Sept. | 16 Sept. | 12 Sept. | ? | 6 May | 6 May* |  | 239 | 236 |
| 325 | F | ? | 9 Sept. | 16 Sept. | 12 Sept. |  |  |  |  |  |  |
| 328 | F | ? | 16 Sept. | 22 Sept. | 19 Sept. |  |  |  |  |  |  |
|  | $\begin{aligned} & \text { MEAN } \\ & \text { "S" } \end{aligned}$ |  | $\underset{8}{\overline{15} \text { Sept. }}$ | $\begin{gathered} 23 \text { Sept. } \\ 7 \end{gathered}$ | $\underset{6}{19} \text { Sept. }$ | $\begin{aligned} & 9 \text { May } \\ & 4 \end{aligned}$ | $\begin{gathered} \overline{13} \text { May } \\ 6 \end{gathered}$ | $\overline{{ }_{5}^{11 ~ M a y}}$ | $\overline{227}$ 9 | 240 9 | 234 8 |
|  | n |  | 19 | 19 | 19 | 14 | 17 | 17 | 14 | 17 | 17 |

[^11]Table 70. Den entrance and emergence dates of radio-collared black bears for the winter of 1982-83 ("S" is the standard deviation, but it includes variability from the fluctuating time between observations, as well as variability in denning times).

|  |  | Reproductive status | 1982 Entrance |  |  |  | Emergenc |  | Days in Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bear ID | Sex | at exit | Min. | Max. | Mid. | Min. | Max. | Mid. | Min. | Max. | MIC. |
| 289 | F | $2 @ 0$ | 28 Sep | 6 Oct | 2 Oct | 10 May | 15 May | 13 May | 216 | 230 | 223 |
| 303 | M |  | 29 Sep | 20 Oct | 10 0ct | 4 May | 10 May | 7 May | 196 | 223 | 210 |
| 317 | F | 2 @ | 20 Sep | 29 Sep | 24 Sep | 10 May | 23 May | 17 May | 223 | 245 | 234 |
| 318 | F | 2@0 | 6 Oct | 15 Oct | 10 Oct | 10 May | 23 May | 17 May | 207 | 229 | 218 |
| 321 | F | w/o | 20 Sep | 29 Sep | 24 Sep | 10 May | 15 May | 13 May | 223 | 237 | 230 |
| 323 | M |  | 6 Oct | 15 Oct | 10 Oct | 25 Apr | 4 May | 30 Apr | 192 | 210 | 201 |
| 324 | M |  | 29 Sep | 6 Oct | 2 Oct | 25 Apr | 4 May | 30 Apr | 201 | 217 | 209 |
| 327 | F | $2 @ 0$ | 6 Oct | 15 Oct | 10 Oct | 4 May | 10 May | 7 May | 201 | 216 | 209 |
| 329 | F | w/o | 29 Sep | 6 Oct | 2 Oct | 25 Apr | 4 May | 30 Apr | 201 | 217 | 209 |
| 343 | M |  | 6 Oct | 20 Oct | 13 Oct | 4 May | 10 May | 7 May | 196 | 216 | 206 |
| 346 | M |  | 6 Oct | 15 Oct | 10 Oct | 25 Apr | 4 May | 30 Apr | 192 | 210 | 201 |
| 349 | F | w/o | 29 Sep | 6 Oct | 2 Oct | 10 May | 18 May | 14 May | 216 | 231 | 224 |
| 354 | F | $1 @ 1$ | 6 Oct | 15 Oct | 10 Oct | 10 May | 23 May | 17 May | 207 | 229 | 218 |
| 357 | M |  | 6 0ct | 15 Oct | 10 Oct | (BEAR | ED DURING | ER) | - | - | - |
| 358 | M |  | 29 Sep | 6 Oct | 2 Oct | 4 May | 10 May | 7 May | 210 | 223 | 217 |
| 359 | M |  | 60 ct | 150 ct | 10 Oct | 4 May | 10 May | 7 May | 201 | 216 | 209 |
| 360 | M |  | 6 Oct | 15 Oct | 10 Oct | 25 Apr | 4 May | 30 Apr | 192 | 210 | 201 |
| 361 | F | $3 @ 0$ | 6 Oct | 15 Oct | 10 Oct | 10 May | 23 May | 17 May | 207 | 229 | 218 |
| 363 | F | w/o | 6 Oct | 15 Oct | 10 Oct | 25 Apr | 4 May | 30 Apr | 192 | 210 | 201 |
| 365 | M |  | 6 Oct | 20 Oct | 13 Oct | 25 Apr | 4 May | 30 Apr | 187 | 210 | 199 |
| 367 | F | w/o | 6 Oct | 15 Oct | 10 Oct | 10 May | 19 May | 15 May | 207 | 225 | 216 |
| 369 | F | w/o | 6. Oct | '15 Oct | 10 Oct | 25 Apr | 4 May | 30 Apr | 192 | 210 | 201 |
| 370 | F | 2 @0 | 6 Oct | 15 Oct | 10 Oct | 4 May | 10 May | 7 May | 201 | 216 | 209 |
| 372 | F | 3@0 | 29 Sep | 6 Oct | 2 Oct | 10 May | 19 May | 15 May | 216 | 232 | 224 |
| 375 | F | $2 @ 0$ | 29 Sep | 6 Oct | 20 ct | 25 Apr | 4 May | 30 Apr | 201 | 217 | 209 |
| 376 | F | $3 @ 0$ | 6 Oct | 15 Oct | 10 Oct | 25 Apr | 4 May | 30 Apr | 192 | 210 | 201 |
| 377 | F | $1 @ 0$ | 29 Sep | 6 Oct | 2 Oct | 4 May | 10 May | 7 May | 210 | 223 | 217 |
| 378 | F | $2 @ 0$ | 20 Sep | 29 Sep | 24 Sep | 4 May | 10 May | 7 May | 217 | 232 | 225 |
| 379 | F | $3 @ 0$ | N. D. | N. D. | N. D. | 4 May | 10 May | 7 May | - | $\stackrel{-}{2}$ | 5 |
| 301 | F | 2 @ | N. D. | N. D. | N. D. | 4 May | 10 May | 7 May | - | - | - |
| 374 | F | $3 @ 0$ | N. D. | N. D. | N. D. | 10 May | 19 May | 15 May | - | - | - |
|  | MEAN |  | $20 c t$ | 11 Oct | 60 CE | 3 May | 17 May | 7 May | 204 | 221 | 213 |
|  | "S" |  | 5 | 6 | 6 | 6 | 7 | 6 | 10 | 10 | 10 |
|  | $n$ |  | 28 | 28 | 28 | 30 | 30 | 30 | 27 | 27 | 27 |

Table 71. Black bear den entrance and emergence dates, winter of 1983/84.


Table 72. Black bear den entrance and emergence dates, winter of 1984/85:

| Bear ID |  |  | Repro- <br> ductive <br> status <br> at exit | 1983 Entrance |  |  | 1984 Emergence |  |  | Days in Den |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Sex |  | earliest | latest | Mid. | earliest | latest | Mid. | Min. | Max. | Mid. |
|  | B289 | $F$ | $2 @ 0$ | 1 Oct | 11 Oct | 6 Oct | 23 May | 1 June | 28 May | 224 | 243 | 234 |
|  | B317 | F | $2 @ 0$ | 1 Oct | 11 Oct | 6 Oct | 23 May | 1 June | 28 May | 224 | 243 | 234 |
|  | B321 | F | 101 | 1 Oct | 11 Oct | 6 Oct | 9 May | 16 May | 13 May | 210 | 227 | 219 |
|  | B329 | F | w/o | 11 Oct | 24 Oct | 18 Oct | 9 May | 16 May | 13 May | 197 | 217 | 207 |
|  | B354 | F | w/o | 1 Oct | 11 Oct | 6 Oct | 23 May | 4 June | 29 May | 224 | 246 | 235 |
|  | B359 | M |  | 1 Oct | 11 Oct | 60 ct | 9 May | 16 May | 13 May | 210 | 227 | 219 |
|  | B361 | $F$ | $3 @ 2$ | 11 Oct | 24 Oct | 18 Oct | 9 May | 16 May | 13 May | 197 | 217 | 207 |
|  | B363* | F | 2@1 | 1 Oct | 11 Oct | 6 Oct | 9 May | 16 May | 13 May | 210 | 227 | 219 |
| $\stackrel{N}{N}$ | B369* | F | $1 @ 1$ | 11 Oct | 24 Oct | 18 Oct | 9 May | 16 May | 13 May | 197 | 217 | 207 |
| ¢ | B375* | F | ? | 11 Oct | 24 Oct | 18 Oct | 23 May | 31 May | 27 May | 211 | 232 | 221 |
|  | B376* | F | w/o | 11 Oct | 24 Oct | 18 Oct | 9 May | 16 May | 13 May | 197 | 217 | 207 |
|  | B377* | F | $2 @ 0$ | 1 Oct | 11 Oct | 6 Oct | 16 May | 23 May | 20 May | 212 | 234 | 226 |
|  | B378* | F | $1 @ 0$ | 21 Sep | 1 Oct | 26 Sep | 23 May | 5 June | 30 May | 234 | 257 | 246 |
|  | B387 | M |  | 1 Oct | 11 Oct | 6 Oct | 30 Apr | 9 May | 5 May | 201 | 220 | 211 |
|  | B401 | M |  | 1 Oct | 24 Oct | 13 Oct | 30 Apr | 9 May | 5 May | 189 | 220 | 204 |
|  | B402* | $F$ | $2 @ 0$ | 24 Oct | 7 Nov | 31 Oct | 16 May | 23 May | 20 May | 190 | 211 | 201 |
|  | B404* | F | $3 @ 0$ | 11 Oct | 24 Oct | 18 Oct | 16 May | 23 May | 20 May | 204 | 224 | 214 |
|  | B405* | F | 2@2 | 21 Sep | 1 Oct | 26 Sep | 23 May | 5 June | 30 May | 234 | 257 | 246 |
|  | B408* | M |  | 11 Oct | 24 Oct | 18 Oct | No effort | -- | -- | -- | -- | -- |
|  | B409* | F | w/o | 11 Oct | 24 Oct | 18 Oct | 16 May | 23 May | 20 May | 204 | 224 | 214 |
|  | B411* | F | 201 | 1 Oct | 11 Oct | 6 Oct | 16 May | 23 May | 20 May | 212 | 234 | 226 |
|  | B328 | F | 3@0 | 6 Sep | 21 Sep | 14 Sep | 16 May | 23 May | 20 May | 237 | 259 | 248 |
|  | B349 | $F$ | 2@0 | 1 Oct | 11 Oct | 6 Oct | 16 May | 23 May | 20 May | 212 | 234 | 226 |
|  | B364 | F | w/o | 21 Sep | 1 Oct | 26 Sep | 23 May | 3 June | 28 May | 234 | 255 | 244 |
|  | B416 | M |  | 21 Sep | 1 Oct | 26 Sep | 16 May | 23 May | 20 May | 227 | 244 | 236 |
|  | B302 | M |  | 1 Oct | 24 Oct | 130 ct | 9 May | 16 May | 13 May | 197 | 227 | 212 |
|  |  | Mean |  | 30 ct | 15 Oct | 9 Oct | 14 May | 23 May | 19 May | 212 | 233 | 223 |
|  |  | "S" |  | 9.5 | 10.5 | 9.9 | 7.0 | 8.1 | 7.5 | 14.6 | 14.3 | 14. 5 |
|  |  | $n$ |  | 28 | 27 | 27 | 25 | 25 | 25 | 25 | 25 | 25 |

* Downstream bear

Table 73. Characteristics of black bear dens in the Susitna study area during winters of 1980/1981, 1981/1982, 1982/1983, 1983/84, 1984/85.


Table 73, (continued)


Table 73. (continued)


Table 73. (continued)

|  | $\begin{aligned} & \text { Den } \\ & \text { No. } \end{aligned}$ |  | $\begin{gathered} \text { Bear } \\ \text { ID No. } \end{gathered}$ | $\begin{gathered} \text { Age at } \\ \text { Exit } \end{gathered}$ | $\begin{gathered} \text { Eleva- } \\ \text { tion } \\ \text { (feet) } \\ \hline \end{gathered}$ | Slope (Degrees) | $\begin{aligned} & \text { Aspect** } \\ & \text { (True N) } \end{aligned}$ | Vegetation | \% Canopy Tree Coverage | ENTRANCE |  | CHAMBER |  |  | Total <br> Length (cin) | Previously Used? (Yes/No) | A | B | C |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} \mathrm{Ht} \\ \left(\mathrm{~cm}_{0}\right) \end{gathered}$ |  |  |  |  |  |  | $\begin{aligned} & \text { WIdth } \\ & \text { (cm.) } \end{aligned}$ | $\begin{aligned} & \operatorname{Ln} . \\ & (\mathrm{cm} .) \end{aligned}$ | $\begin{aligned} & \text { Width } \\ & \left(\mathrm{cm}_{\mathrm{s}}\right) \end{aligned}$ | $\begin{gathered} \mathrm{Ht} \\ (\mathrm{~cm} .) \\ \hline \end{gathered}$ |  |  |  |  |  |
| FEMALES w/offspring (at exit) (continued) |  |  |  |  |  |  |  |  | 0 | - | - | - | - | - | - | No? | 1 | - | No |
| w/1@2? |  | 174 |  | B364 | 12 | 2145 | 22 | 326 | Spruce-Birch | 40 | 33 | 39 | 110 | 113 | 73 | 183 | No? | 2 | - | Yes |
| w/ 2 @0 |  | 181 | B317 | 12 | 2055 | 32 | 287 | Alder-Birch | 20 | 50 | 59 | 152 | 133 | 78 | 152 | No | 3 | No | - |
| w/3@0 |  | 186 | B404 | 13 | 1975 | 26 | 214 | Alder-Spruce | 10 | 27 | 67 | 193 | 91 | 72 | 193 | Yes | 3 | - | - |
| w/2@0 |  | 187 | B402 | 12 | 1910 | 21 | 133 | Alder | 0 | 38 | 63 | 130 | 98 | 54 | 134 | No? | 3 | - | - |
| W/2@0 |  | 188* | B377 | 7 | 1500 | 35 | 38 | Alder | 0 | - | - | - | - | - | - | - | - | - | - |
| w/2@1 |  | 198* | B369 | 7 | 1100 | - | - | Alder-Birch | - | - | - | - | - | - | - | - | - | - | - |
| w/2@0 |  | 203* | B289 | 14 | 1600 | - | - | Spruce | - | - | - | - | - | - | - | - | - | - | - |
| FEMALES w/o offspring (at exit) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| N |  | 34 | B321 | 12 | 2125 | 22 | 184 | Alder | 10 | 29 | 43 | 99 | 118 | 79 | 193 | No | 2 | No | - |
| - |  | 43 | B31 7 | 9 | 2250 | 8 | 153 | Dwarf Birch | 0 | 32 | 36 | 92 | 89 | 63 | 150 | No | 2 | No | - |
|  |  | 55 | B349 | 5 | 2650 | 21 | 207 | Alder/Spruce | 10 | 39 | 54 | 56 | 92 | 55 | 124 | No |  | - | No |
|  |  | 58 | B327 | 7 | 1675 | 26 | 321 | Birch/Alder | 70 | 35 | 49 | 86 | 73 | 61 | 160 | No | 3 | - | Yes |
|  |  | 67 | B369 | 5 | 1410 | 21 | 78 | Grass/Alder/Spruce | - 25 | 36 | 51 | - | 91 | 71 | 104 | No | 3 | - | - |
|  |  | 80 | B329 | 3 | 1725 | 31 | 28 | Alder | 90 | 24 | 43 | 102 | 84 | 53 | 165 | No | 5 | - | Yes |
|  |  | 82 | B367 | 5 | 1960 | 30 | 323 | Alder/Fern | 80 | 36 | 38 | 102 | 130 | 81 | 152 | No | 4 | - | - |
|  |  | 99* | B363 | 5 | 2775 | 21 | 177 | Alder | 90 | 30 | 74 | - | 112** | 53** | 94** | No | 3 | - | No |
|  |  | 142 | B411 | 9 | 1475 | 7 | 105 | Alder/Birch/Spruce | - 100 | 34 | 57 | 139 | 117 | 57 | 220 | Yes | 3 | - | - |
| MALES | \#\#\# | 20*** | B323* | 3 | 1950 | 71 | 176 | Alder/Birch/Spruce | - 80 | 166 | 25 | 217 | 76 | 36 | 454 | Yes | 3 | - | Yes |
|  |  | 35 | B304 | 12 | 1650 | 36 | 79 | Birch | 25 | 53 | 147 | 100 | 173 | - | 660 | Yes | 2 | No | - |
|  |  | 38* | B343 | 6 | 1200 | 39 | 313 | Birch/Alder/Spruce | - 60 | 35 | 62 | - | - | - | - | No | $?$ | - | - |
|  |  | 39 | B348 | 10 | 1375 | 43 | 240 | Birch/Spruce | 20 | 57 | 91 | 116 | 172 | 183 | 530 | Yes | 1 | - | - |
|  |  | 57 | B302 | 10 | 2025 | 41 | 236 | Spruce/Birch | 40 | 55 | 63 | 94 | 138 | 101 | 188 | Yes | 2 | - | Yes |
|  |  | 71 | B365 | 6 | 900** | 10** | - | Alder/Birch/Spruce |  | - | - | - | - | - | - | - |  |  |  |

Table 73. (continued)

|  | $\begin{aligned} & \text { Den } \\ & \text { No. } \end{aligned}$ | $\begin{gathered} \text { Bear } \\ \text { ID No. } \\ \hline \end{gathered}$ | Age at Exit | $\begin{gathered} \text { Eleva- } \\ \text { (fion } \\ \text { (feet) } \\ \hline \end{gathered}$ | Slope (Degrees) | Aspect** <br> (True N) | Vegetation | 8. Canopy Tree Coverage | ENTRANCE |  | CHAMBER |  |  | Total Length (cm) | $\begin{aligned} & \text { Previously } \\ & \text { Used? } \\ & \text { (Yes/No) } \\ & \hline \end{aligned}$ | A | B | C |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  | $\begin{aligned} & \text { Ht. } \\ & \text { (cm.) } \end{aligned}$ | $\begin{aligned} & \text { Width } \\ & \text { (cm.) } \end{aligned}$ | $\begin{gathered} \mathrm{Ln}_{\bullet} \\ \left(\mathrm{cm}_{+}\right) \end{gathered}$ | $\begin{aligned} & \text { Width } \\ & \text { (cm.) } \end{aligned}$ | $\begin{aligned} & \begin{array}{l} \mathrm{Ht} . \\ (\mathrm{cm}) \end{array} \end{aligned}$ |  |  |  |  |  |
| MALES (continued) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | B387 | 5 | 3375 | 25 | 111 | Alder/D. Birch | 80 | - | 40 | - | - | - | - | No | 4 | - | No |
|  | 126* | B359 | 6 | 2375 | 0 | 9 | Spruce/D. Birch | 50 | - | - | - | - | - | 354** | No | 2 | - | No |
|  | 128 | B360 | 9 | 2150 | 14 | 239 | Alder/Spruce | 110 | 54 | 57 | 90 | 160 | 84 | 146 | No | 3 | - | Yes |
|  | 159 | B302 | 13 | 2030 | 29 | 34 | Alder | 0 | 47 | 77 | 142 | 111 | 64 | 200 | Yes | 2 | - | Yes |
|  | 202* | B416 | 10 | 1700 | - | - | - | - | - | - | - | - | - | - | - | - | No | - |
| SPECIES UNKNOWN |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UNKNOWN CAVITY TYPE MALES |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 40 | B324 | 7 | 1400** | - | - | -- | - | - | - | - | - | - | - | - | - | ? | - |
| $\begin{aligned} & \mathrm{N} \\ & \mathrm{~N} \end{aligned}$ | 51\#\#\# | \# B346 | 10 | 2370** | 30 | 168** | Spruce/Birch | 0 | 38 | 53 | - | - | 48 | - | Yes | - | - | No |
|  | 62 | B319 | 4 | 1600** | 60** | 118** | Spruce/Alder | - | - | - | - | - | - | - | - | - | - | - |
| FEMALES | 65* | B329 | 1 | 1900** | 45** | 28** | -- | - | - | - | - | - - | - | Yes | - |  |  |  |
|  | 63* | B290 | 9 | 1850** | 15** | 73** | -- | - | - | - | - | - | - | - | - | - | - | No |
|  | 64* | B290 | 9 | 1700** | 15** | 28** | -- | - | - | - | - | - | - | - | - | - | - | No |
| w/1@0 | 190* | B378 | 9 | 2000 | 62 | 308 | Alder | 0 | - | - | - | - | - | - | - | - | - | - |
| UNKNOWN SEX | 61 | ? | $?$ | 2400 | 35** | 163** | Spruce/Alder/Birch | - 80 | - | - | - | - | - | - | No | 4 | - | No |

[^12]Table 73. (continued)

* Actual den site not found or too difficult to enter or collapsed.
** Approximate value.
A Subjective characterization of quality, $1=$ highest and $5=$ lowest.
B Will be flooded by Devil's Canyon tmpoundment?
C Will be flooded by Watana impoundment?
*** Den not located first year known
but thought to be the same location as subsequently found den. Den No. 158=171.
**** Mag. $\mathrm{N}+28^{\circ}=$ True N. of hillside.
\# Used by the same bear two consecutive winters.
\#\# Used by the offspring during natal winter and subsequent winter. \#\#\# Used by different radio-collared bear during subsequent winter.

Dens No. 8, 19, 6, 7, 9 10, 13, 18, 2, 4, 11, 12, 21, 20, 62, 63, 64 used during winter of 1980/1981.

Dens No. 32, 33, 50, 34, 43, 55, 58, 35, 38, 39, 57, 40, 49, 51, 61, 65, 7, 9, 10, 4, 21, used during winter of 1981/1982.

Dens No. $73,88,92,93,85,51,66,95,96,98,100,72,68,69,70$, $74,75,81,83,84,90,91,97,67,80,82,99,71,10,7,9$, 19 used during winter 1982/1983.

Dens No. 113, 129, 20, 115, 144, 49, 146, 154, 145, 114, 127, 138, 141, 143, 142, 116, 126, 128, 140, 152, 156, 147, 9, 51, 88, 92, and 73 used during winter 1983/84.

Dens No. 168, 169, 172, 180, 184, (158), 185, 191, 167, 173, 160, 174, 181, 186, 187, 188, 198, 203, (159), 202, 190, (85), (49), (74), used during winter 1984/85.

Table 74. (Continued)

|  |  | 1982/83 |  |  | 1983/84 |  |  | 1984/85 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bear No. | Sex | Cavity <br> Type | Den\# | $\text { Assoc }{ }^{* *}$ | Cavity <br> Type | Den\# | $\text { As soc }^{* *}$ | $\begin{aligned} & \text { Cavity } \\ & \text { Type } \end{aligned}$ | Den\# | $\text { Assoc }{ }^{\star *}$ |
| 376 | F | Dug | 91 | w/3@0 | Natural | 144 | w/o | Nat. | 85 | w/o? |
| 377 | F | Natural | 85 | w/o | Tree | 146 | W/?@0? | Dug | 188 | w/2@0 |
| 378 | F | Dug | 90 | w/2@0 | Tree | 154 | w/2@1 | Nat. | 190 | w/100 |
| 379 | F | Natural | 19 | w/3@0 | Dead-- |  |  |  |  |  |
| 387 | M |  |  |  | Dug | 116 | w/o | Nat. | 167 | -- |
| 401 | M |  |  |  | Natural | 157 | w/o | Nat. | 49 | -- |
| 402 | F |  |  |  | Tree | 145 | w/o | Dug | 187 | w/2c |
| N 404 | F |  |  |  | Natural | 92 | w/o | Dug | 186 | w/3@0 |
| N 405 | F |  |  |  | Dug | 143 | w/2@1 | Nat. | 185 | w/o |
| 408 | M |  |  |  | Natural | 157 | w/o | Unk. | 201 | w/o |
| 411 | F |  |  |  | Dug | 142 | w/2@0 | Nat. | 184 | w/2@1 |
| 416 | M |  |  |  |  |  |  | Dug | 202 | -- |
| 364 | F |  |  |  |  |  |  | Dug | 174 | w/1@2? |

** Associations are at time of emergence
*** Den 158 was capture site of B289 (mother of B329) in spring 1980. Den not flagged until winter 84/85, assumed was 79/80 den of B289

Table 74. History of den use by individual radio-marked black bears, 1980/81-1984/85.


Table 75. History of use of individual black bear dens by radio-marked black bears, 1980/81-1984/85 (blanks indicate no data available, den not revisited and no radio-marked bear there). "Flooded" means would be inundated by impoundment.

| Den No. | Den Type | Flooded | Location ${ }^{* * *}$ | 80/81 |  | 81/82 | 82/83 | 83/84 | 84/85 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 158 | Dug | Yes | W | [B289 in 79/80 spring | W/2@1] | Unk. 80/81, 81/82 | -- | -- | B3 29 female |
| 2 | Dug | Yes | W | B301 female w/2@0 |  | Vacant | Vacant | Vacant |  |
| 4 | Dug | Yes | W | B289 female w/3@0 |  | B289 female w/2@1 | Vacant | Vacant | Vacant |
| 6 | Nat | No | D | B325 female w/o |  |  |  |  |  |
| 7 | Nat | No | D | B287 male |  | B287 male | B321 female w/o |  |  |
| 8 | Nat | No | D | B321 female w/2@0 |  |  |  |  |  |
| 9** | Nat | No | D | Bs 24 male |  | B3 25 female w/o | B3 24 male | B3 24 male | Vacant |
| 10 | Nat | No | D | B303 male |  | B303 male | B303 male | Vacant |  |
| 11 | Dug | No | D | B317 female w/2@1 |  |  |  |  |  |
| 12 | Dug | No | D | B318 female w/1@1 <br> (B330 male) |  | Collapsed--------- | ----------------* |  |  |
| 13 | Nat | No | D | B304 male |  |  |  |  |  |
| 18 | Nat | Yes | W | B3 22 male |  |  |  |  |  |
| 19 | Nat | No | D | B328 female w/2@0 |  |  | B379 female w/3@0 |  |  |
| 20 | Nat | Yes | W | B3 23 male |  |  |  | B317 female <br> w/1@1 | Vacant |
| 21 | Dug | Yes | W | B327 female w/B329@1 |  | B329 female w/o | Collapsed--------- | - ----- |  |
| 32 | Nat | No | D |  |  | B328 female w/l@l | Vacant |  | Vacant |
| 33 | Nat | No | D |  |  | B318 female w/o |  |  |  |
| 34 | Dug | No | D |  |  | B321 female w/o |  |  |  |
| 35 | Dug | No | D |  |  | B304 male | Vacant------------ |  |  |
| 38 | Dug | No | DS |  |  | B343 male | Collapsed--------- | - ----------- |  |
| 39 | Dug | No | DS |  |  | B348 male | Vacant |  |  |
| 40 | - | Yes | D |  |  | B3 24 male |  |  |  |
| 43 | Dug | No | D |  |  | B317 female w/o |  |  |  |
| 49 | Nat | Yes | W |  |  | B3 23 male (?) |  |  | B401 male |
| 51* | Nat | No | W |  |  | B346 male | B323 male | B346 male |  |
| 50 | Dug | No | W |  |  | B301 female w/2@1 | Vacant | Vacant |  |
| 55 | Dug | No | W |  |  | B349 female w/o |  |  |  |
| 57 | Dug | Yes | W | . |  | B302 male | Vacant | Vacant | Vacant |
| 58 | Dug | Yes | W |  |  | B327 female w/o | Vacant |  |  |
| 61 | Dug | No | W | - |  | Unmarked BKB |  |  |  |
| 62 | - | No | D | B319 male |  |  |  |  |  |
| 63 | - | No | D | B390 female w/o |  |  |  |  |  |
| 64 | - | No | D | B390 female w/o |  |  |  |  |  |
| 65 | - | Yes | W |  |  | B329 female w/o |  |  |  |
| 66 | Nat | No | D |  |  |  | B343 male |  |  |
| 67 | Dug | No | DS |  |  |  | B369 female w/o | -------- |  |
| 68 | Dug | No | D |  |  |  | B318 female w/2@0 | Collapsed-- |  |
| 69 | Dug | No | D |  |  |  | B317 female w/2@0 |  |  |
| 70 | Dug | No | W |  |  |  | B301 female w/2@0 | Vacant | Vacant |
| 71 | Dug | No | DS | - |  |  | B365 male |  |  |

(continued on next page)

Table 75. (Continued)

| Den No. | Den Type | Flooded | Location ${ }^{* * *}$ | 80/81-81/82 | 82/83 | 83/84 | 84/85 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 72 | Nat | No | W |  | Unmarked BKB |  |  |
| 73 | Nat | Yes | W |  | B327 female w/2@0 | B329 Female w/1@ | Vacant |
| 74 | Dug | No | W |  | B349 female w/2@0 |  | B349 |
| 75 | Dug | No | W |  | B361 female w/4@0 |  |  |
| 80 | Dug | Yes | W |  | B329 female w/o |  |  |
| 81 | Dug | Yes | W |  | B389 female w/2@0 | Vacant |  |
| 82 | Dug | No | DS |  | B367 female w/o |  |  |
| 83 | Dug | No | DS |  | B370 female w/2@0 |  |  |
| 84 | Dug | No | DS |  | B372 female w/3@0 |  |  |
| 85 | Nat | No | DS |  | B377 female w/o |  | B376 |
| 88 | Nat | No | DS |  | B375 female w/2@0 | B375 female w/2@1 |  |
| 90 | Dug | No | DS |  | B378 female w/2@0 |  |  |
| 91 | Dug | No | DS |  | B376 female w/3@0 |  |  |
| 92 | Nat | No | DS |  | B374 female w/3@0 | B404 female w/o |  |
| 93 spring | Nat | No | DS |  | B374 female w/3@0 |  |  |
| 95 | Nat | Yes | W |  | B360 male | Vacant |  |
| 96 | Nat | Yes | W |  | B346 male |  |  |
| 97 | Dug | No | W |  | B354 female w/1@1 | Collapsed-------- | --------------------- |
| 98 | Nat | Yes | W |  | B359 male | Vacant | Vacant |
| 99 | Dug | No | W |  | B363 female w/o | Collapsed-------- | ------- |
| 100 | Nat | No | W |  | B358 male | Collapsed-----.-- | ------ |
| 113 | Nat | No | W |  |  | B354 female w/2@0 |  |
| 114 | Dug | No | W |  |  | B363 female w/2@0 | Vacant |
| 116 | Dug | No | W |  |  | B387 male | Collapsed------------ |
| 126 | Dug | No | W |  |  | B359 male | Collapsed------------ |
| 127 | Dug | Yes | W | . |  | B361 female w/3@1 | Vacant |
| 128 | Dug | Yes | W |  | . | B360 male |  |
| 129 | Nat | Yes | W |  |  | B289 female w/1@1 | Vacant |
| 157 | Nat | Yes | W |  |  | B401 male |  |
| 138 | Dug | No | ${ }_{\text {D }}^{\text {D }}$ |  |  | B321 female w/?@0 | Collapsed------------ |
| 140 | - | No | DS |  |  | B406 female w/2@1 |  |
| 141 | Dug | No | DS |  |  | B369 female w/2@0 |  |
| 142 | Dug | No | DS |  |  | B411 female w/o B405 female w/2@1 |  |
| 144 | Nat | No | DS |  |  | B376 female w/o |  |
| 145 | Tree | No | DS |  |  | B402 female w/o | Vacant |

(continued on next page)

Table 75. (Continued)

| Den No. | Den Type | Flooded | Location ${ }^{* * *}$ | 80/81-82/83 | 83/84 | 84/85 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 146 | Tree | No | DS |  | B377 female w/?@o | Vacant |
| 147 | - | - | D |  | B343 male |  |
| 152 | - | No | DS |  | B409 female w/o |  |
| 154 | Tree | No | DS |  | B378 female w/2@1 |  |
| 156 | Nat | No | DS |  | B408 male |  |

* Attempted initial denning location for B323, B346, \& B360 in 1982/1983. B346 \& B360 subsequently moved.
** At tempted dening location for B3 $24 \&$ B325 in 1981/1982. B324 subsequently moved.
*** $\mathrm{W}=$ Watana, $\mathrm{D}=$ Devils Canyon, $\mathrm{DS}=$ Downstream of impoundment zone.
SUMMARY OF TABLE:
103 dens identified to date throughout entire study area (reused dens counted only once).
51 (49.5\%) dug dens, $40(38.8 \%)$ natural cavity dens, $9(8.7 \%)$ unknown cavity type. $3(2.9 \%)$ tree dens.
Watana dens ( $\mathrm{N}=44$ ) Devils Canyon dens ( $\mathrm{N}=30$ ) Downstream dens ( $\mathrm{N}=\mathrm{N}^{\prime} 29$ )

|  |  |  | Tree | 3(10.3\%) |  |
| :--- | :---: | :--- | :---: | ---: | ---: |
| Dug | $24(54.5 \%)$ | Dug | $10(33.3 \%)$ | Dug | $17(58.6 \%)$ |
| Natural | $18(40.9 \%)$ | Natural | $13(43.3 \%)$ | Natural | $9(31.0 \%)$ |
| Unknown | $2(4.5 \%)$ | Unknown | $7(23.3 \%)$ |  |  |
| Flooded | $24(54.5 \%)$ | Flooded | $1(3.3 \%)$ | Flooded | $0(0.0 \%)$ |
| Not flooded | $20(45.5 \%)$ | Not flooded | $28(93.3 \%)$ | Not flooded | $29(100.0 \%)$ |

Table 76. Daily search effort for each quadrat for the spring 1985 bear population estimate of the Su-Hydro study area. Commuting and circling time not included.

| For each day: <br> Search time (minutes) /Spotter Plane Number* |  |  |  |  |  |  |  |  |  |  | Total Minutes | Total$\text { Min/mi }{ }^{2}$ | Total$\mathrm{Min} / \mathrm{km}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Quadrat $\qquad$ | $\mathrm{Mi}^{2}$ | $\mathrm{Km}^{2}$ | 6/1 | 6/2 | 6/3 | 6/4 | 6/5** | 6/9 | 6/10 | 6/11 |  |  |  |
| 1 | 56.52 | 146.38 | (19)/2 | 132/2 | 93/2 | 100/2 | 233/1 | (209)/1 | (17)/1 | 156/2 | 959 | 17.0 | 6.6 |
| 2 | 63.64 | 142.89 | -- | 253/1 | 183/1 | 121/1 | 172/2 | (90)/2 | (72)/2 | (93)/3 | 984 | 15.4 | 6.9 |
| 3 | 38.62 | 100.02 | 120/3 | 131/3 | 82/2 | -- | 175/2,3 | -- | 110/1 | 85/2 | 703 | 18.2 | 7.0 |
| 4 | 49.30 | 127.67 | 106/1 | 89/1 | 96/1 | 120/1 | 168/3 | (4)/1 | 157/2 | 116/1 | 856 | 17.4 | 6.7 |
| 5 | 55.17 | 142.89 | 49/1 | 167/3 | 138/2 | 120/3 | -- | 121/1 | 103/1 | (10)/1 | 708 | 12.8 | 5.0 |
| 6 | 33.76 | 87.42 | 148/2 | 79/2 | 93/1 | 149/2 | (16)/2 | (12)/2 | 180/3 | (107)/1 | 784 | 23.2 | 9.0 |
| 8 | 64.72 | 167.62 | 214/3,2 | (62)/3 | 173/1,3 | 174/3 | -- | 166/1 | 210/2 | 169/1 | 1168 | 18.1 | 7.0 |
| 9 | 83.29 | 215.71 | 118/2 | 104/1 | -- | 151/1 | -- | 211/2 | (166)/1,3 | (61)/3 | 811 | 9.7 | 3.8 |
| 10 | 75.10 | 194.50 | 96/1 | (50)/2 | (77)/2 | 148/2 | (7)/2 | 120/2 | 217/1 | -- | 715 | 9.5 | 3.7 |
| Total 5 | 520.12 | 1325.10 | $\begin{gathered} 870 \\ 4.5 \mathrm{~h} \end{gathered}$ | 1,067 17.8 h | $\begin{gathered} 935 \\ 15.6 \mathrm{~h} \end{gathered}$ | 1,083 18.1 h | $\begin{aligned} & 771 \\ & 12.9 \mathrm{~h} \end{aligned}$ | $\begin{array}{r} 933 \\ 15.6 \mathrm{~h} \end{array}$ | $\begin{aligned} & 1,232 \\ & 20.5 \mathrm{~h} \end{aligned}$ | $\begin{gathered} 797 \\ 13.3 \mathrm{~h} \end{gathered}$ | $\begin{aligned} & 7,688 \\ & 128.1 \mathrm{~h} \end{aligned}$ | 14.8 | 5.8 |

* Spotter Pilot \#1 = McMahan, \#2 = Lee, \# 3 = Deering
** Bad weather on $6 / 5 / 85$ and on the 3 days following
()$=$ partially done



Figure 2. Capture locations for 53 brown bears xadiocollared in the upstream study area. Polygon
incorpora
0 incorporates an area of $2,169 \mathrm{kn}^{2}$. Females are indicated with a hexagon, males ith an asterisk.



Figure 4. grown bear study area, Illustrated poipgons are a thean hoate range diameter $137.5 \mathrm{~km}=\mathrm{miapant}$ between average alale and average femaie home range diameter atound inpoument zones. Defined ingoundment zone is between Devils canyon and confluence of oshetna and Susiona Rivers. Total area of impondmenc zone
 overiap. Portion of each polygon that is above 5,000 feet slevation definea as rot brom beas holicery





Figure 5. Point locations $(N=2 ; 195)$ for radiomarked brown bears captured in the downstream study area, 1980-1984. Polygon incorporates an area of

Scale 1:1200000 $2,946 \mathrm{~km}^{2}$, females indicated with a hexagon, males

0 60 with an asterisk.
 downstream black bears.



Figure 8. Capture locations of 22 black bears radio-collared in the

Scale 1:500000 Km.
0 downstream study area. Polygon incorporates an area of $250 \mathrm{~km}^{2}$, females indicated with a hexagon, males with an asterisk.
 377, 378, 402, 404, 405, 406, 408, 409, 410, and 411.


Eigure 10. Tllustration of proximity polygons that are l mile and 5 miles trom the shoreline of proposed Watand and Devils
Canyon Impoundments.


7igare 11. Percent of brown bear point locations in each of 4 inpoundment proxinity zones, by month. hll radiolocations in 1980-1984 are incivded except for den site locations. Number of point locations for months 5 ( (Hay), 6 (June), 7 (JuIy), 8 (August), 9 (September), and 10-4 (October through April) are, respectiveif: 339, 633, 211, 184, 159, and 92 for Hataza Impoundreat zones (above), and 104, 174, 125, 90, 68, and 30 for Depils canyon Impoundment zopes (ietiol). BR.BEAR USE OF DEVILS CAN. PROX. ZONES








Figure 17. Annual variation in ean home range size of radio-narked black bear wales and females loniy fenales without newborn cubs included). Rumber indicates sample size used in calculation of mean and standard deviation.

ANNUAL BROWN BEAR HOME RANGE SIZES







## APRIL 1 SNOW DEPTHS, 4 STATIONS

 IN SU-HYDRO VCINITY

Rigure 23 . Amual snow depths on 1 hpril and 1 May during 1980 through 1985 at 4 snov surver stations in the picinity of the proposed inpoundments. Data provided by v.S. Soil Conservation Servise, snow Survey.


## Aspects of Brown Bear Dens

 Includes $+/-22.5$ Degrees, true north

Figure 24. Aspects of 89 brown bear den side. Aspects includes arc of $\pm 22.5$ degrees on from true north. corrected for magnetic deviation

Den aspect based on sex and status


Figure 25. Aspects of brown bear dens based on reproductive status of 27 females with newborn cubs (COY) at exit from den cavities, 30 females without newborn cubs at exit, and 12 males. Dens for 20 brown bears of unknown sex or reproductive status are not included. Indicated direction includes arc of +22.5 degrees on either side. Aspects corrected for magnetic deviation from true north..


Figure 26. Location of den sites for radio-marked brown bears in the su-Hydro study area upstream from the Devils Canyon dam site in springs of 1981

Beale 1:800000
(triangles, $N=9$ ), 1982 (squares with $X, N=12$ ),


1983 (squares with tails, $N=18$ ), 1984 (6-pointed
stars, $N=25)$, and $1985(*, N=19)$. Diamonds are
for den sites of unmarked bears ( $N=8$ )



Figure 28. Percent of black bear point locations in each of 4 Tatana Das iapoundment proxinity 20nes, by month. Ali radio-locations in 1980-1984 are included except for den site locations. Ruber of point locations for months 5 (Hay), 6 (June), 7 (July), 8 (August), 9 (Septeaber), and $10-4$ (October-April) are, respectively: 222, 465, 203, 226, 154, and 35 for Fiatana iapoundment zones and 141, 289, 98, 84, 58, and 9 for Devils Canyon impoundrent zones.

BK. BEAR USE OF DEVILS CAN. PROX. ZONES



Figure 29 . Upstream movement of black bear 321
during poor berry summer of 1981.


Figure 30 . Upstream movement of black bear 318 during poor berry summer of 1981.


Figure 31 . Upstream movement of black bear 342 b during poor berry summer of 1981 .


Figure 32. Downstream movement of black bear 343
during poor berry summer of 1981.


Figure 33. Downstream movement of black bear 324
during poor berry summer of 1981.

pigure 34. hanual variation in nean hone range size of radio-narked brom bear nales and fenales lonly fenales without nenborn cubs included). Nuber indicates sanple size used in calculation of nean and standard deviation.

ANNUAL BLACK BEAR HOME RANGE SIZES

 upstream from the Devils Canyon dam site during spring of 1981 (triangles, $N=16$ ), 1982
(squares with X, $N=13$ ), 1983 (squares with
feale 1:600000
tails, $N=21$ ), 1984 (6-pointed stars, $N=12$ ),
and 1985 (*, $N=15$ ). Diainonds are for den
sites of unmarked bears ( $10=3$ ).


Aspect of Black Bear Dens


Figure 37 a. Aspects of black bear dens in 45 dug cavities and 36 natural rock cavities. Indicated direction includes arc of +22.5 degrees on either side. Aspects corrected for magnetic deviation from true north.

Aspect of Black Bear Dens


Figure 37 b : Aspects of black bear dens for 27 females with newborn cubs (COY) at exit from den cavities, 30 females without newborn cubs at exit, and 24 males. Indicated direction includes arc of +22.5 degrees on either side. Aspects corrected for magnetic deviation from true north.



## BLUEBERRY RIPENESS



Pigure 10. Ripeness phenology ( $\$$ of all plots read during period with berries in that category) for blueberry ( P ig. 40a), crouberry (Fig. 40b), and bighbush cranberry (Pig. 40c). Saple sizes indicated on left of point for green berries, on right of point for tart berries, and above point for ripe berries.

CROWBERRY RIPENESS


LOWBUSH CRANBERRY RIPENESS


Fiqure 40. Ripeness phenology (t) of all plots read during period with berries in that category) for blueberry (fig. 40al, crouberry ( Fig . 40bl, and bighbush cranberry ( Fig . 40c). Sample sizes indicated on left of point for green berries, on right of point for tart berries, and above point for ripe berries.

Figure 41. Canopy coverage for blueberries in each inpoundment zone and above 2200 feet elevation. Chi square analysis was based on estimated proportion in each class times the number of transects. Last two classes were lunped.

|  | Devils Canyon |  |
| :---: | :---: | ---: |
| Coverage | POPULATION D |  |
| Class | Est. Frop. SE w/Cov |  |
| None | 0.228 | 0.0481 |
| $\langle 5 \%$ | 0.238 | 0.0390 |
| $5-25 \%$ | 0.272 | 0.0330 |
| $26-50 \%$ | 0.179 | 0.0292 |
| $51-75 \%$ | 0.052 | 0.0120 |
| $775 \%$ | 0.031 | 0.0117 |

No. transects

## Watana

POPULATION A
Est. Prop. SE w/eov

| 0.244 | 0.0190 |
| :--- | :--- |
| 0.223 | 0.0124 |
| 0.292 | 0.0111 |
| 0.147 | 0.0113 |
| 0.067 | 0.0087 |
| 0.027 | 0.0050 |

$>2200$ feet elev. POPULATION 8
Est. Prop. SE v/COV

| 0.253 | 0.0205 |
| :--- | :--- |
| 0.145 | 0.0131 |
| 0.263 | 0.0194 |
| 0.197 | 0.056 |
| 0.100 | 0.0122 |
| 0.042 | 0.0085 |

## BLUEBERRY CANOPY COVERAGE



Figure 42. Canopy coverage for crowberries in each impoundment zone and above 2200 feet elevation. Chi square analysis was based on estinated proporticn in each class times the number of transects. Last three classes were iumped.
Coyerage
Class
Mone
$<5 \%$
$5-25 \%$
$26-50 \%$
$51-75 \%$
$>75 \%$

Devils Canyon

No. transects
FOPULATION D
Est. Prop. SE w/cov
0.3790 .0639
$0.256 \quad 0.0439$
$0.181 \quad 0.0225$
$0.131 \quad 0.0251$
$0.036 \quad 0.0119$
$0.016 \quad 0.0117$
43.00

Watana
POPULATION A
Est. Prop. SE w/cov

| 0.376 | 0.0210 |
| :--- | :--- |
| 0.140 | 0.0127 |
| 0.224 | 0.0138 |
| 0.148 | 0.0106 |
| 0.085 | 0.0104 |
| 0.028 | 0.0061 |

165.00
$>2200$ feet elev. POPULATION B
Est. Prop. SE w/COv

| 0.611 | 0.0257 |
| :--- | :--- |
| 0.178 | 0.0175 |
| 0.143 | 0.0160 |
| 0.059 | 0.0110 |
| 0.010 | 0.0029 |
| 0.000 | 0.0000 |

Chi Square ( 6 d.f.) $=29.3, P<0.001$


Figure 43. Canopy coverage for lowbush cranberries in each impoundment zone ang above 2200 feet elevation. Chi square analysis was basecu on estimatad proporaio: in each class times the number of transects. Last chree classes were lumped.
Coverage
Class
None
$65 \%$
$5-25 \%$
$26-50 \%$
$51-75 \%$
$>75 \%$

## Devils Canyon

 POPULATION DEst. Prop. SE $\mathrm{N} / \mathrm{COV}$

| 0.171 | 0.0483 |
| :--- | :--- |
| 0.409 | 0.0459 |
| 0.289 | 0.0385 |
| 0.097 | 0.0193 |
| 0.031 | 0.0118 |
| 0.003 | 0.0026 |

Watana
POFULATION A
Est. Prop. SE w/COV

| 0.262 | 0.0187 |
| :--- | :--- |
| 0.454 | -0.0138 |
| 0.214 | 0.0128 |
| 0.055 | 0.0084 |
| 0.012 | 0.0033 |
| 0.004 | 0.0020 |

>2200 feet elev. FOFULATION B
Est. Prop. SE w/00\%
0.1880 .0182
$0.306 \quad 0.0218$
$0.330^{\circ} \quad 0.0202$
$0.138 \quad 0.0150$
$0.033 \quad 0.0067$
$0.006 \quad 0.0032$


Figure 44. Canopy coverage for Equisetum in each impoundment zone and above 2200 feet elevation. Chi square analysis was based on estimated paporitat , in each class times the number of transects. Last four classes mere lumped.

|  | Devils Canyon |  |
| :---: | :---: | :---: |
| Coverage | POPULATION D |  |
| Class | Est. Prop SE W/cov |  |
| None | 0.503 | 0.0651 |
| $\langle 5 \%$ | 0.361 | 0.0627 |
| $5-25 \%$ | 0.075 | 0.0172 |
| $26-50 \%$ | 0.035 | 0.0120 |
| $51-75 \%$ | 0.012 | 0.0119 |
| $>75 \%$ | 0.014 | 0.0086 |

Watana
POPULATION A
Est. Prop. SE w/COV 0.6920 .0120
$0.178 \quad 0.0208$
$0.074 \quad 0.0090$
$0.029 \quad 0.0058$
$0.016 \quad 0.0034$
$0.011 \quad 0.0032$
) 2200 feet elev. FOPULATION B
Est. Frop. SE w/cov
$0.358 \quad 0.0255$
0.2180 .0186
$0.122 \quad 0.0150$
$0.046 \quad 0.0083$
$0.028 \quad 0.0060$
0.0190 .0059

125

EQUISETUM CANOPY COVERAGE
Chi Square (4. d.f.) $=11.5, P=0.02$


Figure 45. Abundance data for blueberries in each impoundment zone and above 2200 feet elevation. Chi square analysis was based on estimated proportion in each class times the number of transects. Last two classes were lumped.

| Abundance | Devils Canyon |  | Watana |  | >2200 feet elev. |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | POPULATION | D | POPULATTON | A | POPULATION | B |
| Class | Est. Prop. | SE W/COV | Est. Prop. | SE w/COV | Est. Prop. | SE W/Cov |
| None | 0.764 | 0.0421 | 0.796 | 0.0171 | 0.681 | 0.0210 |
| 1-4 | 0.110 | 0.0244 | 0.117 | 0.0119 | 0.155 | 0.0163 |
| 5-20 | 0.086 | 0.0181 | 0.065 | 0.0074 | 0.123 | 0.0118 |
| )20 | 0.040 | 0.0136 | 0.021 | 0.0060 | 0.041 | 0.0081 |
| Mo. transe |  | 43 |  | 165 |  | 126 |

## BLUEBERRY ABUNDANCE



Figure 45. Abundance of crowberries in each impoundment zone and
above 2200 feet elevation. Chi square analysis was based on estimated propoucian in each class times the number of transects. Last three classes vere lumpei.

|  | Devils Canyon <br> Abundance <br> POPULATION D |  |
| :---: | ---: | ---: |
| Class | EST. FROF. SE w/COV |  |
| None | 0.758 | 0.0391 |
| $1-4$ | 0.102 | 0.0241 |
| $5-20$ | 0.073 | 0.0141 |
| 20 | 0.067 | 0.0241 |

Watana
FOPULATTON A
EST. PROF. SE w/COV

| 0.618 | 0.0144 |
| :--- | :--- |
| 0.102 | 0.0080 |
| 0.125 | 0.0105 |
| 0.153 | 0.0134 |

32200 feat elev. POPULATION B
EST. FROP. SE \%/CO\%

| 0.875 | 0.0153 |
| :--- | :--- |
| 0.052 | 0.0100 |
| 0.048 | 0.0073 |
| 0.026 | 0.0056 |

No. transects
43
165
46

CROWBERRY ABUNDANCE
Chi Square (2 d.f.) $=24.0, P<0.001$


Figure 47. Abundance of lowbush cranberry berries in each impoundment zone anc above 2200 feet eievation. Chi square analysis was based on estimated proportion in each class times the number of transects. Last three classes were lumped.

|  | Devils Canyon |  | Watana |  | >2200 feet elev. |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Abundance | POPULATION | D | POPULATION | A | POPULATION | B |
| Class | Est. Prop. | SE w/COV | Est. Prop. | SE m/COV | Est. Prop. | SE w/cov |
| None | 0.845 | 0.0315 | 0.873 | 0.0136 | 0.744 | 0.0205 |
| 1-4 | 0.101 | 0.0288 | 0.038 | 0.0064 | 0.068 | 0.0103 |
| 5-20 | 0.040 | 0.0128 | 0.047 | 0.0074 | 0.119 | 0.0134 |
| >20 | 0.014 | 0.0075 | 0.042 | 0.0074 | 0.069 | 0.0110 |

Ho. transects
43
165
125



[^0]:    * Reject null hypothesis, p.less than 0.10 .
    ** Reject null hypothesis, p less than 0.05.

[^1]:    * Reject null hypethesis, p less than 0.10 .
    ** Reject null hypothesis, $p$ less than 0.05 .

[^2]:    ** possible unreported hunter kill, collar failure, or emigration.

[^3]:    * Flight on $8 / 6 / 85$ was in a $180 \mathrm{w} / 3$ observers and area was incompletely covered.

[^4]:    Not Included:
    Subadults @2 in $1979=259$.
    Subadults @i in $1979=275,262$ or $263,256,257,252,253,245,271,239,238$.
    ** Recaptured in Su-Hydro area (6/85).

[^5]:    (continued on next page)

[^6]:    * Standard minimum grid method (Mohr 1947).

[^7]:    * Reject null hypothesis, p less than 0.10.
    ** Reject null hypothesis, $p$ less than 0.05.

[^8]:    * Downstream study area

[^9]:    (continued on next page)

[^10]:    (continued on next page)

[^11]:    * Dates were designated from a point value rather than a time period, because a more accurate mean emergence date was produced.

[^12]:    (continued on next page)

